Chapter 10

Agricultural Waste Management System Component Design

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Chapter 10

Agricultural Waste Management System Component Design

651.1000 Introduction

Alternatives for managing agricultural waste are available for any given agricultural operation. As described in chapters 2 and 9, an agricultural waste management system can consist of any one or all of the following functions: production, collection, storage, treatment, transfer, and utilization. These functions are carried out by planning, applying, and operating individual components.

A component can be a piece of equipment, such as a pump; a structure, such as a waste storage tank; or an operation, such as composting. The combination of the components should allow the flexibility needed to efficiently handle all forms of waste generated for a given enterprise. In addition, the components must be compatible and integrated with each other. All components should be designed to be simple, manageable, and durable, and they should require low maintenance. In this chapter, components are discussed under section headings that describe the function that they are to accomplish.

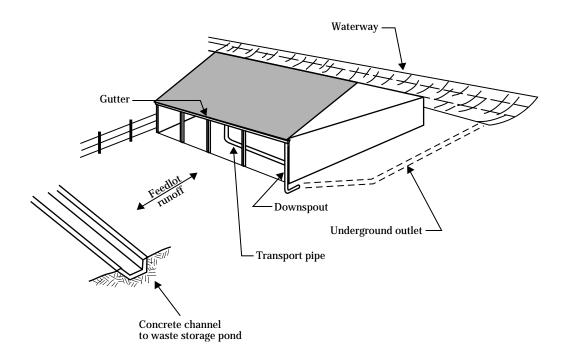
651.1001 Production

Components that affect the volume and consistency of agricultural waste produced are included in the production function. Roof gutters and downspouts and diversion to exclude clean water from areas of waste are examples of components that reduce the volume of waste material that needs management. Fences and walls that facilitate collection of waste confine the cattle, thus increase the volume.

(a) Roof runoff management

Roof runoff should be diverted from feedlots and manure storage areas unless it is needed for some use, such as dilution water for waste storage ponds or treatment lagoons. This can be accomplished by roof gutters and downspouts with underground or open channel outlets (fig. 10–1). Gutters and downspouts may not be needed if the roof drainage will not come into contact with areas accessible to livestock.

Figure 10-1 Roof gutter and downspout



The area of a roof that can be served by a gutter and downspout system is controlled by either the flow capacity of the gutter (channel flow) or by the capacity of the downspout (orifice flow). The gutter's capacity may be computed using Manning's equation. Design of a gutter and downspout system is based on the runoff from a 10-year frequency, 5-minute rainfall except that a 25-year frequency, 5-minute rainfall is used for exclusion of roof runoff from waste treatment lagoons, waste storage ponds, or similar practices.

Rainfall intensity maps are in appendix 10B. Caution should be used in interpolating these maps. Rainfall probabilities are based on measured data at principal weather stations that are mostly in populated regions. The 10-year, 5-minute rainfall in the 11 Western States was based on NOAA Atlas 1, and that in the 37 Eastern States was based on the National Weather Service HYDRO 35. Both of these publications state their limitations in areas of orographic effect. In the Western States, the 10-year, 5-minute rainfall generally is larger in mountain ranges than in valleys. Rainfall in all mountain ranges could not be shown on these maps because of the map scale and readability considerations. Many of these differences were in the range of 0.05 inch and fall within the contour interval of 0.10 inch.

A procedure for the design of roof gutters and downspouts follows:

Step 1—Compute the capacity of the selected gutter size. This may be computed using the Manning's equation. Using the recommended gutter gradient of 1/16 inch per foot and a Manning's roughness coefficient of 0.012, this equation can be expressed as follows:

$$q_g = 0.01184 \times A_g \times r^{0.67}$$

where:

 q_g = capacity of gutter, ft³/sec A_g = cross sectional area of gutter, in² r = A_g/wp , inches

wp =wetted perimeter of gutter, inches

Step 2—Compute capacity of downspout. Using an orifice discharge coefficient of 0.65, the orifice equation may be expressed as follows:

$$q_d = 0.010457 \times A_d \times h^{0.5}$$

 q_d = capacity of downspout, ft³/sec

 A_d^u = cross sectional area of downspout, in² h = head, inches (generally the depth of the gutter minus 0.5 inch)

Step 3—Determine whether the system is controlled by the gutter capacity or downspout capacity and adjust number of downspouts if desired.

$$N_d = \frac{q_g}{q_d}$$

where:

 N_d = number of downspouts

If N_d is less than 1, the system is gutter capacity controlled. If it is equal to or greater than 1, the system is downspout capacity controlled unless the number of downspouts is equal to or exceeds N_d .

Step 4—Determine the roof area that can be served based on the following equation:

$$A_r = \frac{q \times 3,600}{P}$$

where:

 A_r = Area of roof served, ft²

= capacity of system, either q_g or $q_{d'}$ whichever is smallest, ft³/sec

= 5-minute precipitation for appropriate storm event, inches

The above procedure is a trial and error process. Different sizes of gutters and downspouts should be evaluated along with multiple downspouts to determine the best gutter and downspout system to serve the roof area involved.

(1) Design example 10-1—Gutters and downspouts

Mrs. Linda Worth of Pueblo, Colorado, has requested assistance in developing an agricultural waste management system for her livestock operation. The selected alternatives include gutters and downspouts for a barn having a roof with a horizontally projected area of 3,000 square feet. The 10-year, 5-minute precipitation is 0.5 inches. The procedure above is used to size the gutter and downspouts.

Step 1—Compute the capacity of the selected gutter size. Try a gutter with a 6-inch depth and 3-inch bottom width. One side wall is vertical, and the other is sloping, so the top width of the gutter is 7 inches. Note that a depth of 5.5 inches is used in the computations to allow for 0.5 inch of freeboard.

$$A_g = (3 \times 5.5) + (0.5 \times 3.67 \times 5.5)$$

$$= 26.6 \text{ in}^2$$

$$wp = 3 + 5.5 + (3.67^2 + 5.5^2)^{0.5}$$

$$= 15.1 \text{ in}$$

$$r = \frac{A_g}{\text{wp}}$$

$$= \frac{26.6}{15.1}$$

$$= 1.76 \text{ in}$$

$$q_g = 0.01184 \times A_g \times r^{0.67}$$

$$= 0.01184 \times 26.6 \times 1.76^{0.67}$$

$$= 0.46 \text{ ft}^3 / \text{sec}$$

Step 2—Compute capacity of downspout. Try a 3-inch diameter downspout

H = depth of gutter -0.5 in²
= 5.5 in

$$A_d = 3.1416 \times \left(\frac{3}{2}\right)^2$$

= 7.06 in²
 $q_d = 0.010457 \times 7.06 \times 5.5^{0.5}$
= 0.17 ft³ / sec

Step 3—Determine whether the system is controlled by the gutter capacity or downspout capacity and make adjustments to number of downspouts if desired. By inspection it can be determined that the gutter capacity (0.46 ft³/sec) exceeds the capacity of one downspout (0.17 ft³/sec) Unless a larger downspout or additional downspouts are used, the system capacity would be limited to the capacity of the downspout. Try using multiple down-

spouts. Determine number required to take advantage of gutter capacity.

$$N_d = \frac{q_g}{q_d}$$
$$= \frac{0.46}{0.17}$$
$$= 2.7$$

 N_d is greater than 1; therefore, with one downspout the system would be downspout controlled. With three, it would be controlled by the gutter capacity, or 0.46 ft³/sec. Use three downspouts to take full advantage of gutter capacity.

Step 4—Determine the roof area that can be served based on the following equation:

$$A_r = \frac{q \times 3,600}{P}$$
$$= \frac{0.46 \times 3,600}{0.5}$$
$$= 3,312 \text{ ft}^2$$

This exceeds the roof area to be served; therefore, the gutter dimension selected and the three downspouts with dimensions selected are okay.

(b) Runoff control

Essentially all livestock facilities in which the animals are housed in open lots or the manure is stored in the open must deal with runoff. "Clean" runoff from land surrounding livestock facilities should be diverted from barns, open animal concentration areas, and waste storage or treatment facilities (fig. 10–2). Runoff from feedlots should be channeled into waste storage facilities.

Appendix 10C presents a series of maps indicating the amount of runoff that can be expected throughout the year for paved and unpaved feedlot conditions. "Clean" runoff should be estimated using information in chapter 2 of the NRCS Engineering Field Manual or by some other hydrologic method.

Diversions are to be designed according to NRCS Conservation Practice Standard, Diversion, Code 362 (USDA 1985). Diversion channels must be maintained to remain effective. If vegetation is allowed to grow tall, the roughness increases and the channel velocity decreases causing possible channel overflow. Therefore, vegetation should be periodically mowed. Earth removed by erosion from earthen channels should be replaced. Unvegetated, earthen channels should not be used in regions of high precipitation because of potential erosion.

651.1002 Collection

Livestock and poultry manure collection often depends on the degree of freedom that is allowed the animal. If animals are allowed freedom of movement within a given space the manure produced will be deposited randomly. Components that provide efficient collection of animal waste include paved alleys, gutters, and slatted floors with associated mechanical and hydraulic equipment as described below.

(a) Alleys

Alleys are paved areas where the animals walk. They generally are arranged in straight lines between animal feeding and bedding areas. On slatted floors, animal hoofs work the manure through the slats into the alleys below, and the manure is collected by flushing or scraping the alleys.

(1) Scrape alleys and open areas

Two kinds of manure scrapers are used to clean alleys (fig. 10–3). A mechanical scraper is dedicated to a given alley. It is propelled using electrical drives attached by cables or chains. The drive units are often

Figure 10-2 Diversion of "clean" water around feedlot

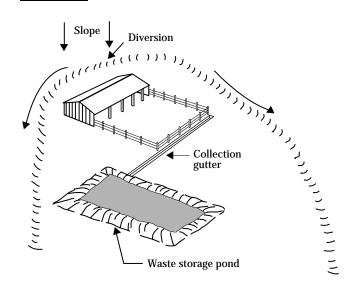
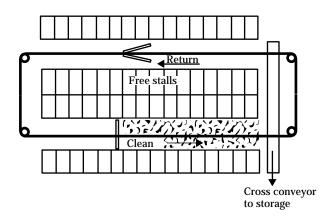


Figure 10–3 Scrape alley used in dairy barns



used to power two mechanical scrapers that are traveling in opposite directions in parallel alleys in an oscillating manner. Some mechanical scrapers are in alleys under slatted floors.

A tractor scraper can be used in irregularly shaped alleys and open areas where mechanical scrapers cannot function properly. It can be a blade attached to either the front or rear of a tractor or a skid-steer tractor that has a front-mounted bucket.

The width of alleys depends on the desires of the producer and the width of available equipment. Scrape alley widths typically vary from 8 to 14 feet for dairy and beef cattle and from 3 to 8 feet for swine and poultry.

(2) Flush alleys

Alleys can also be cleaned by flushing. Grade is critical and can vary between 1.25 and 5 percent. It may change for long flush alleys. The alley should be level perpendicular to the centerline. The amount of water used for flushing is also critical. An initial flow depth of 3 inches for underslat gutters and 4 to 6 inches for open alleys is necessary.

The length and width of the flush alley are also factors. Most flush alleys should be less than 200 feet long. The width generally varies from 3 to 10 feet depending on animal type. For underslat gutters and alleys, channel width should not exceed 4 feet. The width of open flush alleys for cattle is frequently 8 to 10 feet.

Table 10–1 Recommended total daily flush volumes (MWPS 1985)

Animal type	gal/head	
Swine		
Sow and litter	35	
Prenursery pig	2	
Nursery pig	4	
Growing pig	10	
Finishing pig	15	
Gestating sow	25	
Dairy cow	100	
Beef feeder	100	

Flush alleys and gutters should be cleaned at least twice per day. For pump flushing, each flushing event should have a minimum duration of 3 to 5 minutes.

Tables 10–1 and 10–2 indicate general recommendations for the amount of flush volume. Table 10–3 gives the minimum slope required for flush alleys and gutters. Figures 10–4 and 10–5 illustrate flush alleys.

Several mechanisms are used for flushing alleys. The most common rapidly empties large tanks of water or use high-volume pumps. Several kinds of flush tanks are used (fig. 10–6). One known as a tipping tank pivots on a shaft as the water level increases. At a certain design volume, the tank tips, emptying the entire amount in a few seconds, which causes a wave that runs the length of the alley.

Table 10-2 Flush tank volumes and discharge rates (MWPS 1985)

Initial flow depth, in.	Tank volume, gal/ft of gutter width	Tank discharge rate, gpm/ft of gutter width	Pump discharge, gpm/ft of gutter width
1.5	30	112	55
2.0	40	150	75
2.5	45	195	95
3.0	55	255	110
4.0	75	615	150
5.0	100	985	175
6.0	120	1,440	200

Table 10–3 Minimum slope for flush alleys (MWPS1985)

	Underslat alley	na	pen Al rrow v (<4')		wi	en Al de wie 4')	ley dth
Initial flow depth, in.	3.0	1.5	2.0	2.5	4.0	5.0	6.0
Slope, %	1.25	2.0	1.5	1.25	5.0	4.0	3.0

Some flush tanks have manually opened gates. These tanks are emptied by opening either a valve, a standpipe, a pipe plug, or a flush gate. Float switches can be used to control flushing devices.

Another kind of flush tank uses the principle of a siphon. In this tank the water level increases to a given point where the head pressure of the liquid overcomes the pressure of the air trapped in the siphon mechanism. At this point the tank rapidly empties, causing the desired flushing effect.

Most flush systems use pumps to recharge the flush tanks or to supply the necessary flow if the pump flush technique is used. Centrifugal pumps typically are used. The pumps should be designed for the work that they will be doing. Low volume pumps (10 to 150 gpm) may be used for flush tanks, but high volume pumps (200 to 1,000 gpm) are needed for alley flushing. Pumps should be the proper size to produce the desired flow rate. Flush systems may rely on recycled lagoon water for the flushing liquid.

In some parts of the country where wastewater is recycled from lagoons for flush water, salt crystals (struvite) may form inside pipes and pumps and cause decreased flow. Use of plastic pipe and fittings and pumps that have plastic impellers can reduce the frequency between cleaning or replacing pipes and

pumps. If struvite formation is anticipated, recycle systems should be designed for periodic clean out of pumps and pipe. A mild acid, such as dilute hydrochloric acid (1 part 20 mole hydrochloric acid to 12 parts water), can be used. A separate pipe may be needed to accomplish acid recycling. The acid solution should be circulated throughout the pumping system until normal flow rates are restored. The acid solution should then be removed. Caution should be exercised when disposing of the spent acid solution to prevent ground or surface water pollution.

(b) Gutters

Gutters are narrow trenches used to collect animal waste. They are often employed in confined stall or stanchion dairy barns and in some swine facilities.

(1) Gravity drain gutters

Deep, narrow gutters can be used in swine finishing buildings (fig. 10–7). These gutters are at the lowest elevation of the pen. The animal traffic moves the waste to the gutter. The gutter fills and is periodically emptied. Gutters that have Y, U, V, or rectangular cross sectional shapes are used in farrowing and nursery swine facilities. These gutters can be gravity drained periodically.

Figure 10-4 Dairy flush alley

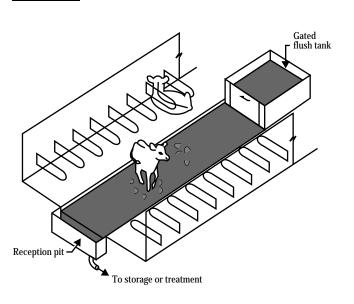
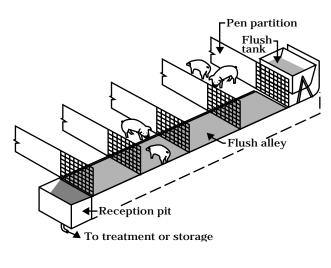


Figure 10–5 Swine flush alley



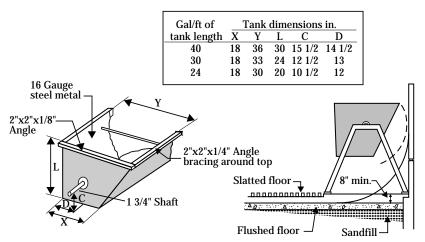
(2) Step-dam gutters

Step-dam gutters, which are also known as gravity gutters or gravity flow channels, provide a simple alternative for collecting dairy manure (fig. 10–8). A 6-inch high dam holds back a lubricating layer of manure in a level, flat-bottomed channel. Manure drops through a floor grate or slats and flows down the gutter under its own weight. The gutter is about 30 inches wide and steps down to a deeper cross channel below the dam.

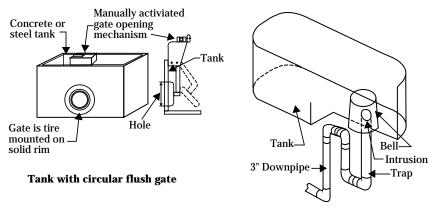
(3) Scrape gutters

Scrape gutters are frequently used in confined stall dairy barns. The gutters are 16 to 24 inches wide, 12 to 16 inches deep, and generally do not have any bottom slope. They are cleaned using either shuttle-stroke or chain and flight gutter cleaners (figs. 10–9 & 10–10). Electric motor driven shuttle stroke gutter cleaners have paddles that pivot on a drive rod. The drive rod travels alternately forward for a short distance and then backwards for the same distance. The paddles are designed to move manure forward on the forward stroke and to collapse on the drive rod on the return stroke. This action forces the manure down the gutter. Shuttle stroke gutter cleaners can only be used on straight gutters.

Figure 10-6 Flush tanks



Tipping tank



Automatic siphon tank

Chain and flight scrapers are powered by electric motors and are used in continuous loops to service one or more rows of stalls.

(4) Flush gutters

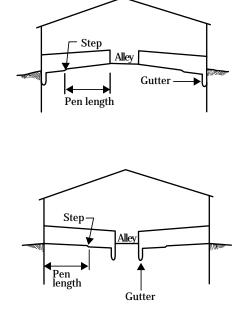
Narrow gutters can also be cleaned by flushing. Flush gutters are usually a minimum of 2 feet deep on the shallow end. The depth may be constant or increase as the length of the gutter increases. The bottom grade can vary from 0 to 5 percent depending or storage requirements and clean out technique. Flushing tanks or high volume pumps may be used to clean flush gutters (refer to the section on flush alternatives for alleys).

(c) Slatted floors

Waste materials are worked through the slats by the animal traffic into a storage tank or alley below. Most slats are constructed of reinforced concrete (fig. 10–11); however, some are made of wood, plastic, or aluminum. They are manufactured either as individual units or as gangs of several slats. Common slat openings range from 3/8 inch to 1 3/4 inches, depending on animal type. For swine, openings between 3/8 and 3/4 inch are not recommended.

Slats are designed to support the weight of the slats plus the live loads (animals, humans, and mobile equipment) expected for the particular facility. Reinforcing steel is required in concrete slats to provide needed strength.

Figure 10-7 Flush and gravity flow gutters for swine manure



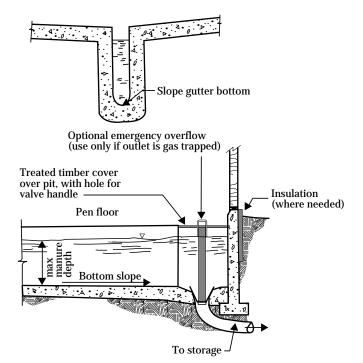
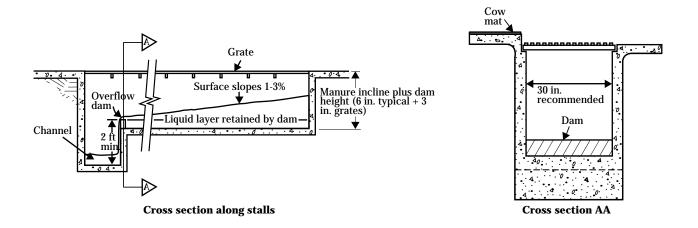


Figure 10-8 Gravity gutter for dairy manure



 $\textbf{Figure 10-9} \quad \text{Shuttle-stroke gutter cleaner} \\$

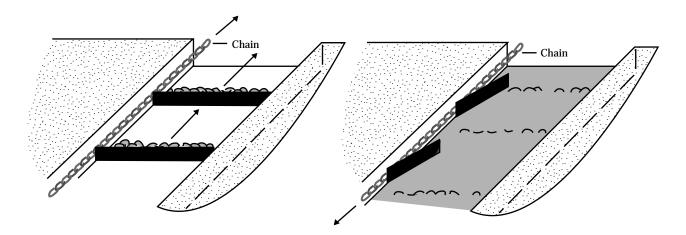


Figure 10-10 Chain and flight gutter cleaner

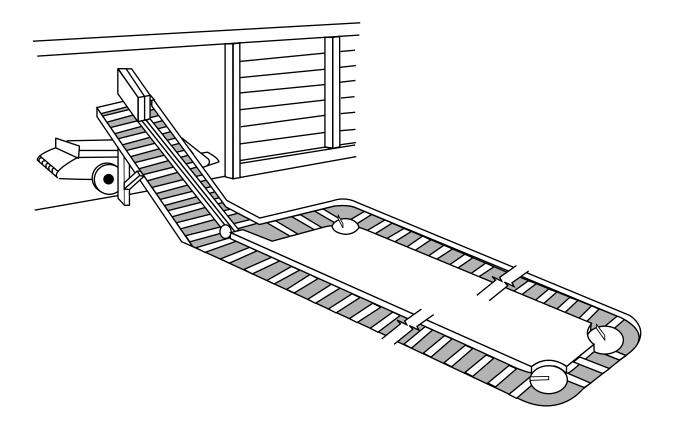
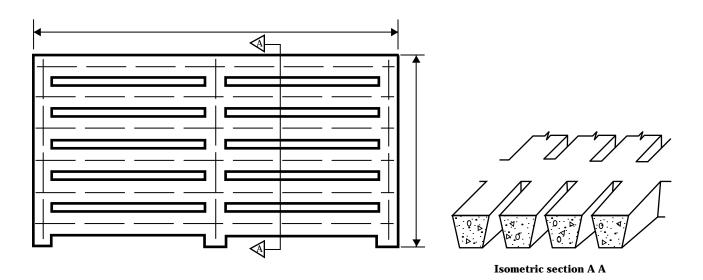


Figure 10-11 Concrete gang slats



651.1003 Storage

Waste generally must be stored so that it can be used when conditions are appropriate. Storage facilities for wastes of all consistencies must be designed to meet the requirements of a given enterprise.

Determining the storage period for a storage facility is crucial to the proper management of an agricultural waste management system. If too short a period is selected, the facility may fill before the waste can be used in an environmentally sound manner. Too long a period may result in an unjustified expenditure for the facility.

Many factors are involved in determining the storage period. They include the weather, crop, growing season, equipment availability, soil, soil condition, labor requirements, and management flexibility. Generally, when waste utilization is by land application, a storage facility must be sized so that it can store the waste during the nongrowing season. A storage facility that has a longer storage period generally will allow more flexibility in managing the wastes to accommodate weather variability, equipment availability, equipment breakdown, and overall operation management.

(a) Waste storage facilities for solids

Storage facilities for solid manure include waste storage ponds and waste storage structures. Waste storage ponds are earthen impoundments used to retain manure, bedding, and runoff liquid. Solid and semi-solid manure placed into a storage pond will most likely have to be removed as a liquid unless precipitation is low or a means of draining the liquid is available. The pond bottom and entrance ramps should be paved if emptying equipment will enter the pond.

Waste storage structures can be used for manure that will stack and can be handled by solid manure handling equipment. These structures must be accessible for loading and hauling equipment. They can be open or covered. Roofed structures are used to prevent or reduce excess moisture content. Open stacks can be

used in either arid or humid climate. Seepage and runoff must be managed. Structures for open and covered stacks often have wooden, reinforced concrete, or concrete block sidewalls. The amount of bedding material often dictates whether or not the manure can be handled as a solid.

In some instances manure must be stored in open stacks in fields. Runoff and seepage from these stacks must be managed to prevent movement into streams or other surface or ground water. Figures 10–12 and 10–13 show various solid manure storage facilities.

(1) Design considerations

Solid waste storage ponds and structures must be designed correctly to ensure desired performance and safety. Considerations include materials selection, control of runoff and seepage, necessary storage capacity, and proper design of structural components, such as sidewalls, floors, and roofs.

The primary materials used in constructing timber structures for solids storage are pressure-treated or rot-resistant wood and reinforced concrete. These materials are suitable for long-term exposure to animal waste without rapid deterioration. Structural grade steel is also used, but it corrodes and must be protected against corrosion or be periodically replaced. Similarly, high quality and protected metal fasteners must be used with timber structures to reduce corrosion problems.

Seepage and runoff, which frequently occur from manure stacks must be controlled to prevent access into surface and ground water. One method of control is to channel any seepage into a storage pond. At the same time uncontaminated runoff, such as that from the roof and outside the animal housing and lot area, should be diverted around the site.

Concrete ramps are used to gain access to solid manure storage areas. Ramps and floors of solid manure storage structures need to be designed so that handling equipment can be safely operated. Ramp slopes of 8 to 1 (horizontal to vertical) or flatter are considered safe. Slopes steeper than this are difficult to negotiate. Concrete pavement for ramps and storage units should be rough finished to aid in traction. Ramps need to be wide enough that equipment can be safely backed and maneuvered.

Factors to consider in the design of storage facilities for solids include type, number and size of animals, number of days storage desired, and the amount of bedding that will be added to the manure. Equation 10–1 can be used to calculate the manure storage volume:

$$VMD = AU \times DVM \times D$$
 [10-1]

where:

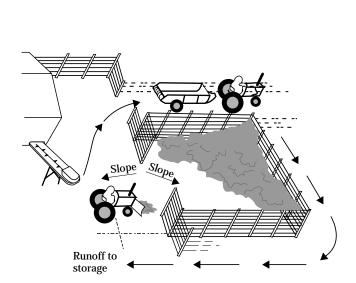
VMD = volume of manure production for animal type for storage period, ft³

AU = number of 1,000 pound animal units by animal type

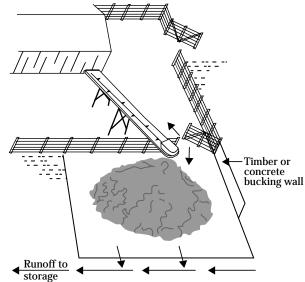
DVM = daily volume of manure production for animal type, ft³/AU/day

D = Number of days in storage period

Figure 10–12 Solid manure stacking facilities

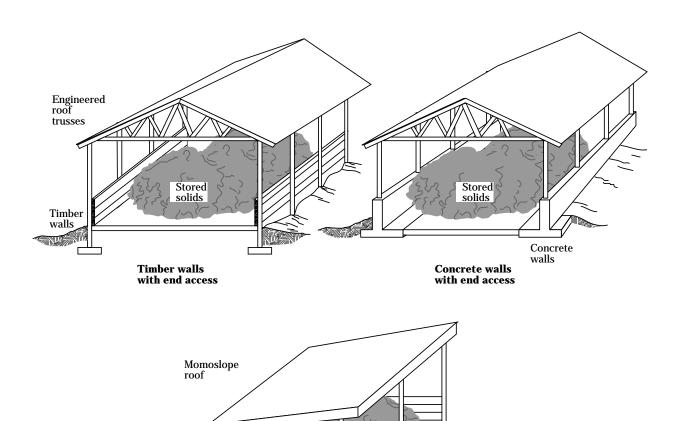


Barn cleaner to spreader or tractor stacking



To storage and/or spreader from elevator stacker

Figure 10-13 Roofed solid manure storage



Stored solids

Timber walls with side access

The bedding volume to be stored can be computed using:

$$BV = \frac{FR \times WB \times AU \times D}{BUW}$$
 [10-2]

where:

FR = volumetric void ratio (ASAE 1982) (values range from 0.3 to 0.5)

WB = weight of bedding used for animal type, lb/ AU/day

BUW= bedding unit weight, lb/ft3

Using the recommended volumetric void ratio of 0.5, the equation becomes:

$$BV = \frac{0.5 \times WB \times AU \times D}{BUW}$$

Characteristics of manure and bedding are described in chapter 4. Other values may be available locally or from the farmer or rancher.

Allowance must be made for the accumulation of precipitation that may fall directly into the storage. Contaminated runoff should be handled separately from a solid manure storage facility. Uncontaminated runoff should be diverted from the storage unit.

(2) Design example 10-2—Waste stacking facility

Mr. Ralph Kilpatrick of Hoot Ridge, Kentucky, has requested assistance in developing a waste management system. He selected an alternative that includes solid manure storage for his 100 Holstein milking cows and 52 heifers. His nutrient management plan indicates the need for 90 days storage. He uses sawdust bedding for both the milking cows and the heifers. Because of space limitations the storage can be no wider than 50 feet. He would prefer that the facility be no more than 7 feet deep. The structure will not be roofed, so stacking above sidewalls will not be considered in design. Determine the necessary volume and facility dimensions using worksheet 10A–1.

Manure production—The animal descriptions, average weight, and numbers are entered on **lines 1** and **2**. The number of equivalent animal units for each animal type is calculated and entered on **line 4**. Daily manure production **(line 4)** is in table 4–5 in chapter

4. The number of days in storage is entered on **line 5**. The manure volume **(line 7)** is calculated using equation 10–1. Add the calculated manure volume for each animal type (VMD) and enter the sum (TVM) on **line 8**.

Wastewater volume—Because this design example involves a waste stacking facility, it would not be appropriate to include wastewater in the storage facility. Therefore, **lines 9, 10,** and **11** are not involved in estimating the waste volume for this example.

Bedding volume—The weight of bedding used daily per animal unit for each animal type is entered on **line 12.** The bedding unit weight, which may be taken from table 4–4, is entered on **line 13**. The bedding volume for each animal type for the storage period is calculated using equation 10–2 and entered on **line 14.** The total bedding volume (TBV) is the sum of the bedding volume for all animal types. Sum the calculated bedding volume (BV) for each animal type and enter it on **line 15.**

Waste volume—The total waste volume (WV) **(line 16)** is the sum of the total manure production (TVM) and the total bedding volume (TBV). The storage width and depth are known, so the length **(line 17)** is calculated using the equation:

$$L = \frac{WV}{WI \times H}$$

A waste storage structure for solids should be designed to withstand all anticipated loads. Loadings include internal and external loads, hydrostatic uplift pressure, concentrated surface and impact loads, water pressure because of the seasonal high water table, and frost or ice pressure.

The lateral earth pressure should be calculated from soil strength values determined from results of appropriate soil tests. If soil strength tests are not available, the minimum lateral earth pressure values indicated in the NRCS Conservation Practice Standard, Waste Storage Facility, Code 313, are to be used (NRCS 1995).

Timber sidewalls for storage structures should be designed with the load on the post based on full wall height and spacing of posts.

Completed worksheet for Design example 10-2

Decisionmaker: Ralph Kilpatrick	Date: 6/13/91
Hoot Ridge, KY	
Animal units	
	3. Number of animals (N)
2. Animal weight, lbs (W)1,4001,000	4. Animal units, AU = W x N =
Manure volume	
5. Daily volume of daily manure production per AU, ft³/AU/day (DVM)= 1.30 1.30	7. Total volume of manure production for animal type for storage period, ft ³
6. Storage period, days (D) =90	VMD = AU x DVM x D = 10,300
c. c.o.ago portar, aayo (c)	8. Total manure production for storage period, t^3 (TVM)
Wastewater volume	
9. Daily wastewater volume per AU, ft ³ /AU/day (DWW) =	11. Total wastewater volume for O storage period, ft ³ (TWW)
10. Total wastewater volume for animal	
description for storage period, ft ³ WWD = DWW x AU x D =	
Bedding volume	
12. Amount of bedding used daily	14. Bedding volume for animal type for storage period, ft ³ (BV) = $\frac{1,628}{604}$
12. Amount of bedding used daily for animal type, 1bs/AU/day (WB) = 3.1 3.1	_
13. Bedding unit weight, 12 lbs/ft ³ (BUW) =	2 BV= 0.5 x WB x AU x D BUW
IDS/IL (BUW) =	
	15. Total bedding volume for storage 2,232
	15. Total bedding volume for storage period, ft ³ (TBV) = $\frac{2,232}{}$
Waste volume requirement	15. Total bedding volume for storage 2,232 period, ft ³ (TBV) =
Waste volume requirement	
•	15. Total bedding volume for storage (TBV) = 2,232
16. Waste volume, ft ³ (WV) = TVM + TWW + TBV = 22,40 Waste stacking structure sizing	64 + 0 + 2,232 = 24,696
16. Waste volume, ft ³ (WV) = TVM + TWW + TBV = 22,40 Waste stacking structure sizing	64 + 0 + 2,232 = 24,696
16. Waste volume, ft ³ (WV) = TVM + TWW + TBV = $22,40$ Waste stacking structure sizing 17. Structure length, ft L = $\frac{WV}{WI \times H}$ = 88.2	(USE 90) 19. Structure height, ft H = WV = 7
16. Waste volume, ft ³ (WV) = TVM + TWW + TBV = $22,40$ Waste stacking structure sizing 17. Structure length, ft L = $\frac{WV}{WI \times H}$ = 88.2	(USE 90) 19. Structure height, ft H = WV = 7
16. Waste volume, ft ³ (WV) = TVM + TWW + TBV = $\frac{22,40}{\text{Waste stacking structure sizing}}$ 17. Structure length, ft L = $\frac{\text{WV}}{\text{WI x H}} = \frac{88.2}{\text{L x H}}$ 18. Structure width, ft WI = $\frac{\text{WV}}{\text{L x H}} = \frac{40}{\text{L x H}}$	(USE 90) 19. Structure height, ft H = WV = 7
16. Waste volume, ft ³ (WV) = TVM + TWW + TBV = $\frac{22,40}{40}$ Waste stacking structure sizing 17. Structure length, ft L = $\frac{WV}{WI \times H}$ = $\frac{88.2}{L \times H}$ 18. Structure width, ft WI = $\frac{WV}{L \times H}$ = $\frac{40}{L \times H}$ Notes for waste stacking structure:	(USE 90) 19. Structure height, ft H = WV L x WI =
Waste stacking structure sizing 17. Structure length, ft $L = \frac{WV}{WI \times H} = \frac{88.2}{40}$ 18. Structure width, ft $WI = \frac{WV}{L \times H} = \frac{40}{100}$ Notes for waste stacking structure: 1. The volume determined (WV) does not include any volume for freeboard. It is recommended that a minimum of 1 foot of	(USE 90) 19. Structure height, ft H = WV = 7
16. Waste volume, ft ³ (WV) = TVM + TWW + TBV = $\frac{22,40}{40}$ Waste stacking structure sizing 17. Structure length, ft L = $\frac{WV}{WI \times H}$ = $\frac{88.2}{L \times H}$ 18. Structure width, ft WI = $\frac{WV}{L \times H}$ = $\frac{40}{L \times H}$ Notes for waste stacking structure: 1. The volume determined (WV) does not include any volume for freeboard. It is recommended that a minimum of 1 foot of freeboard be provided for a waste stacking structure.	(USE 90) 19. Structure height, ft H = WV L x WI =
16. Waste volume, ft ³ (WV) = TVM + TWW + TBV = $\frac{22,40}{\text{Waste stacking structure sizing}}$ 17. Structure length, ft L = $\frac{\text{WV}}{\text{WI x H}}$ = $\frac{88.2}{\text{L x H}}$ 18. Structure width, ft WI = $\frac{\text{WV}}{\text{L x H}}$ = $\frac{40}{\text{Notes for waste stacking structure:}}$ 1. The volume determined (WV) does not include any volume for freeboard. It is recommended that a minimum of 1 foot of freeboard be provided for a waste stacking structure.	(USE 90) 19. Structure height, ft H = \frac{WV}{L x WI} = \frac{7}{L x WI} = \frac{7}{L x WI} = \frac{7}{L x WI} = \frac{1}{100} = \frac{1}{1
Waste stacking structure sizing 17. Structure length, ft $L = \frac{WV}{WI \times H} = \frac{88.2}{40}$ 18. Structure width, ft $WI = \frac{WV}{L \times H} = \frac{40}{100}$ Notes for waste stacking structure: 1. The volume determined (WV) does not include any volume for freeboard. It is recommended that a minimum of 1 foot of freeboard be provided for a waste stacking structure. Tank sizing	(USE 90) 19. Structure height, ft H = \frac{WV}{L \times WI} = \frac{7}{L \times WI} 2. The equations for L, WI, and H assume manure is stacked to average height equal to the sidewall height. Available storage volume must be adjusted to account for these types of variations. 22. Rectangular tank dimensions Total height, ft (H) = \frac{VV}{VVV} Selected width, ft (WI) = \frac{VVV}{VVV}
Waste stacking structure sizing 17. Structure length, ft L = \frac{WV}{WI x H} = \frac{40}{L x H} 18. Structure width, ft WI = \frac{WV}{L x H} = \frac{40}{L x H} Notes for waste stacking structure: 1. The volume determined (WV) does not include any volume for freeboard. It is recommended that a minimum of 1 foot of freeboard be provided for a waste stacking structure. Tank sizing 20. Effective depth, ft. (EH) Total height (or depth) of tank desired, ft (H) Less precipitation for storage period, ft	(USE 90) 19. Structure height, ft H = \frac{WV}{L x WI} = \frac{7}{L x WI} 2. The equations for L, WI, and H assume manure is stacked to average height equal to the sidewall height. Available storage volume must be adjusted to account for these types of variations. 22. Rectangular tank dimensions Total height, ft (H) = \frac{SA}{WI} = \frac{SA}{WI}
16. Waste volume, ft³ (WV) = TVM + TWW + TBV = 22,40 Waste stacking structure sizing 17. Structure length, ft L = WV WI x H 18. Structure width, ft WI = WV L x H Notes for waste stacking structure: 1. The volume determined (WV) does not include any volume for freeboard. It is recommended that a minimum of 1 foot of freeboard be provided for a waste stacking structure. Tank sizing 20. Effective depth, ft. (EH) Total height (or depth) of tank desired, ft (H) Less precipitation for storage period, ft. — - (uncovered tanks only)	(USE 90) 19. Structure height, ft H = \frac{WV}{L \times WI} = \frac{7}{L \times WI} 2. The equations for L, WI, and H assume manure is stacked to average height equal to the sidewall height. Available storage volume must be adjusted to account for these types of variations. 22. Rectangular tank dimensions Total height, ft (H) = \frac{VV}{VVV} Selected width, ft (WI) = \frac{VVV}{VVV}
Waste stacking structure sizing 17. Structure length, ft L = \frac{WV}{WI \times H} = \frac{40}{L \times H} 18. Structure width, ft WI = \frac{WV}{L \times H} = \frac{40}{L \times H} Notes for waste stacking structure: 1. The volume determined (WV) does not include any volume for freeboard. It is recommended that a minimum of 1 foot of freeboard be provided for a waste stacking structure. Tank sizing 20. Effective depth, ft. (EH) Total height (or depth) of tank desired, ft (H) Less precipitation for storage period, ft	(USE 90) 19. Structure height, ft H = \frac{WV}{L x WI} = \frac{7}{L x WI} 2. The equations for L, WI, and H assume manure is stacked to average height equal to the sidewall height. Available storage volume must be adjusted to account for these types of variations. 22. Rectangular tank dimensions Total height, ft (H) = \frac{SA}{WI} = \frac{SA}{WI}
16. Waste volume, ft³ (WV) = TVM + TWW + TBV = 22,40 Waste stacking structure sizing 17. Structure length, ft L = WV WI x H 18. Structure width, ft WI = WV L x H Notes for waste stacking structure: 1. The volume determined (WV) does not include any volume for freeboard. It is recommended that a minimum of 1 foot of freeboard be provided for a waste stacking structure. Tank sizing 20. Effective depth, ft. (EH) Total height (or depth) of tank desired, ft (H) Less precipitation for storage period, ft. — - (uncovered tanks only)	(USE 90) 19. Structure height, ft H = \frac{WV}{L \times WI} = \frac{7}{L \times WI} = \frac{7}{L \times WI} = \frac{1}{L \ti
16. Waste volume, ft³ (WV) = TVM + TWW + TBV = 22,40 Waste stacking structure sizing 17. Structure length, ft L = \frac{WV}{WI \times H} = 40 18. Structure width, ft WI = \frac{WV}{L \times H} = 40 Notes for waste stacking structure: 1. The volume determined (WV) does not include any volume for freeboard. It is recommended that a minimum of 1 foot of freeboard be provided for a waste stacking structure. Tank sizing 20. Effective depth, ft. (EH) Total height (or depth) of tank desired, ft (H)	(USE 90) 19. Structure height, ft H = \frac{WV}{L x WI} = \frac{7}{L
Waste stacking structure sizing 17. Structure length, ft L = \begin{array}{c} WV & WV	22. The equations for L, WI, and H assume manure is stacked to average height equal to the sidewall height. Available storage volume must be adjusted to account for these types of variations. 22. Rectangular tank dimensions Total height, ft (H) = Selected width, ft (WI)= Length, ft L = SA = 23. Circular tank dimensions Total height, ft H =

(b) Liquid and slurry waste storage

Liquid and slurry manure can be stored in waste storage ponds or in aboveground or below-ground tanks. Solids separation of manure and bedding is a problem that must be considered in planning and design. Solids generally can be resuspended with agitation before unloading, but this involves a cost in time, labor, and energy. Another option allows solids to accumulate if the bottom is occasionally cleaned. This requires a paved working surface for equipment.

Earthen storage is frequently the least expensive type of storage; however, certain restrictions, such as limited space availability, high precipitation, water table, permeable soils, or shallow bedrock, can limit the types of storage considered.

Storage ponds are earthen basins designed to store wastewater and manure (figs. 10–14, 10–15, 10–16). They generally are rectangular, but may be circular or any other shape that is practical for operation and

maintenance. The inside slopes range from 1.5 to 1 (horizontal to vertical) to 3 to 1. The combined slopes (inside plus outside) should not be less than 5 to 1 for embankments. The soil, safety, and operation and maintenance need to be considered in designing the slopes. The minimum top width of embankments should be 8 feet; however, greater widths should be provided for operation of tractors, spreaders, and portable pumps.

Storage ponds should provide capacity for normal precipitation and runoff (less evaporation) during the storage period. Appendix 10C provides a method for determining runoff and evaporation volumes. A minimum of 1 foot of freeboard is provided.

Inlets to storage ponds can be of any permanent material designed to resist erosion, plugging, or, if freezing is a problem, damage by ice. Typical loading methods are pipes and ramps, which are described in section 651.1005. Flow of wastes away from the inlet should be considered in selecting the location of the inlet

Figure 10–14 Cross section of waste storage pond without a watershed

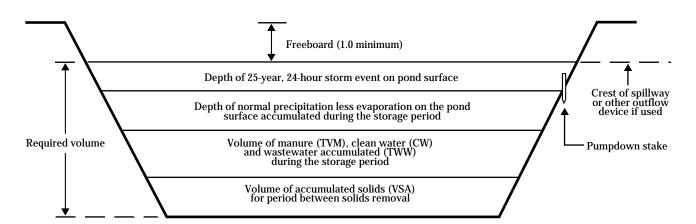
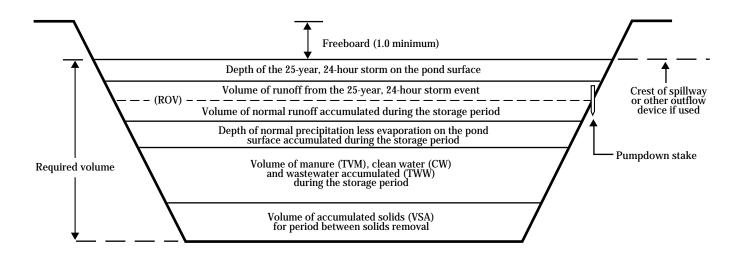
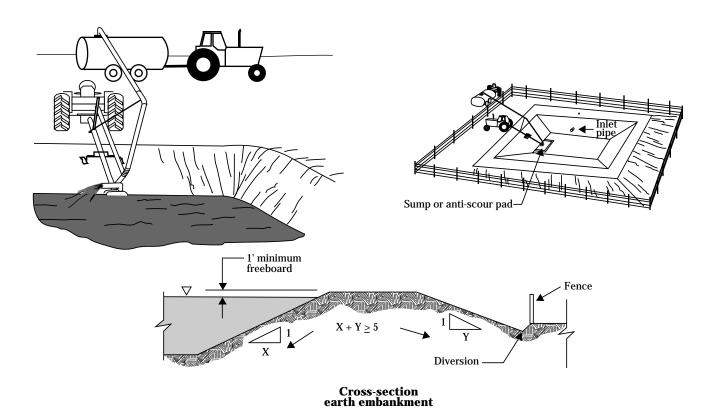


Figure 10-15 Cross section of waste storage pond with watershed



*or other outflow device

Figure 10-16 Waste storage ponds



(210-vi-AWMFH, rev. 1, July 1996)

Gravity pipes, pumping platforms, and ramps are used to unload storage ponds. A method for removing solids should be designed for the storage pond. If the wastes will be pumped, adequate access must be provided to thoroughly agitate the contents of the pond. A ramp should have a slope of 8 to 1 or flatter and be wide enough to provide maneuvering room for unloading equipment.

Pond liners are used in many cases to compensate for site conditions or improve operation of the pond. Concrete, geomembrane, and clay linings reduce permeability and can make an otherwise unsuitable site acceptable. See Appendix 10D, Geotechnical design and construction guidelines for waste impoundment Liners, for detail on clay liners. Concrete also provides a wear surface if unloading equipment will enter the pond.

Figures 10–17, 10–18, and 10–19 represent various kinds of storage ponds and tanks.

Liquid manure can be stored in aboveground (fig. 10-18) or below-ground (fig. 10-19) tanks. Liquid manure storage tanks can be constructed of metal,

concrete, or wood. Below-ground tanks can be loaded using slatted floors, push-off ramps, gravity pipes or gutters, or pumps. Aboveground tanks are typically loaded by a pump moving the manure from a reception pit. Tank loading can be from the top or bottom of the tank depending on such factors as desired agitation, minimized pumping head, weather conditions, and system management.

Storage volume requirements for tanks are the same as those for ponds except that provisions are normally made to exclude outside runoff from waste storage tanks because of the relative high cost of storage. Of course, if plans include storage of outside runoff, accommodation for its storage must be included in the tank's volume.

Tanks located beneath slatted floors can sometimes be used for temporary storage with subsequent discharge into lagoons or other storage facilities. Recycled lagoon effluent is added to a depth of 6 to 12 inches in underslat pits to reduce tendency for manure solids to stick to the pit floor. Wastes are allowed to collect for several days, typically 1 to 2 weeks, before the pits are gravity drained.

Figure 10–17 Layout of waste storage pond

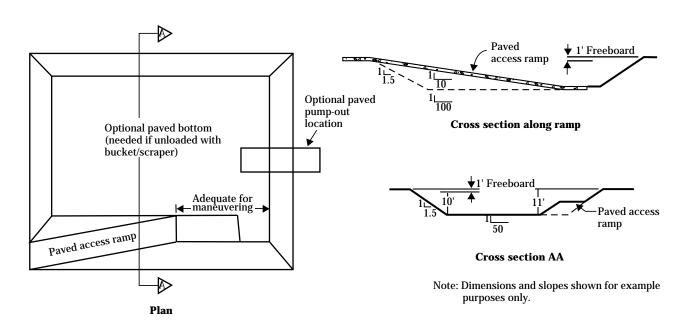


Figure 10-18 Aboveground waste storage tank

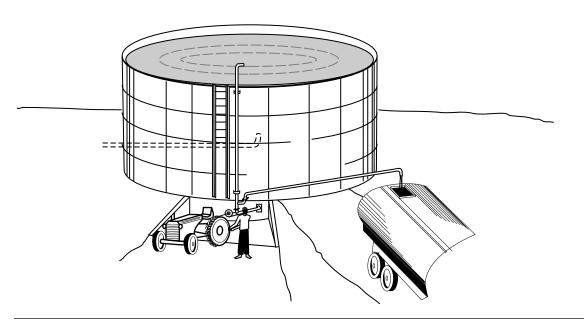
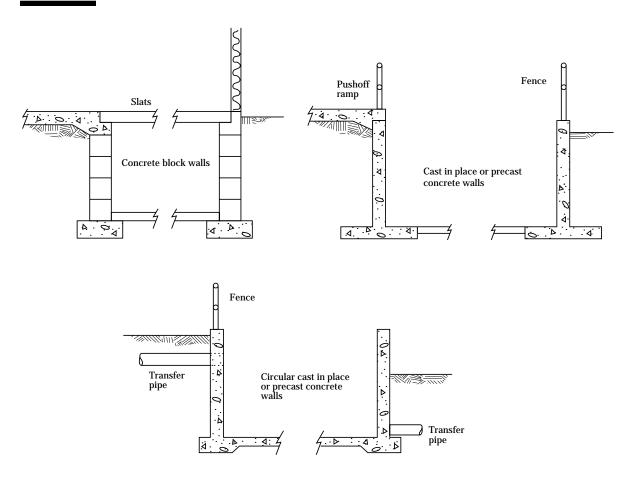


Figure 10–19 Below-ground waste storage structure



Agricultural Waste Management System Component Design

Part 651 Agricultural Waste Management Field Handbook

(1) Design considerations

Tank material types—The primary materials used to construct manure tanks are reinforced concrete, metal, and wood. Such tanks must be designed by a professional engineer and constructed by experienced contractors. A variety of manufactured, modular, and cast-in-place tanks are available from commercial suppliers. NRCS concurs in the standard detail drawings for these structures based on a review and approval of the drawings and supporting design calculations. A determination must be made that the site conditions are compatible with the design assumptions on which the design is based. Structures can also be designed on an individual site-specific basis.

Cast-in-place, reinforced concrete, the principal material used in below-ground tanks, can be used in above-ground tanks as well. Tanks can also be constructed of precast concrete panels that are bolted together. Circular tank panels are held in place with metal hoops. The panels are positioned on a concrete foundation or have footings cast as an integral part of the panel. Tank floors are cast-in-place slabs.

Other above-ground tanks are constructed of metal. Glass-fused steel panels are widely used. Such tanks are manufactured commercially and must be constructed by trained crews. Other kinds of metal panels are also used.

At least one company offers a wooden above-ground tank for liquid storage. The preservative treated boards have tongue-and-groove edges and are held in place using metal hoops similar to those used for concrete panel tanks. All manure tanks should meet the standards identified in the section on solid manure storage.

Sizing—Liquid waste storage ponds and structures should be sized to hold all of the manure, bedding, wastewater from milkhouse, flushing, and contaminated runoff that can be expected during the storage period. Equation 10–3 can be used to compute the waste volume:

$$WV = TVM = TWM = TBV$$
 [10-3]

where:

WV = Waste volume for storage period, ft³

TVM = Total volume of manure for storage period, ft³ (see equation 10–1)

TWW= Total wastewater volume for storage period, ft³

TBV = Total bedding volume for storage period, ft³ (see equation 10–2)

Data on wastewater production are available in chapter 4 or from the farmer or rancher. Appendix 10C provides a method of estimating contaminated runoff volume.

In addition to the waste volume, waste storage tanks must, if uncovered, provide a depth to accommodate precipitation less evaporation on the storage surface during the most critical storage period. The most critical storage period is generally the consecutive months that represent the storage period that gives the greatest depth of precipitation less evaporation. Appendix 10C gives a method for estimating precipitation less evaporation. Waste storage tanks must also provide a depth of 0.5 feet for material not removed during emptying. A depth for freeboard of 0.5 feet is also recommended.

Waste storage ponds must also provide a depth to accommodate precipitation less evaporation during the most critical storage period. If the pond does not have a watershed, the depth of the 25-year, 24-hour precipitation on the pond surface must be included. Appendix 10B includes a map giving the precipitation amount for the 25-year, 24-hour precipitation. Frequently, waste storage ponds are designed to include outside runoff from watersheds. For these, the runoff volume of the 25-year, 24-hour storm must be included in the storage volume.

Appendix 10C gives a procedure for estimating the runoff volume from feedlots. The NRCS Engineering Handbook for Conservation Practice, chapter 2, may be used to estimate runoff volumes for other watershed areas.

Chapter 10	Agricultural Waste Management System
	Component Design

(2) Design of sidewalls and floors

The information on the design of sidewalls and floors in section 651.1003(a) on solid manure storage material is applicable to these items used for liquid manure storage. All possible influences, such as internal and external hydrostatic pressure, flotation and drainage, live loads from equipment and animals, and dead loads from covers and supports, must be considered in the design.

Pond sealing—Waste storage ponds must not allow excess seepage. The soil in which the pond is to be located must be evaluated and, if needed, tested during planning and design to determine need for an appropriate liner. Refer to Appendix 10D, Geotechnical design and construction guidelines for waste impoundment liners, for detail on determining need for and design of clay liners. Also refer to Chapter 7, Geology and Ground Water Considerations, for more information on site evaluation, investigations, and testing.

(3) Design example 10-3—Waste storage tank Mr. Bill Walton of Middlesburg, Tennessee, has requested assistance on a waste management system. The selected alternative includes a below-ground, covered, slurry storage tank for his Holstein dairy herd. He has 150 milkers that average 1,400 pounds and 75 heifers that are about 1,000 pounds each. Bedding material is not used with these animals. Based on crop utilization of the nutrients, storage is needed for 75 days. The critical storage periods are January 1 to March 15 and July 1 to September 15. The wash water from the milkhouse and parlor is also stored. No runoff will be directed to the storage. Worksheet 10A-1 shows how to determine the necessary volume for the storage tank and several possible sets of tank dimensions. It also shows how to estimate

Manure production—The animal type, average weight, and number are entered on lines 1, 2, and 3. The equivalent 1,000 pound animal units (AU) for the animal type is calculated and entered on line 4. The daily volume of manure (DVM) production for each animal type is selected from table 4–5 and entered on line 5. The storage period (D) is entered on line 6. The total manure volume (VMD) is calculated for each animal type and entered on line 7. Add the VMD for each animal type and enter the sum (TVM) on line 8.

the total solids content of the stored waste.

Wastewater volume—The daily wastewater volume per animal unit description (DWW) is selected from table 4–6 and entered on **line 9.** The wastewater volume for the animal type for the storage period (WWD) is calculated and entered on **line 10.** Add the wastewater volumes for each animal type and enter the sum (TWW) on **line 11.**

Bedding volume—Bedding is not used in this example. If bedding were used, however, its volume for the storage period would be determined using **lines** 12 through 15.

Waste volume—WV is the total volume of waste material that will be stored including total manure (TVM), total wastewater (TWW), and total bedding volume (TBV). Provisions are to be made to assure that outside runoff does not enter the tank. In addition, if the tank is not covered, the depth of precipitation less evaporation on the tank surface expected during the most critical storage period must be added to the depth requirements.

Total depth available—The desired depth is the total planned depth based on such considerations as foundation condition, tank wall design, and standard drawing depth available.

Surface area—The surface area **(line 21)** dimensions are calculated using the equation for SA.

Tank dimensions—Because tanks are rectangular or circular, various combinations of length and width can be used to provide the SA required. If the depth is held constant, only one solution for the diameter of a circular tank is possible. The dimensions of either shape can be rounded upward to match a standard detail drawing or for convenience.

Total solids content—The initial TS content of the manure is given in table 4–5 in chapter 4. Because there are two sources of manure, the solids content of the total manure must be weighted by the contribution from each animal type. The adjusted total solids content of the stored manure is determined from figure 10–40 using the added water from the milkhouse and parlor, the runoff (none in this example), and the net rainfall during the storage period. Because the total solids content of milking center wastewater is so low, it can be ignored.

Initial TS =
$$\frac{(12.5\% \times 210\text{AU}) + (10.7\% \times 75\text{AU})}{210\text{AU} + 75\text{AU}}$$
$$= 12\%$$

Added water:

$$[9,450 \text{ ft}^3 + (0.3 \text{ ft} \times 33,580 \text{ ft}^3)] \times 7.48 \text{ gal} / \text{ft}^3$$

= 78,720 gal

Added water/ft3 manure:

$$\frac{78,20}{20,472+7,313} = 2.8 \text{ gal } / \text{ ft}^3$$

From figure 10-40, adjusted TS = 8.8%.

Completed worksheet for Design example 10-3

Worksheet 10A-1—Waste	storage structure capacity design
Decisionmaker: Bill Walton	Date: 6/13/87
Middlesburg, TN	
Animal units	
	3. Number of animals (N)
2. Animal weight, lbs (W)1,4001,000	4. Animal units, AU = W x N =
Manure volume	7. Total volume of manure production for animal type for storage period, ft ³ = 20,475 7,312 WMD = AU x DVM x D = 20,475 7,312 8. Total manure production for storage period, ft ³ (TVM) 27,78
Wastewater volume	
9. Daily wastewater volume per AU, ft³/AU/day (DWW) = 0.6 0 10. Total wastewater volume for animal description for storage period, ft³ 9,450 0 0	
Bedding volume	
12. Amount of bedding used daily for animal type, lbs/AU/day (WB) =	14. Bedding volume for animal type for storage period, ff ³ =
13. Bedding unit weight, lbs/fb ³ (BUW) =	
	15. Total bedding volume for storage Oreriod, ft ³ (TBV) =
Minimum waste storage volume requirement 16. Waste storage volume, ft ³ (WV) = TVM + TWW + TBV = Waste stacking structure sizing	27,787 ₊ 9,450 ₊ 0 = <u>37,2</u>
	19_Structure beight ft H WV
18. Structure width, ft WI =WV =	19. Structure height, ft H = WV L x WI =
Notes for waste stacking structure:	
1. The volume determined (WSV) does not include any volume for freeboard. It is recommended that a minimum of 1 foot of freeboard be provided for a waste stacking structure.	The equations for L, WI, and H assume manure is stacked to average height eq to the sidewall height. Available storage volume must be adjusted to account for these types of variations.
Tank sizing	22. Rectangular tank dimensions Total height, ft (H) = $\frac{12}{12}$ Selected width, ft (WI) = $\frac{30}{12}$
20. Effective depth, ft. (EH)	9
Total height (or depth) of tank desired, ft (H) 0	Length, ft L = <u>SA</u> = <u>112.8</u> (USE 115)
(uncovered tanks only) Less depth allowance for accumulated solids, ft – 0.5	23. Circular tank dimensions Total height, ft H =12
(0.5 ft. minimum) Less depth for freeboard (0.5 ft. recommended), ft	Diameter, ft DIA = (1.273 x SA)0.5 = <u>65.6</u> (USE 6
Effective depth, ft (EH) =11	Notes for waste storage tank structure: 1. Final dimensions may be rounded up to whole numbers or to use
21. Surface area required, ft ² SA = $\frac{WV}{E H}$ = $\frac{3,385}{E}$	increments on standard drawings. 2. Trial and error may be required to establish appropriate dimensions.

(4) Design example 10-4—Waste storage pond

Mr. Joe Green of Silverton, Oregon, has requested assistance in developing an agricultural waste management system for his dairy. He has selected an alternative that includes a waste storage pond component. He has a Holstein herd composed of 500 milkers averaging 1,400 pounds; 150 dry cows averaging 1,400 pounds; and 150 heifers averaging 1,000 pounds. He has a freestall barn that has flush alleys. He uses foam pads for bedding. The alternative selected includes land application. A storage period of 180 days is required for storage through the winter months of high precipitation. A solid separator will be used to minimize solid accumulation in the waste storage pond and to allow recycling of the flush water. Water from the milkhouse and parlor will be stored in the pond. Use worksheet 10A-2 to determine the required capacity and size of the pond.

Manure production—The animal type, average weight, and numbers are entered on lines 1, 2, and 3. The number of 1,000 pound animal units for each animal type (AU) is calculated and entered on line 4. The volume of daily manure production (DVM) from table 4–5 is entered on line 5. The storage period (D) is entered on line 6. The manure volume for the storage period for each animal type (VMD) is then calculated and entered on line 7. The total volume (TVM) is added and then entered on line 8.

Wastewater volume—In this example, only the wastewater from the milkhouse and parlor is accounted for in the waste storage volume requirements because the alley flush water is recycled. The daily wastewater volume per animal unit (DWW) from table 4-6 is entered on line 9. The wastewater volume for each animal type for the storage period (WWD) is calculated using the equation and entered on line 10. The wastewater volume from each animal type (WWD) is added, and the sum (TWW) is entered on line 11.

Clean water volume—In this example, no clean water is added. However, if clean water (CW) is added for dilution, for example, the amount added during the storage period would be entered on **line 12**.

Runoff volume—For this example, the waste storage pond does not have a watershed and storage for runoff is not needed. However, waste storage ponds are frequently planned to include the runoff from a watershed, such as a feedlot. The ponds that have a watershed must include the normal runoff for the storage period and the runoff volume for the 25-year, 24-hour storm. The runoff volume from feedlots may be calculated using the procedures in appendix 10C. For watersheds or parts of watersheds that have cover other than feedlots, the runoff volume may be determined using the procedure in chapter 2 of the Engineering Field Manual for Conservation Practices. The value for watershed runoff volume (ROV) is entered on line 13. Documentation showing the procedure and values used in determining the volume of runoff should be attached to the worksheet.

Volume of accumulated solids—This volume is to accommodate the storage of accumulated solids for the period between solids removal. The solids referred to are those that remain after the liquid has been removed. An allowance for accumulated solids is required mainly for ponds used to store wastewater and polluted runoff. Solids separation, agitation before emptying, and length of time between solids removal all affect the amount of storage that must be provided. Enter the value for accumulated solids (VSA) on line 14. In this example, the solids from the manure are separated and solids accumulation will be minimal. No storage is provided for accumulated solids.

Waste volume—The total waste storage volume (WV) is determined by adding the total volume of manure (TVM), total wastewater volume (TWW), clean water added (CW), and volume allowance for solids accumulation (VSA). Waste storage ponds that have a watershed must also include the normal runoff volume for the storage period and the volume of the 25-year, 24-hour storm runoff (ROV). WSV is calculated on line 15. The waste storage pond must be sized to store this volume plus additional depth as explained in "depth adjustment."

Completed worksheet for Design example 10-4

Decisionmaker:	loe Green			Date:	10/4/9	0	
Site: Silve	erton, OR						
Animal units							
Animal type	Milkers D	ry Heifers	3. Number of animals (N)		500	150	1
2. Animal weight, lbs (400 1,000	4. Animal units. AU = W		700	210	
Manure volume							
 Daily volume of ma per AU, ft³/AU/day 	nure production (DVM)= 1.30	1.30 1.30	Total volume of manure animal type for storage	e production fo e period, ft ³	r <u>163,800</u>	49,140	35
	ys (D) =		VMD = AU x DVM x D 8. Total manure production	= on for storage	period, ft ³ (TVM))	<u>248</u>
Wastewater vo			11. Total wastewater volum	o for			
 Daily wastewater v AU, ft³/AU/day (E 	volume per 0.6)WW) =	0 0	11. Iotal wastewater volum storage period, ft ³ (TW	/W)			75,
10. Total wastewater v description for storag WWD = DWW x AU x	e period, ft ³ 75 / 00)					
Clean water vol	ume	0	Runoff Volume				
12. Clean water added	during storage period, f	ft ³ (CW)	13. Runoff volume, ft3 (Includes the volume of r				
Solids accumul	ation	_	due to normal runoff for				
		\circ	runoff volume from the				
Waste volume 15. Waste volume, ft ³	(WV) = TVM +	+ TWW + CW + ROV + V	runoff volume from the	25-year, 24-ho	our storm.	23,640	
Waste volume 15. Waste volume, ft ³ Pond sizing 16. Sizing by tria Side slope ratio, Rectangular pond,	requirement (WV) = TVM + = 248 I and error (Z) = 3	+ TWW + CW + ROV + V: B,040 + 75,600	SA $0 + 0 + 0$ or greater than WV = 32 Circular pond	25-year, 24-hc 25-year, 24-hc 3,640 ft ³	our storm. O = 32		
Waste volume 15. Waste volume, ft ³ Pond sizing 16. Sizing by tria Side slope ratio, Rectangular pond, $V = \left(\frac{4 \times Z^2 \times d^3}{3}\right)$ Trial Bottom width no. ft (BW)	requirement $(WV) = TVM + \frac{248}{}$ $= 248$ I and error $(Z) = 3$ $-(Z \times BL \times d^2) + (Z \times BL \times d^2)$ Bottom length perfect (BL) fit	+ TWW + CW + ROV + V: $8,040 + 75,600$ $V \text{ must be equal to } 0$ $8W \times d^2 + (BW \times BL \times BL)$ $\text{epth*} \text{ Volume}$ $\text{t (d)} \text{ ft}^3 \text{ (V)}$	SA $0 + 0 + 0$ or greater than WV = 32 Circular pond	25-year, 24-hc 25-year, 24-hc 3,640 ft ³	O = 32	x d ²) + (0.7 Depth* ft (d)	Vo
Waste volume 15. Waste volume, ft ³ Pond sizing 16. Sizing by tria Side slope ratio, Rectangular pond, $V = \left(\frac{4 \times Z^2 \times d^3}{3}\right) + C$ Trial Bottom width	requirement $(WV) = TVM + \frac{248}{4}$ I and error $(Z) = \frac{3}{4}$ Bottom length of t (BL) first (BL) first (BL) for the second of t (BL) for the s	+ TWW + CW + ROV + V: $\frac{8,040}{0} + \frac{75,600}{0}$ $W = V \text{ must be equal to } 0$ $\frac{3W \times d^2}{0} + (BW \times BL \times BU)$ $\text{epth*} \text{Volume}$	SA $\frac{O}{O} + \frac{O}{O} + \frac{O}{O}$ or greater than WV = $\frac{32}{O}$ Circular pond $V = (1.05 \times 20)$ Trial	$\frac{25 - \text{year}, 24 - \text{ho}}{25 - \text{year}, 24 - \text{ho}}$ $\frac{23,640}{1}$ ft ³ I, $\frac{22 \times d^3}{1}$ + (O = 32	x d ²) + (0.7 Depth* ft (d)	Vo ft
Waste volume 15. Waste volume, ft ³ Pond sizing 16. Sizing by tria Side slope ratio, Rectangular pond, $V = \left(\frac{4 \times Z^2 \times d^3}{3}\right) + \frac{1}{3}$ Trial Bottom width no. ft (BW) 1 100 2 100 3 100 4 100 * Depth must be adjusted.	requirement $(WV) = TVM + \frac{248}{248}$ I and error $(Z) = \frac{3}{48}$ Bottom length Defit (BL) fit \frac{500}{400} \frac{6}{425} \frac{6}{425} \frac{6}{425} \frac{1}{425}	+ TWW + CW + ROV + V: $(3,040) + 75,600$ — V must be equal to $(3,040) + (3,040) + (3,040)$ epth* Volume t (d) $(3,040) + (3,040) + (3,040)$ 6 367,392 6 296,592 6 314,292	SA $O + O + ($ or greater than WV = $\frac{32}{}$ Circular pond $V = (1.05 \times 2)$ Trial $no.$	$\frac{25 - \text{year}, 24 - \text{ho}}{25 - \text{year}, 24 - \text{ho}}$ $\frac{23,640}{1}$ ft ³ I, $\frac{22 \times d^3}{1}$ + (O = 32	x d ²) + (0.7 Depth* ft (d)	Vo
Waste volume 15. Waste volume, ft ³ Pond sizing 16. Sizing by tria Side slope ratio, Rectangular pond, $V = \left(\frac{4 \times Z^2 \times d^3}{3}\right) + \frac{1}{3}$ Trial Bottom width no. ft (BW) 1 100 2 100 3 100 4 100 4 100 * Depth must be adjusted.	requirement $(WV) = TVM + \frac{248}{2}$ I and error $(Z) = \frac{3}{4}$ Bottom length be fit (BL) fit $\frac{500}{400}$ ($\frac{425}{425}$ ($\frac{100}{400}$ ($\frac{100}{425}$ ($\frac{100}{400}$ (+ TWW + CW + ROV + V: $(3,040) + 75,600$ — V must be equal to $(3,040) + (3,040) + (3,040)$ epth* Volume t (d) $(3,040) + (3,040) + (3,040)$ 6 367,392 6 296,592 6 314,292	SA $O + O + ($ or greater than WV = $\frac{32}{}$ Circular pond $V = (1.05 \times 2)$ Trial $no.$	$\frac{25 - \text{year}, 24 - \text{ho}}{25 - \text{year}, 24 - \text{ho}}$ $\frac{23,640}{1}$ ft ³ I, $\frac{22 \times d^3}{1}$ + (O = 32	x d ²) + (0.7 Depth* ft (d)	Vo
Waste volume 15. Waste volume, ft ³ Pond sizing 16. Sizing by tria Side slope ratio, Rectangular pond, $V = \left(\frac{4 \times Z^2 \times d^3}{3}\right) + \frac{1}{3}$ Trial Bottom width no. ft (BW) 1 100 2 100 3 100 4 100 * Depth must be adjusted 17. Depth adjustme 17. Depth adjustme	requirement $(WV) = TVM + \frac{248}{248}$ I and error $(Z) = \frac{3}{48}$ Bottom length be fit (BL) fit $\frac{500}{400}$ $\frac{3}{425}$	+ TWW + CW + ROV + V: B,040 + 75,600 — V must be equal to a BW×d²)+(BW×BL× epth* Volume t (d) ft³ (V) 6 367,392 6 296,592 6 314,292 6 326,903	SA $O + O + ($ or greater than WV = $\frac{32}{}$ Circular pond $V = (1.05 \times 2)$ Trial $no.$	$\frac{25 - \text{year}, 24 - \text{ho}}{25 - \text{year}, 24 - \text{ho}}$ $\frac{23,640}{1}$ ft ³ I, $\frac{22 \times d^3}{1}$ + (O = 32	x d ²) + (0.7 Depth* ft (d)	Vo
Waste volume 15. Waste volume, ft ³ Pond sizing 16. Sizing by tria Side slope ratio, Rectangular pond, $V = \left(\frac{4 \times Z^2 \times d^3}{3}\right) + \frac{1}{3}$ Trial Bottom width no. ft (BW) 1 100 2 100 3 100 4 100 * Depth must be adjusted 17. Depth adjustme 17. Depth adjustme Add depth of pre	requirement $(WV) = TVM + \frac{248}{2}$ I and error $(Z) = \frac{3}{4}$ Bottom length Deft (BL) fit (BL) fit (BL) fit (AD) (AD) (AD) (AD) (AD) (AD) (AD) (AD)	+ TWW + CW + ROV + V: B,040 + 75,600 V must be equal to a BW×d²)+(BW×BL× epth* Volume t (d) ft³ (V) 6 367,392 6 296,592 6 314,292 6.2 326,903	SA $O + O + ($ or greater than WV = $\frac{32}{}$ Circular pond $V = (1.05 \times 2)$ Trial $no.$	25-year, 24-ho 25-year, 24-ho 3,640 ft ³ I, Z ² x d ³) + (Bottom dia (DIA)	O = 32	x d ²) + (0.7 Depth* ft (d)	Vo ft:
Waste volume 15. Waste volume, ft ³ Pond sizing 16. Sizing by tria Side slope ratio, Rectangular pond, $V = \left(\frac{4 \times Z^2 \times d^3}{3}\right) + \frac{1}{3}$ Trial Bottom width no. ft (BW) $\frac{1}{2} = \frac{100}{100}$ $\frac{3}{3} = \frac{100}{100}$ * Depth must be adjusted 17. Depth adjust Depth, ft (d) Add depth of pre (For the storage)	requirement $(WV) = TVM + \frac{248}{2}$ I and error $(Z) = \frac{3}{4}$ Bottom length Deft (BL) fit (BL) fit (BL) fit (AD) (AD) (AD) (AD) (AD) (AD) (AD) (AD)	+ TWW + CW + ROV + V: B,040 + 75,600 V must be equal to 0 BW×d²)+(BW×BL× epth* Volume t (d) fi³ (V) 6 367,392 6 296,592 6 314,292 6.2 326,903	SA $\frac{O}{O} + \frac{O}{O} + \frac{O}{O}$ or greater than WV = $\frac{32}{O}$ Circular pond $\frac{O}{O} = \frac{1.05 \times 3}{O}$ Trial $\frac{O}{O} = \frac{O}{O}$ $\approx \frac{O}{O} = \frac{O}{O}$	25-year, 24-ho 25-year, 24-ho 3,640 ft ³ I, Z ² x d ³) + (Bottom dia (DIA)	O = 32	x d ²) + (0.7 Depth* ft (d)	Vo ft:

Waste storage pond sizing—The waste storage pond is sized by trial and error for either a rectangular or circular shaped pond by using the procedure on **line 16**. Figure 10-20 is a simple BASIC computer program that can be used to compute the volume by inputting the bottom width, bottom length, and depth.

Depth adjustment—The depth required to store the waste storage volume with the selected pond dimensions must be adjusted by adding depth for the precipitation less evaporation and the depth of the 25-year, 24-hour storm on the pond surface. The minimum freeboard is 1 foot. The adjustment for final depth is made using **line 17**.

Figure 10-20 BASIC computer program for determining pond volume

o :			: o
0:	100	REM********	: 0 : 0
0:	110	REM * BASIC program for solving the rectangular pond volume *	: 0
o :	120	DEM * equation *	: o
0:		REM*********	: 0
0:	130		: 0 : 0
0:	140	INPUT "Side Slope Ratio, Z";Z	: 0
0:	150	INPUT "Trial No.";T	: o
o :	160	INPUT "Trial Bottom Width, BW";W	: o
0:	170	INPUT "Trial Bottom Length, BL";L	: 0 : 0
0:	180	INPUT "Trial Depth, d";D	: 0
o :	190	$V = (W^*L^*D) + (Z^*D^2L) + (Z^*D^2W) + ((4^*Z^2D^3)/3)$: o
o :	200	PRINT "V = ";V;"cubic feet"	: o
0:			: 0 : 0
0:	210	GOTO 150	: 0
o :	220	END	: o
o :			: 0
0:			: 0 : 0
0:			: 0
o :			: o
0:	100	REM ********	: 0
0:	110	REM * BASIC program for solving the circular pond volume *	: 0 : 0
0:	120		: 0
o :		REM * equation	: o
0:	130		: 0
0:	140	INPUT "Side Slope Ratio, Z";Z	: 0 : 0
0:	150	INPUT "Trial No.";T	: 0
o :	160	INPUT "Trial Bottom Diameter, DIA";W	: o
0:	170	INPUT "Trial Depth, d";D	: 0
0:	180	$V = (1.05*Z^2*D^3) + (1.57*W*Z*D^2) + (.79*W^2*D)$: 0 : 0
0:	190	PRINT "V = ";V;"cubic feet"	: 0
o :	200	GOTO 150	: o
0:			: 0
0:	210	END	: 0 : 0
0:			: 0
0:			: 0

651.1004 Treatment

In many situations it is necessary to treat agricultural waste before final utilization. The purpose of treatment is to reduce pollution potential of the waste through biological, physical, and chemical processes using such components as lagoons, oxidation ditches, and composting. These types of components reduce nutrients, destroy pathogens, and reduce total solids. Composting also reduces the volume of the waste. Treatment also includes any step that might be considered pretreatment, such as solids separation, drying, and dilution that prepares the waste for facilitating another function. By their nature, treatment facilities require a higher level of management than that of storage facilities.

(a) Anaerobic lagoons

Anaerobic lagoons are widely accepted in the United States for the treatment of animal waste. Anaerobic treatment of animal waste helps to protect water quality by reducing much of the organic concentration (BOD, COD) of the waste. Anaerobic lagoons also reduce the nitrogen content of the waste through ammonia volatilization and effectively reduce animal waste odors if the lagoon is managed properly.

(1) Design

The maximum operating level of an anaerobic lagoon is a volume requirement plus a depth requirement. The volume requirement is the sum of the following volumes:

- Minimum treatment volume, ft³ (MTV)
- Manure volume, wastewater volume, and clean water, ft³ (WV)
- Sludge volume, ft³ (SV)

The depth requirement is the normal precipitation less evaporation on the lagoon surface.

Polluted runoff from a watershed must not be included in a lagoon unless a defensible estimate of the volatile solid loading can be made. Runoff from a watershed, such as a feedlot, is not included in a lagoon because loading would only result during storm events and because the magnitude of the loading would be difficult, if not impossible, to estimate. As a result, the lagoon would be shocked with an overload of volatile solids.

If an automatic outflow device, pipe, or spillway is used, it must be placed at a height above the maximum operating level to accommodate the 25-year, 24-hour storm precipitation on the lagoon surface. This depth added to the maximum operating level of the lagoon establishes the level of the required volume or the outflow device, pipe, or spillway. A minimum of 1 foot of freeboard is provided above the outflow and establishes the top of the embankment. Should state regulation preclude the use of an outflow device, pipe, or spillway or if for some other reason the lagoon will not have these, the minimum freeboard is 1 foot above the top of the required volume.

The combination of these volumes and depths is illustrated in figure 10–21. The terms and derivation are explained in the following paragraphs.

Anaerobic waste treatment lagoons are designed on the basis of volatile solids loading rate (VSLR) per 1,000 cubic feet. Volatile solids represent the amount of solid material in wastes that will decompose as opposed to the mineral (inert) fraction. The rate of solids decomposition in anaerobic lagoons is a function of temperature; therefore, the acceptable VSLR varies from one location to another. Figure 10-22 indicates the maximum VSLR's for the United States. If odors need to be minimized, VSLR should be reduced by 25 to 50 percent.

The minimum treatment volume (MTV) represents the volume needed to maintain sustainable biological activity. The minimum treatment volume for VS can be determined using equation 10–4.

$$MTV = \frac{TVS}{VSLR}$$
 [10-4]

where:

MTV = Minimum treatment volume, ft³

TVS = Total daily volatile solids loading (from all sources), lb/day

VSLR = Volatile solids loading rate, lb/1,000 ft³/day (from fig. 10–22)

Daily volatile solids production for various wastes can be determined using tables in chapter 4. If feed spillage exceeds 5 percent, *VSP* should be increased by 4 percent for each additional 1 percent spillage.

Waste volume (WV) should reflect the actual volume of manure, wastewater, flush water that will not be recycled, and clean dilution water added to the lagoon during the treatment period. The treatment period is either the detention time required to obtain the desired reduction of pollution potential of the waste or the time between land application events, whichever is longer. State regulations may govern the minimum detention time. Generally, the maximum time between land application events determines the treatment period because this time generally exceeds the detention time required.

$$WV = TVM + TWW + CW$$
 [10-5]

where:

WV = Waste volume for treatment period, ft³

TVM = Total volume of manure for treatment period. ft³

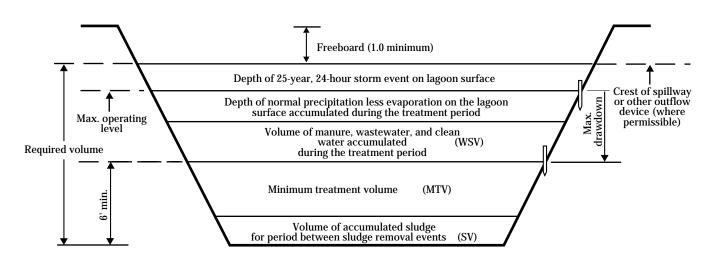
TWW = Total volume of wastewater for treatment period, ft³

CW = Clean water added during treatment period, ft³

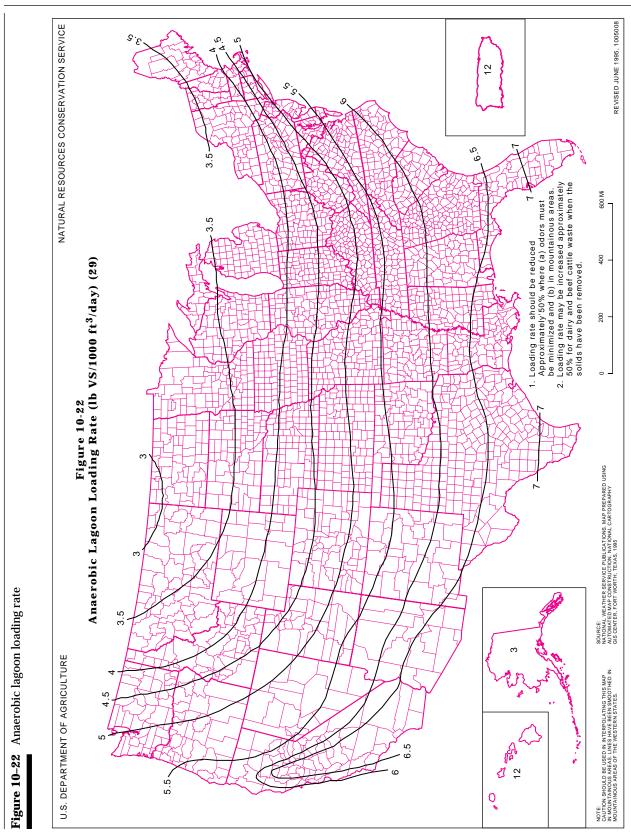
In the absence of site-specific data, values in chapter 4 may be used to make estimates of the volumes.

As the manure is decomposed in the anaerobic lagoon only part of the total solids (TS) is reduced. Some of the TS is mineral material that will not decompose, and some of the VS require a long time to decompose. These materials, referred to as sludge, gradually accumulate in the lagoon. To maintain the minimum treatment volume (MTV), the volume of sludge accumulation over the period of time between sludge removal

Figure 10–21 Anaerobic lagoon cross section



Note: The minimum treatment volume for an anaerobic waste treatment lagoon is based on volatile solids.



must be considered. Lagoons are commonly designed for a 15- to 20-year sludge accumulation period. The sludge volume (SV) can be determined using equation 10–6.

$$SV = 365 \times AU \times TS \times SAR \times T$$
 [10-6]

where:

SV = Sludge volume (ft³)

AU = Number of 1,000-pound animal units

T = Sludge accumulation time (years)

TS = Total solids production per animal unit

per day (lb/AU/day)

SAR = Sludge accumulation ratio (ft³/lb TS)

Total solids values can be obtained from the tables in chapter 4. Sludge accumulation ratios should be taken from table 10-4. An SAR is not available for beef, but it can be assumed to be similar to that for dairy cattle.

The lagoon volume requirements are for accommodation of the minimum treatment volume, the sludge volume, and the waste volume for the treatment period. This is expressed in equation 10–7.

$$LV = MTV + SV + WV$$
 [10-7]

where:

LV = Lagoon volume requirement, ft³

MTV = Minimum treatment volume, ft³ (see equation 10–4)

SV = Sludge volume accumulation for period between sludge removal events, ft³ (see equation 10–6)

WV = Waste volume for treatment period, ft³ (see equation 10–5)

In addition to the lagoon volume requirement (LV), a provision must be made for depth to accommodate the normal precipitation less evaporation on the lagoon surface; the 25-year, 24-hour storm precipitation; the depth required to operate the emergency outflow; and freeboard. Normal precipitation on the lagoon surface is based on the critical treatment period that produces the maximum depth. This depth can be offset to some degree by evaporation losses on the lagoon surface. This offset varies, according to the climate of the region, from a partial amount of the precipitation to an amount in excess of the precipitation. Precipitation and evaporation can be determined from local climate data.

The minimum acceptable depth for anaerobic lagoons is 6 feet, but in colder climates at least 10 feet is recommended to assure proper operation and odor control.

The design height of an embankment for a lagoon should be increased by the amount needed to ensure that the design elevation is maintained after settlement. This increase should not be less than 5 percent of the design fill height. The minimum top width of the lagoon should be as shown in table 10–5, although a width of 8 feet and less is difficult to construct.

The combined side slopes of the settled embankment should not be less than 5 to 1 (horizontal to vertical). The inside slopes can vary from 1 to 1 for excavated slopes to 3 to 1 or flatter where embankments are used. Construction technique and soil type must also be considered. In some situations a steep slope may be used below the design liquid level, while a flatter slope is used above the liquid level to facilitate maintenance

 Table 10-4
 Sludge accumulation ratios (Barth 1985)

Animal type	SAR	
Poultry		
Layers	0.0295	
Pullets	0.0455	
Swine	0.0485	
Dairy cattle	0.0729	

Table 10–5 Minimum top width for lagoon embankments (USDA 1984, Waste...)

Maximum height of embankment, ft	Top width, ft
10 or less	6
11-14	8
15–19	10
20-24	12
25-34	14
35 or more	15

and bank stabilization. The minimum elevation of the top of the settled embankment should be 1 foot above the maximum design water surface in the lagoon.

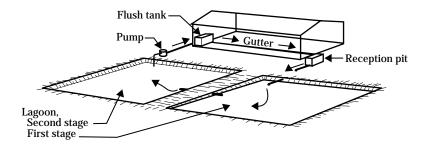
A lagoon should be constructed to avoid seepage and potential ground water pollution. Care in site selection, soils investigation, and design can minimize the potential for these problems. In cases where the lagoon needs to be sealed, the techniques discussed in Appendix 10D, Geotechnical design and construction guidelines for waste impoundment liners, can be used. Also refer to Chapter 7, Geology and Ground Water Considerations, for more information on site evaluation, investigations, and testing. Figure 10–23 shows a two lagoon systems.

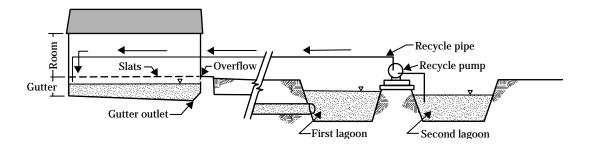
If overtopping can cause embankment failure, an emergency spillway or overflow pipe should be provided. A lagoon can have an overflow to maintain a constant liquid level if the overflow liquid is stored in a

waste storage pond or otherwise properly managed. The inlet to a lagoon should be protected from freezing. This can be accomplished by using an open channel that can be cleaned out or by locating the inlet pipe below the freezing level in the lagoon. Because of possible blockages, access to the inlet pipe is needed. Venting inlet pipes prevents backflow of lagoon gases into the animal production facilities.

Sludge removal is an important consideration in the design. This can be accomplished by agitating the lagoon and pumping out the mixed sludge or by using a drag-line for removing floating or settled sludge. Some pumps can remove sludge, but not deposited rocks, sand, or grit. The sludge removal technique should be considered when determining lagoon surface dimensions. Many agitation pumps have an effective radius of 75 to 100 feet. Draglines may only reach 30 to 50 feet into the lagoon.

Figure 10-23 Anaerobic lagoon recycle systems





Part 651 Agricultural Waste Management Field Handbook

(2) Management

Anaerobic lagoons must be managed properly if they are to function as designed. Specific instructions about lagoon operation and maintenance must be included in the overall waste management plan that is supplied to the decisionmaker. Normally an anaerobic lagoon is managed so that the liquid level is maintained at or below the maximum operating level as shown in figure 10–21. The liquid level is lowered to the minimum treatment level at the end of the treatment period. It is good practice to install markers at the minimum treatment and maximum operating levels.

The minimum liquid level in an anaerobic lagoon before wastes are added should coincide with the MTV. If possible a lagoon should be put into service during the summer to allow adequate development of bacterial populations. A lagoon operates more effectively and has fewer problems if loading is by small, frequent (daily) inflow, rather than large, infrequent slug loads.

The pH should be measured frequently. Many problems associated with lagoons are related to pH in some manner. The optimum pH is about 6.5. When pH falls below this level, methane bacteria are inhibited by the free hydrogen ion concentration. The most frequent cause of low pH in anaerobic digestion is the shock loading of organic material that stimulates the facultative acid-producing bacteria. Add hydrated lime or lye if pH is below 6.5. Add 1 pound per 1,000 square feet daily until pH reaches 7.

Lagoons are designed based on a given loading rate. If an increase in the number of animals is anticipated, sufficient capacity to handle all of the expected waste load should be available. The most common problem in using lagoons is overloading, which can lead to odors, malfunctioning, and complaints. When liquid removal is needed, the liquid level should not be dropped below the MTV plus SV levels. If evaporation exceeds rainfall in a series of dry years, the lagoon should be partly drawn down and refilled to dilute excess concentrations of nutrients, minerals, and toxics. Lagoons are typically designed for 15 to 20 years of sludge accumulation. After this time the sludge must be cleaned out before adding additional waste.

Sometimes operators want to use lagoon effluent as flush water. To polish and store water for this purpose, waste storage ponds can be constructed in series with the anaerobic lagoon. The capacity of the waste storage pond should be sized for the desired storage volume. A minimum capacity of the waste storage pond is the volume for rainfall (RFV), runoff (ROV), and emergency storm storage (ESV). By limiting the depth to less than 6 feet, the pond will function more nearly like an aerobic lagoon. Odors and the level of ammonia, ammonium, and nitrate will be more effectively reduced.

(3) Design example 10–5—Anaerobic lagoon Mr. Oscar Smith of Rocky Mount, North Carolina, has requested assistance in developing an agricultural waste management system for his 6,000 pig finishing facility. The alternative selected includes an anaerobic lagoon. The animals average 150 pounds. The 25-year, 24-hour storm for the area is 6 inches (appendix 10B). Mr. Smith needs 180-day intervals between lagoon pumping. During this time the net precipitation should be 2 inches, based on data from appendices 10B and 10C. He wants to use the lagoon for at least 5 years before removing the sludge. Worksheet 10A–3 is used to determine the necessary volume for this lagoon.

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Completed worksheet for Design example 10--5

Decisionmaker: Oscar Smith	Date: 6/13/90
Site: Rocky Mount, NC	
Animal units	
1. Animal type Growers	3. Number of animals (N)
2. Animal weight, lbs (W)150	4. Animal units, AU = <u>W x N</u> = <u>900</u>
Manure volume	
5. Daily volume of daily manure production per AU, ft ³ /AU/day (DVM)= 1.0	7. Total volume of manure production for animal type for treatment period, ft ³ VMD = AU x DVM x D = 162,000
6. Treatment period, days (D) =	8. Total manure production for treatment period, ft ³ (TVM)
Wastewater volume	
9. Daily wastewater volume per AU, ft ³ /AU/day (DWW) =	11. Total wastewater volume for treatment period, ft ³ (TWW)
10. Total wastewater volume for animal	
description for treatment period, ft ³ WWD = DWW x AU x D =	
Clean water volume	
12. Clean water added during treatment period, ft ³ (CW)	
Waste volume	
13. Waste volume for treatment period, ft ³ $WV = TVM + TWW + CW = \frac{1}{2}$	62,000 ₊
Manure total solids	34
15. Daily manure total solids production for animal type. Ibs/day	34 16. Total manure total solids production,
MTSD = MTS x AU = 570	06 lbs/day (TMTS) = -
Manure volatile solids 17. Daily manure volatile solids production per AU, lbs/AU/day (MVS) =	5.4
17. Daily manure volatile solids production per AU, lbs/AU/day (MVS) =	1860
18. Daily manure volatile solids production for animal type per day, lbs/day	
19. Total manure volatile solids production, lbs/day (TMVS)	
Wastewater volatile solids	
Daily wastewater volatile solids production, lbs/1000 gal (DWVS) Total wastewater volatile solids production for animal type, lbs/day.	
21. Total wastewater volatile solids production for animal type, lbs/day	
WVSD = DWVS x DWW x 7.48 D x 1.000	=
22. Total wastewater volatile solids production, lbs/day (TWVS)	===
Total volatile solids (manure and wastewater)	40/0
23. Total daily volatile solids production, lbs/day TVS = TMVS + TWVS =	4860 + =
Minimum treatment volume	25. Minimum treatment volume, ft ³
24. Selected lagoon VS loading rate, lbs VS/1,000 ft ³ (VSLR) =	$\frac{6}{\text{MTV}} = \frac{\text{TVS x 1000}}{\text{VSLR}} = \frac{(4860) \times 1000}{(6)} = \frac{8}{1000}$
Sludge volume requirement 26. Sludge accumulation ratio. ft ³ /lb TS (SAR) = 0.0485	28. Sludge volume requirement, ft ³
	$SV = 365 \times TMTS \times T \times SAR$
27 Sludge accumulation period, years (T) =5	= 365 x (5706)(5)(0.0485) = 5
Minimum lagoon volume requirement	
29. Minimum lagoon volume requirements, ft ³	00 + 505,052 + 162,000 = 1,2

Completed worksheet for Design example 10-5—Continued

_agoon sizing		(4 72 13)	/ 7 8	2) /		
30. Sizing by trial an	d error V	$=\frac{(4 \times 2 \times a)}{3}$)+ (Z x BL x t	d") ₊ (ZxBWxd	²) ₊ (BWx BL x d)	
Side slope ratio,	(Z) = <u>2</u>			V must be equal to	or greater than MLVF	R = 1,477,052 ft ³
Trial no. 1 2 3	Bottom width ft (BW)	150 150 150	Bottom ler ft (BL)	-	Depth* 8 ft (d) 8	Volume ft³ (V) 1,349,931 1,615,531 1,482,731 ≈ MLVR
	- Ci 21			1100		1,402,731 ~ WEVK
Depth must be adjusted i	п эгер зт.					
Depth adjustm	ent					
31. Depth adjustme	ent					
Depth, ft (d)				8		
Add depth of pre (for the treatm	ecipitation less evapora nent period)	tion on lagoon surf	ace+	0.6		
Add depth of 25-	-year, 24-hour storm		+	0.5		
Add for freeboard	d (1.0 foot minimum)		+	1.0		
Final depth				10.1		
						2,014,299
32. Compute total vo	lume using final depth,	ft ³ (use equation	in step 30)			

(b) Aerobic lagoons

Aerobic lagoons can be used if minimizing odors is critical (fig. 10–24). These lagoons operate within a depth range of 2 to 5 feet to allow for the oxygen entrainment that is necessary for the aerobic bacteria.

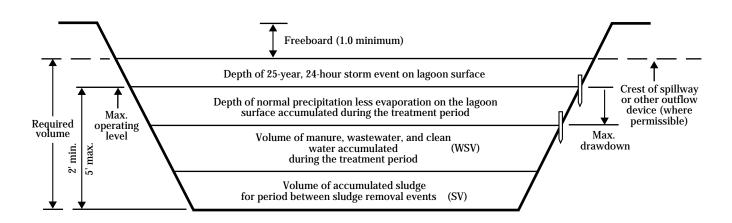
The design of aerobic lagoons is based on the amount of BOD_5 added per day. If local data are not available, use the BOD_5 values from the tables in chapter 4. Figure 10--25 shows the acceptable aerobic loading rates for the United States in lb- BOD_5 /acre/day. The lagoon surface area at the average operating depth is sized so that the acceptable loading rate is not exceeded.

Even though an aerobic lagoon is designed on the basis of surface area, it must have enough capacity to accommodate the waste volume (WV) and sludge volume (SV). In addition, depth must be provided to accommodate the normal precipitation less evapora-

tion on the lagoon surface, the 25-year, 24-hour storm precipitation on the lagoon surface, and freeboard. Should State regulations not permit an emergency outflow or for some other reason one is not used, the minimum freeboard is 1 foot above the top of the required volume. Figure 10–24 demonstrates these volume depth requirements.

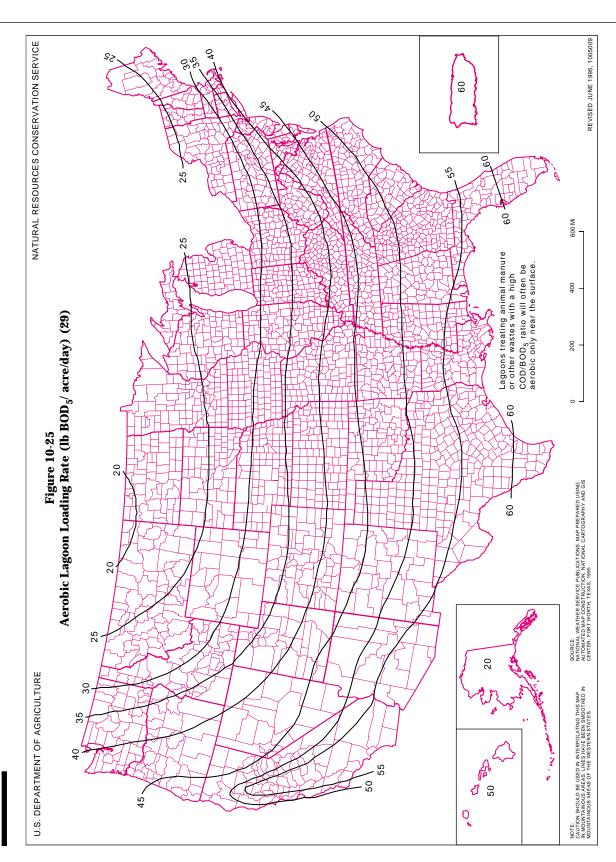
Aerobic lagoons need to be managed similarly to anaerobic lagoons in that they should never be overloaded with oxygen demanding material. The lagoon should be filled to the minimum operating level, generally 2 feet, before being loaded with waste. The maximum liquid level should not exceed 5 feet. The water level must be maintained within the designed operating range. Sludge should be removed when it exceeds the designed sludge storage capacity. Aerobic lagoons should also be enclosed in fences and marked with warning signs.

Figure 10–24 Aerobic lagoon cross section



Note: An aerobic waste treatment lagoon has a required minimum surface area based on ${\rm BOD}_5$

Figure 10-25 Aerobic lagoon loading rate



Chapter	10

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(1) Design example 10-6—Aerobic lagoon Mr. John Sims of Greenville, Mississippi, has requested assistance on the development of an agricultural waste management system. He has requested that an alternative be developed that includes an aerobic lagoon to

treat the waste from his 50,000 caged layers, which have an average weight of 4 pounds. Completed worksheet 10A–4 shows the calculations to size the lagoon for this design example.

Decisionmaker: John Sims	-Aerobic lagoon design Date: 11/16/90	
	11/10/90	
Greenville, MS		
Animal units Caged		
1. Animal type	3. Number of animals (N)	
1		
2. Animal weight, lbs (W)	4. Animal units, AU = W x N =	
Manure volume		
5. Daily volume of daily manure production per AU, $ft^3/AU/day$ (DVM) = 0.93	VMD = ALLY DVM y D = 33,40U	
6. Treatment period, days (D) =	8. Total manure production for treatment period, ft ³ (TVM)3	3,480
Wastewater volume		
9. Daily wastewater volume per	11. Total wastewater volume for treatment period, ft ³ (TWW)	0
AU, ft ³ /AU/day (DWW) = 10. Total wastewater volume for animal description for treatment period, ft ³ WWD = DWW x AU x D =	пеаннен ренов, и (туучу)	
Clean water volume		
		0
Waste volume		
13. Waste volume for treatment period, ft^3 WV = TVM + TV	NW + CW = 33,480 + 0 + 0 = 33,480	0
Manure total solids	15 1	
14. Daily manure total solids production, lbs/AU/day (MTS) =		3020
	lbs/day (TMTS) =	3020
15. Daily manure total solids production for animal type, lb/da MTSD = MTS x AU =	ay 3020	3020
Manure 5-day biochemical oxygen demand	^{ay} 3020	3020
Manure 5-day biochemical oxygen demand 17. Daily manure BOD ₅ production per AU, lbs/AU/day (MBOD)	3020	3020
Manure 5-day biochemical oxygen demand 17. Daily manure BOD ₅ production per AU, lbs/AU/day (MBOD) 18. Daily manure BOD ₅ production for animal type per day, lbs	3020	
Manure 5-day biochemical oxygen demand 17. Daily manure BOD ₅ production per AU, lbs/AU/day (MBOD) 18. Daily manure BOD ₅ production for animal type per day, lbs 19. Total manure production, lbs/day (TMBOD)	3020	740
Manure 5-day biochemical oxygen demand 17. Daily manure BOD ₅ production per AU, lbs/AU/day (MBOD) 18. Daily manure BOD ₅ production for animal type per day, lbs 19. Total manure production, lbs/day (TMBOD) Wastewater 5-day biochemical oxygen dem	ay 3020	
Manure 5-day biochemical oxygen demand 17. Daily manure BOD ₅ production per AU, lbs/AU/day (MBOD) 18. Daily manure BOD ₅ production for animal type per day, lbs 19. Total manure production, lbs/day (TMBOD) Wastewater 5-day biochemical oxygen dem 20. Daily wastewater BOD ₅ production, lbs/1000 gal (DWBOD) =	ay 3020	
Manure 5-day biochemical oxygen demand 17. Daily manure BOD ₅ production per AU, lbs/AU/day (MBOD) 18. Daily manure BOD ₅ production for animal type per day, lbs 19. Total manure production, lbs/day (TMBOD) Wastewater 5-day biochemical oxygen dem 20. Daily wastewater BOD ₅ production, lbs/1000 gal (DWBOD) = 21. Total wastewater BOD ₅ production for animal type, lbs/day	ay 3020	
Manure 5-day biochemical oxygen demand 17. Daily manure BOD ₅ production per AU, lbs/AU/day (MBOD) 18. Daily manure BOD ₅ production for animal type per day, lbs 19. Total manure production, lbs/day (TMBOD) Wastewater 5-day biochemical oxygen dem 20. Daily wastewater BOD ₅ production, lbs/1000 gal (DWBOD) =	ay 3020	740
Manure 5-day biochemical oxygen demand 17. Daily manure BOD ₅ production per AU, lbs/AU/day (MBOD) 18. Daily manure BOD ₅ production for animal type per day, lbs 19. Total manure production, lbs/day (TMBOD) Wastewater 5-day biochemical oxygen dem 20. Daily wastewater BOD ₅ production, lbs/1000 gal (DWBOD) = 21. Total wastewater BOD ₅ production for animal type, lbs/day WBOD = (DWBOD x TWW x 7.48)	ay 3020	
Manure 5-day biochemical oxygen demand 17. Daily manure BOD ₅ production per AU, lbs/AU/day (MBOD) 18. Daily manure BOD ₅ production for animal type per day, lbs 19. Total manure production, lbs/day (TMBOD) Wastewater 5-day biochemical oxygen dem 20. Daily wastewater BOD ₅ production, lbs/1000 gal (DWBOD) = 21. Total wastewater BOD ₅ production for animal type, lbs/day WBOD = (DWBOD x TWW x 7.48) D x 1,000	ay 3020	740
Manure 5-day biochemical oxygen demand 17. Daily manure BOD ₅ production per AU, lbs/AU/day (MBOD) 18. Daily manure BOD ₅ production for animal type per day, lbs 19. Total manure production, lbs/day (TMBOD) Wastewater 5-day biochemical oxygen dem 20. Daily wastewater BOD ₅ production, lbs/1000 gal (DWBOD) = 21. Total wastewater BOD ₅ production for animal type, lbs/day WBOD =(DWBOD x TWW x 7.48) D x 1,000 22. Total wastewater BOD ₅ production, lbs/day (TWBOD)	ay 3020	740 0
Manure 5-day biochemical oxygen demand 17. Daily manure BOD ₅ production per AU, lbs/AU/day (MBOD) 18. Daily manure BOD ₅ production for animal type per day, lbs 19. Total manure production, lbs/day (TMBOD) Wastewater 5-day biochemical oxygen dem 20. Daily wastewater BOD ₅ production, lbs/1000 gal (DWBOD) = 21. Total wastewater BOD ₅ production for animal type, lbs/day WBOD = (DWBOD x TWW x 7.48) D x 1,000 22. Total wastewater BOD ₅ production, lbs/day (TWBOD) TOTAL BOD ₅ (manure and wastewater) 23. Total daily production, lbs/day TBOD = TMBOD + TWBOD	ay 3020	740 0
Manure 5-day biochemical oxygen demand 17. Daily manure BOD ₅ production per AU, lbs/AU/day (MBOD) 18. Daily manure BOD ₅ production for animal type per day, lbs 19. Total manure production, lbs/day (TMBOD) Wastewater 5-day biochemical oxygen dem 20. Daily wastewater BOD ₅ production, lbs/1000 gal (DWBOD) 21. Total wastewater BOD ₅ production for animal type, lbs/day WBOD =(DWBOD x TWW x 7.48)	ay 3020	740 0 740
Manure 5-day biochemical oxygen demand 17. Daily manure BOD ₅ production per AU, lbs/AU/day (MBOD) 18. Daily manure BOD ₅ production for animal type per day, lbs 19. Total manure production, lbs/day (TMBOD) Wastewater 5-day biochemical oxygen dem 20. Daily wastewater BOD ₅ production, lbs/1000 gal (DWBOD) 21. Total wastewater BOD ₅ production for animal type, lbs/day WBOD =(DWBOD x TWW x 7.48)	ay 3020	740 0 740 14.8
Manure 5-day biochemical oxygen demand 17. Daily manure BOD ₅ production per AU, lbs/AU/day (MBOD) 18. Daily manure BOD ₅ production for animal type per day, lbs 19. Total manure production, lbs/day (TMBOD) Wastewater 5-day biochemical oxygen dem 20. Daily wastewater BOD ₅ production, lbs/1000 gal (DWBOD) 21. Total wastewater BOD ₅ production for animal type, lbs/day WBOD =(DWBOD x TWW x 7.48)	ay 3020	740 0 740
Manure 5-day biochemical oxygen demand 17. Daily manure BOD ₅ production per AU, lbs/AU/day (MBOD) 18. Daily manure BOD ₅ production for animal type per day, lbs 19. Total manure production, lbs/day (TMBOD) Wastewater 5-day biochemical oxygen dem 20. Daily wastewater BOD ₅ production, lbs/1000 gal (DWBOD) = 21. Total wastewater BOD ₅ production for animal type, lbs/day WBOD =(DWBOD x TWW x 7.48) D x 1,000 22. Total wastewater BOD ₅ production, lbs/day (TWBOD) TOTAL BOD ₅ (manure and wastewater) 23. Total daily production, lbs/day TBOD = TMBOD + TWBOD Minimum treatment surface area 24. Selected lagoon BOD ₅ loading rate, lbs BOD ₅ /acre (BODLR) Sludge volume requirement 26. Sludge accumulation ratio, ft³/lb TS (SAR)	ay 3020	740 0 740 14.8

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agoon sizing 30. Sizing by trial and	error:					
) =2					
	or greater than MLVR = _	196,069 _{ft³}				
	o or greater than MTA = _					
Rectangular lagoor						
d must be less that						
a must be less that	101001					
SA= <u>(BL + 2Zd</u> 43)(BW + 2Zd) ,560					
Trial no.	Bottom width ft (BW)	Bottom length ft (BL)	Depth* ft (d)	Volume ft ³ (V)	Surface area acres (SA)	
1	600	1100	1	663,405	<u>15.3 OK</u>	
* Depth must be a	djusted in Step 31					
Depth adjustme	nt					
31. Depth adjustment						
Depth , ft (d)		1.	0			
Add depth of precip	itation less evaporation or nt period)	n lagoon surface +O	1.5			
Add depth of 25-ye	ar, 24-hour storm	+ <u>C</u>).6			
Add for freeboard	(1.0 foot minimum)	+1.	.0			
Final depth		3	3.1			

(c) Mechanically aerated lagoons

Much of this material was taken directly from technical notes on the design of mechanically aerated lagoons for odor control (Moffitt 1980).

Aerated lagoons operate aerobically and are dependent on mechanical aeration to supply the oxygen needed to treat waste and minimize odors. This type of design is used to convert an anaerobic lagoon to an aerobic condition, or as an alternative, to a naturally aerated lagoon that would otherwise need to be much larger. Mechanically aerated lagoons combine the small surface area feature of anaerobic lagoons with relative odor free operation of an aerobic lagoon. The main disadvantages of this type of lagoon are the energy requirements to operate the mechanical aerators and the high level of management required.

The typical design includes 1 pound of oxygen transferred to the lagoon liquid for each pound of BOD_5 added. The TS content in aerated lagoons should be maintained between 1 and 3 percent with dilution water. The depth of aerated lagoons depends on the type of aerator used. Agitation of settled sludge needs to be avoided. As with naturally aerobic lagoons, consideration is required for storage of manure and rainfall.

Two kinds of mechanical aerator are used: the surface pump and the diffused air system. The **surface pump** floats on the surface of the lagoon, lifting water into the air, thus assuring an air-water mixture. The **diffused air system** pumps air through water, but is generally less economical to operate than the surface pump.

(1) Lagoon loading

Lagoon loading should be based on 5-day biochemical oxygen (BOD $_5$) or carbonaceous oxygen demand (COD). NRCS designs on the basis of BOD $_5$. The tables in chapter 4 show recommended BOD $_5$ production rates, but local data should be used where available.

(2) Aerator design

Aerators are designed primarily on their ability to transfer oxygen (O_2) to the lagoon liquid. Of secondary importance is the ability of the aerator to mix or disperse the O_2 throughout the lagoon. Where the aerator is intended for minimizing odors, complete mixing is not a consideration except as it relates to the surface area.

For the purpose of minimizing odors, aerators should transfer from 1 to 2 pounds of oxygen per pound of BOD_5 . Even a limited amount of oxygen transfer (as little as 1/3 lb O_2 per lb BOD_5) reduces the release of volatile acids and accompanying gases. For design purposes, use 1 pound of oxygen per pound of BOD_5 unless local research indicates a higher value is needed.

Aerators are tested and rated according to their clean water transfer rate (CWTR) or laboratory transfer rate (LTR), whichever term is preferred. The resulting value is given for transfer at standard atmospheric pressure (14.7 psi), dissolved oxygen equal to 0 percent, and water at 20 $^{\circ}$ C. The actual transfer rate expected in field operation can be determined by using equation 10–8.

$$FTR = CWTR \times \frac{\left(B \times C_{dc}\right) - DO}{C_{sc}} \times O^{t-20} \times a$$
 [10-8]

where:

FTR = lb O₂ per horsepower-hour transferred under field conditions

CWTR = clean water transfer rate in lb per horsepower-hour transferred under standard laboratory conditions

B = salinity-surface tension factor. It is the ration of the saturated concentration in the wastewater to that of clean water. Values range from 0.95 to 1.0.

 C_{dc} = O_2 saturation concentration at design conditions of altitude and temperature (mg/L) from figures 10–26 and 10–27.

DO = Average operating O₂ concentration (mg/L). The recommended value of DO can vary from 1 to 3 depending on the reference material. A value of 1.5 should be considered a minimum. For areas where minimizing odors is particularly critical, a DO of 2 or more should be used.

t = Design temperature (°C)

O = Temperature correction factor; values range from 1.024 to 1.035.

a = The ratio of the rate of O₂ transfer in the wastewater to that of clean water. Generally taken as 0.75 for animal waste.

 C_{sc} =Saturation concentration of O₂ in clean water, 20 °C and sea level (9.17 mg/L).

Unless local information supports using other values, the following values for calculating field transfer rates should be used: B=1.0, DO=1.5, O=1.024, a=0.75, and $C_{\rm sc}=9.17$.

Figure 10–28 provides a quick solution to the term O^{t-20}, where O is equal to 1.024. Designs for both summer and winter temperatures are often necessary to determine the controlling (least) transfer rate.

Having calculated FTR, the next step is to determine horsepower requirements of aeration based on loading rates and FTR as calculated above. Horsepower requirements can be estimated using equation 10–9.

$$HP = \frac{BOD_5}{FTR \times HO}$$
 [10-9]

where:

HP = Horsepower

 $BOD_5 = 5$ -day biochemical oxygen demand

loading of waste, lb/day

HO = Hours of operation per day

Most lagoon systems should be designed on the basis of continual aerator operations.

The actual selection of aerator(s) is a subjective process and often depends on the availability of models in the particular area. In general, multiple small units are preferred to one large unit. The multiple units provide better coverage of the surface area as well as permit flexibility for the real possibility of equipment failure and reduced aeration.

Figure 10–27 Relation of dissolved oxygen saturation to elevation above mean sea level

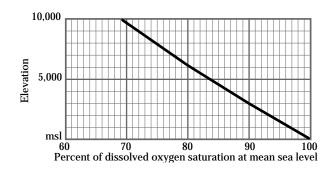


Figure 10-26 Relation of dissolved oxygen saturation to water temperature (clean water at 20 °C and sea level)

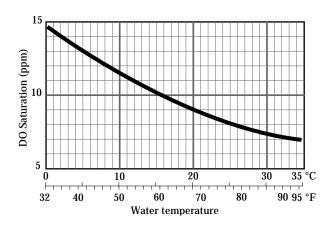
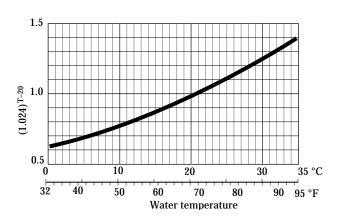


Figure 10–28 Numeral values for O^{t-20} at different temperatures where O=1.024



(d) Oxidation ditches

In some situations sufficient space is not available for a lagoon for treating animal waste, and odor control is critical. One option for treating animal waste under these circumstances is an oxidation ditch (fig. 10–29). The shallow, continuous ditch generally is in an oval layout. It has a special aerator spanning the channel. The action of the aerator moves the liquid waste around the channel and keeps the solids in suspension. Because of the need for continuous aeration, this process can be expensive to operate. Oxidation ditches should only be designed by a professional engineer familiar with the process.

The range of loading for an oxidation ditch is 1 pound of BOD_5 per 30 to 100 cubic feet of volume. This provides for a retention time of 30 to 70 days. Solids accumulate over time and must be removed by settling. The TS concentration is maintained in the 2 to 6 percent range, and dilution water must be added periodically.

If oxidation ditches are not overloaded, they work well for minimizing odors. The degree of management required, however, may be more than desired by some operators. Daily attention is often necessary, and equipment failure can lead to toxic gas generation soon after the aerators are stopped. If the ditches are properly managed, they can be effective in reducing nitrogen to N_2 through cyclic aerobic/anaerobic periods, which allows nitrification and then denitrification.

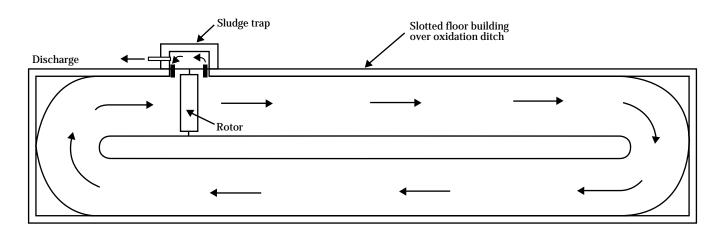
(e) Drying/dewatering

If the water is removed from freshly excreted manure, the volume to handle can be reduced. The process of removing water is referred to as dewatering. In the arid regions of the United States, most manure is dewatered (dried) by evaporation from sun and wind. Some nutrients may be lost in the drying process.

Dried or dewatered manure solids are often sold as a soil conditioner or garden fertilizer. These solids may also be used as fertilizer on agricultural land. They are high in organic matter and can be expected to produce odors if moisture is added and the material is not redried or composted. Because the water is removed, the concentrations of some nutrients and salts will change. Dried manure should be analyzed to determine the nutrient concentrations before land application.

In humid climates dewatering is accomplished by adding energy to drive off the desired amount of moisture. Processes have been developed for drying manure in greenhouse-type facilities; however, the drying rate is dependent on the temperature and relative humidity. The cost of energy often makes the drying process unattractive.

Figure 10-29 Schematic of an oxidation ditch



(f) Composting

Composting is the aerobic biological decomposition of organic matter. It is a natural process that is enhanced and accelerated by the mixing of organic waste with other ingredients in a prescribed manner for optimum microbial growth.

Composting converts an organic waste material into a stable organic product by converting nitrogen from the unstable ammonia form to a more stable organic form. The end result is a product that is safer to use than raw organic material and one that improves soil fertility, tilth, and water holding capacity. In addition, composting reduces the bulk of organic material to be spread; improves its handling properties; reduces odor, fly, and other vector problems; and can destroy weed seeds and pathogens.

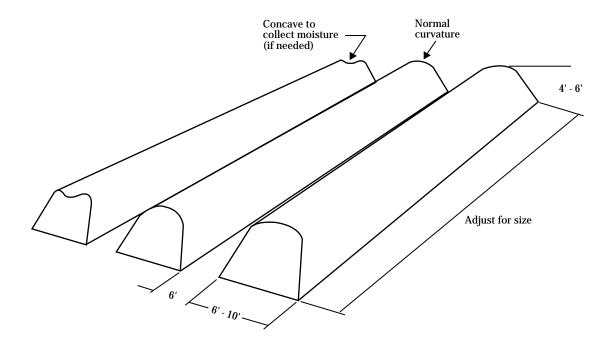
(1) Composting methods

Three basic methods of composting—windrow, static pile, and in-vessel—are described below.

(i) Windrow method—The windrow method involves the arrangement of compost mix in long, narrow piles or windrows (fig. 10–30). To maintain an aerobic condition, the compost mixture must be periodically turned. This exposes the decomposing material to the air and keeps temperatures from getting too high (>170 $^{\circ}$ F). The minimum turning frequency varies from 2 to 10 days, depending on the type of mix, volume, and the ambient air temperature. As the compost ages, the frequency of turning can be reduced.

The width and depth of the windrows are limited only by the type of turning equipment used. Turning equipment can range from a front-end loader to a automatic mechanical turner. Windrows generally are 4 to 6 feet deep and 6 to 10 feet wide.

Figure 10-30 Windrow schematic



Some advantages and disadvantages of the windrow method include:

Advantages:

- Rapid drying with elevated temperatures
- Drier product, resulting in easier product handling
- · Ability to handle high volumes of material
- Good product stabilization
- · Low capital investment

Disadvantages:

- Not space efficient
- High operational costs
- Piles should be turned to maintain aerobic conditions
- Turning equipment may be required
- Vulnerable to climate changes
- Odors released on turning of compost
- Large volume of bulking agent might be required

(ii) Static pile method—The static pile method consists of mixing the compost material and then stacking the mix on perforated plastic pipe or tubing through which air is drawn or forced. Forcing air through the compost pile may not be necessary with small compost piles that are highly porous or with a mix that is stacked in layers with highly porous material. The exterior of the pile generally is insulated with finished compost or other material. In nonlayered

operations, the materials to be composted must be thoroughly blended before pile placement.

The dimensions of the static pile are limited by the amount of aeration that can be supplied by the blowers and the stacking characteristics of the waste. The compost mixture height generally ranges from 8 to 15 feet, and the width is usually twice the depth. Individual piles generally are spaced about a half the distance of the height.

With forced air systems, air movement through the pile occurs by suction (vacuum) or by positive pressure (forced) through perforated pipes or tubing. A filter pile or material is normally used to absorb odor if air is sucked through the pile (fig. 10–31).

Some advantages and disadvantages of the static pile method include:

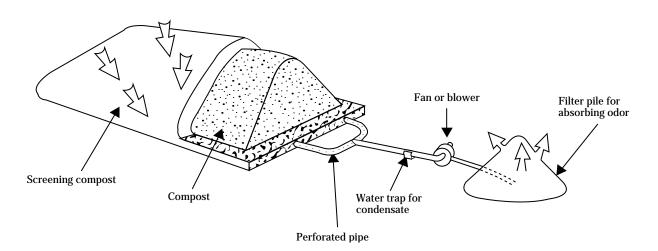
Advantages:

- Low capital cost
- High degree of pathogen destruction
- Good odor control
- Good product stabilization

Disadvantages:

- Not space efficient
- Vulnerable to climate impacts
- Difficult to work around perforated pipe unless recessed
- Operating cost and maintenance on blowers

Figure 10–31 Static pile composting schematic



(iii) In-vessel method—The in-vessel method in-volves the mixing of manure or other organic waste with a bulking agent in a reactor, building, container, or vessel (fig. 10–32) and may involve the addition of a controlled amount of air over a specific detention time. This method has the potential to provide a high level of process control because moisture, aeration, and temperature can be maintained with some of the more sophisticated units. Dead animal composting in a composting bin as discussed in section 651.1007(b), Dead animal disposal, is an example of unsophisticated in-vessel composting.

Some of the advantages and disadvantages of the invessel method include:

Advantages:

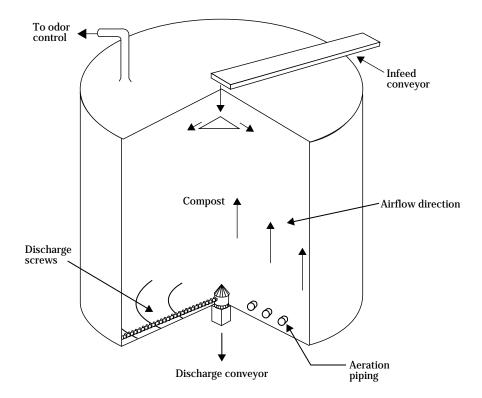
- Space efficient
- Good process control because of self-containment
- · Protection from adverse climate conditions

- Good odor control because of self-containment and process control
- Potential for heat recovery dependent on system design
- Can be designed as a continuous process rather than a batch process

Disadvantages:

- · High capital cost for sophisticated units
- Lack of operating data, particularly for large systems
- · Careful management required
- Dependent on specialized mechanical and electrical equipment
- Potential for incomplete stabilization
- Mechanical mixing needs to be provided
- Less flexibility in operation mode than with other methods

Figure 10–32 In-vessel composting schematic



(2) Method selection

The composting method must fit the individual farm operation. Highly sophisticated and expensive composting operations are not likely to be a viable option for small farming operations. Some factors to consider when selecting the particular method of composting include:

- (i) Operator management capability—The management capability of the operator is an important consideration when selecting the right composting method. Even simple composting methods require that the operator spend additional time in monitoring and material handling. The operator should fully understand the level of management that is required. The windrow method generally is the simplest method to manage, but requires additional labor for periodically turning the compost mix. The static pile is generally next in complexity because of having to maintain blowers and work around perforated pipe. In-vessel composting can be the simplest or the most difficult to manage, depending on the sophistication of the system.
- (ii) Equipment and labor availability—Consider what equipment is available for loading, unloading, turning, mixing, and hauling. The windrow method requires extra equipment and labor to periodically turn the rows. All methods require some type of loading and unloading equipment.
- (iii) Site features—If a limited amount of space is available, then the static pile or in-vessel method may be the only viable composting alternatives. Proximity to neighbors and the appearance of the compost operation may make the windrow and static pile methods unattractive alternatives. If the only composting site has limited accessibility, then the static pile or in-vessel method should be considered because of less mixing requirements. Siting considerations are discussed more fully in the Siting and area considerations section that follows.
- (iv) Compost utilization—If the compost is to be marketed commercially, then a composting method that produces a predictable, uniform product should be considered. Because of varying climatic conditions, the windrow method may not produce a predictable end product. Sophisticated in-vessel methods provide the most process control; therefore, they produce the most uniform and predictable product.

- (v) Climate—In extremely wet climates the static pile and aerated composting methods may become too wet to compost properly unless measures are taken to protect the compost from the weather. In very cold climates, the composting process may slow in the winter. Sheltering the compost pile from the wind helps to prevent a slowdown in the composting process. The windrow and static pile methods are the most vulnerable to freezing temperatures because they are exposed to the elements. All methods may perform unsatisfactorily if the organic waste and amendments are initially mixed in a frozen state.
- (vi) Cost—Composting capital and operating costs vary considerably depending on the degree of sophistication. The windrow method generally has the least capital cost, but also has the most operational costs. The in-vessel method usually has the highest initial capital cost, but the lowest operational cost.

(3) Siting and area considerations

The location of the composting facility is a very important factor in a successful compost operation. To minimize material handling, the composting facility should be located as close as possible to the source of organic waste. If land application is the preferred method of utilization, the facility should also be located with convenient access to the land application sites. Several other important considerations when locating a compost facility are discussed below.

- (i) Wind direction—Improperly managed compost facilities may generate offensive odors until corrective actions are taken. Wind direction and proximity to neighbors should be considered when locating a composting facility.
- (ii) **Topography**—Avoid locating composting facilities on steep slopes where runoff may be a problem and in areas where the composting facility will be subject to inundation.
- (iii) Ground water protection—The composting facility should be located downgradient and at a safe distance from any wellhead. A roofed compost facility, that is properly managed, should not generate leachate that could contaminate ground water. If a compost facility is not protected from the weather, it should be sited to minimize the risk to ground water.

(iv) Area requirements—The area requirements for each composting method vary. The windrow method requires the most land area. The static pile method requires less land area than the windrow method, but more than the in-vessel method. The pile dimensions also affect the amount of land area necessary for composting. A large pile that has a low surface area to total volume ratio requires less composting area for a given volume of manure, but it is also harder to manage. The size and type equipment used to mix, load, and turn the compost should also be considered when sizing a compost area. Enough room must be provided in and around the composting facility to operate equipment. In addition, a buffer area around the compost site should be considered if a visual barrier is needed or desired. In general, given the pile dimensions, a compost bulk density of 35 to 45 pounds per cubit feet can be used to estimate the surface area necessary for stacking the initial compost mix. To this area, add the amount of area necessary for equipment operation, pile turning, and buffer.

(v) Existing areas—To reduce the initial capital cost, existing roofed, concrete, paved, or gravel areas should be used if possible as a composting site.

(4) Compost utilization

Finished compost is used in a variety of ways, but is primarily used as a fertilizer supplement and soil conditioner. Compost improves soil structure and soil fertility, but it generally contains too low a quantity of nitrogen to be considered the only source of crop nitrogen. Nutrients in finished compost will be slowly released over a period of years, thus minimizing the risk of nitrate leaching and high nutrient concentrations in surface runoff. For more information on land application of organic material, see chapter 11.

A good quality compost can result in a product that can be marketed to home gardeners, landscapers, vegetable farmers, garden centers, nursery/greenhouses, turf growers, golf courses, and ornamental crop producers. Generally, the marketing of compost from agricultural operations has not provided enough income to completely cover the cost of composting. If agricultural operations do not have sufficient land to spread the waste, marketing may still be an attractive alternative compared to hauling the waste to another location for land spreading. Often, compost operators generate additional income by charging municipalities and other local governments for composting urban

yard waste with the waste products of the agricultural operations.

Finished compost has also been successfully used as a bedding material for livestock. Because composting generates high temperatures that dry out and sterilize the compost, the finished product is generally acceptable as a clean, dry, bedding material. Refeeding of the poultry compost as a food supplement is currently being tested and may prove to be an acceptable use of poultry compost.

(5) Compost mix design

Composting of organic waste requires the mixing of an organic waste with amendment(s) or bulking agent(s) in the proper proportions to promote aerobic microbial activity and growth and to achieve optimum temperatures. The following must be provided in the initial compost mix and maintained during the composting process:

- A source of energy (carbon) and nutrients (primarily nitrogen).
- Sufficient moisture.
- Sufficient oxygen for an aerobic environment.
- A pH in the range of 6 to 8.

The proper proportion of waste, amendments, and bulking agents is commonly called the "recipe."

A composting amendment is any item added to the compost mixture that alters the moisture content, C:N ratio, or pH. Many materials are suitable for use as a composting amendment. Crop residue, leaves, grass, straw, hay, and peanut hulls are just some of the examples that may be available on the farm. Others, such as sawdust, wood chips, or shredded paper and cardboard, may be available inexpensively from outside sources. Table 10–6 shows typical C:N ratios of common composting amendments. The C:N ratio is highly variable, and local information or laboratory values should be used whenever possible.

A bulking agent is used primarily to improve the ability of the compost to be self supporting (structure) and to increase porosity to allow internal air movement. Wood chips and shredded tires are examples of a bulking agent. Some bulking agents, such as large wood chips, may also alter the moisture content and C:N ratio, in which case they would be both a bulking agent and a compost amendment.

(i) Compost design parameters—To determine the recipe, the characteristics of the waste and the amendments and bulking agents must be known. The characteristics that are the most important in determining the recipe are moisture content (wet basis), carbon content, nitrogen content, and the C:N ratio. If any two of the last three components are known, the remaining one can be calculated.

Carbon to nitrogen (C:N) ratio—The balance between carbon and nitrogen in the compost mixture is a critical factor for optimum microbial activity. After the organic waste and the compost ingredients are mixed together, micro-organisms multiply rapidly and consume carbon as a food source and nutrients to metabolize and build proteins. The C:N ratio of the compost mix should be maintained for most compost operations between 25 and 40 to 1. If the C:N ratio is low, a loss of nitrogen generally occurs through rapid

 Table 10-6
 Typical carbon to nitrogen ratios of common composting amendments*

Material	C:N ratios	Material	C:N ratios
Alfalfa (broom stage)	20	Pig manure	5-8
Alfalfa hay	12-18	Pine needles	225-1000
Asparagus	70	Potato tops	25
Austrian pea straw	59	Poultry manure (fresh)	6-10
Austrian peas (green manure)) 18	Poultry manure (henhouse litter)	12-18
Bark	100-130	Reeds	20-50
Bell pepper	30	Residue of mushroom culture	40
Breading crumbs	28	Rice straw	48-115
Cantaloupe	20	Rotted manure	20
Cardboard	200-500	Rye straw	60 - 350
Cattle manure (with straw)	25-30	Saw dust	300-723
Cattle manure (liquid)	8-13	Sawdust (beech)	100
Clover	12-23	Sawdust (fir)	230
Clover (sweet and young)	12	Sawdust (old)	500
Corn & sorghum stover	60-100	Seaweed	19
Cucumber	20	Shredded tires	95
Dairy manure	10-18	Soil organic matter	10-24
Garden wastes	20-60	Soybean residues	20-40
Grain rice	36	Straw	40-80
Grass clippings	12-25	Sugar cane (trash)	50
Green leaves	30-60	Timothy	80
Green rye	36	Tomato leaves	13
Horse manure (peat litter)	30-60	Tomatoes	25 - 30
Leaves (freshly fallen)	40-80	Watermelon	20
Newspaper	400-500	Water hyacinth	20-30
Oat straw	48-83	Weeds	19
Paper	173	Wheat straw	60-373
Pea vines (native)	29	Wood (pine)	723
Peat (brown or light)	30-50	Wood chips	100-441

^{*} For further information on C:N ratios, see chapter 4 of this handbook.

decomposition and volatilization of ammonia. If it is high, the composting time increases because the nitrogen becomes the limiting nutrient for growth.

Moisture—Micro-organisms need moisture to convert the carbon source to energy. Bacteria generally can tolerate a moisture content as low as 12 to 15 percent; however, with less than 40 percent moisture, the rate of decomposition is slow. At greater than 60 percent moisture, the process turns from one that is aerobic to one that is anaerobic. Anaerobic composting is less desirable because it decomposes more slowly and produces putrid odors. The finished product should result in a material that has a low moisture content.

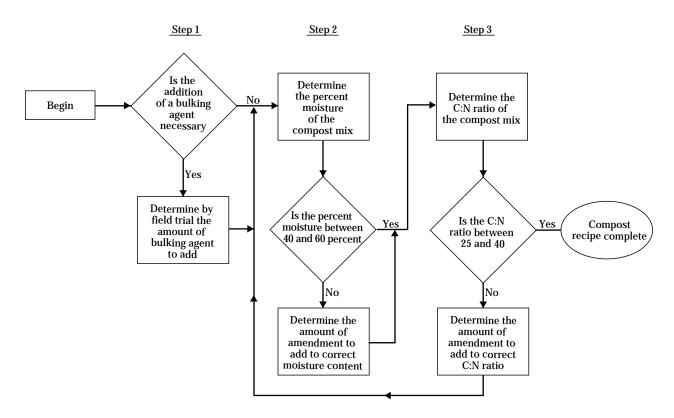
pH—Generally, pH is self-regulating and is not a concern when composting agricultural waste. Bacterial growth generally occurs within the range of pH 6.0 to 7.5, and fungi growth usually occurs within the range of 5.5 to 8.5. The pH varies throughout the compost mixture and during the various phases of the

composting process. The pH in the compost mixture is difficult to regulate once decomposition is started. Optimum pH control can be accomplished by adding alkaline or acidic materials to the initial mixture.

(ii) Compost mix design process—The determination of the compost mix design (recipe) is normally an iterative process of adjusting the C:N ratio and moisture content by the addition of amendments. If the C:N ratio is out of the acceptable range, then amendments are added to adjust it. If this results in a high or low moisture content, amendments are added to adjust the moisture content. The C:N ratio is again checked, and the process may be repeated. After a couple of iterations, the mixture is normally acceptable. Figure 10–33 is a mixture design process flow chart that outlines the iterative procedure necessary in determining the compost recipe.

The iterative process of the compost mix design can be summarized to a series of steps to determine the

Figure 10-33 Compost mixture design flow chart



compost mix design. These steps follow the mixture design process flow chart shown in figure 10-33.

Step 1: Determine the amount of bulking agent to **add.** The process normally begins with determining whether or not a bulking agent is needed. The addition of a bulking agent is necessary if the raw waste cannot support itself or if it does not have sufficient porosity to allow internal air movement. A small field trial is the best method to determine the amount of bulking agent required. To do this, a small amount of raw waste would be weighed and incremental quantities of bulking would be added and mixed until the mix has the structure and porosity desired. The wood chips, bark, and shredded tires are examples of bulking agents commonly used.

Step 2: Calculate the moisture content of the compost mix. After the need for and quantity of bulking agent have been determined, the moisture content of the mixture or raw waste should be calculated. Chapter 4 of this handbook gives typical values for moisture content (wet basis) of excreted manure for various animals. Because water is often added as a result of spillage from waterers and in the cleaning processes, raw waste that is to be composted may have significantly higher moisture content than that of "as excreted" manure. If the amount of water added to the manure can be determined, the moisture content of the mix can be calculated using equation 10–11, ignoring the inappropriate terms.

In addition to extra water, feed spillage and bedding material can constitute a major part of the raw waste to be composted. The moisture content for each additive can be determined individually and used to determine the moisture content of the entire mix (equation 10-11). A sample of the raw waste (including the bedding, wasted feed, and water) can also be taken, weighed, dried, and weighed again to determine the moisture content of the mix. Using this procedure the moisture content can be calculated as follows:

$$M_i = \frac{\text{Wet weight - Dry weight}}{\text{Wet weight}} \times 100$$
 [10-10]

M_i = Percent moisture content (wet basis)

Note: To avoid confusion and repetition, the combination of "as excreted" manure, bedding, water, and bulking agent will be referred to as the "compost mix."

The general equation for the moisture content of the compost mix is as follows. (The equation may contain variables that are not needed in every calculation.)

$$\boldsymbol{M}_{M} = \frac{\left(\boldsymbol{W}_{w} \times \boldsymbol{M}_{w}\right) + \left(\boldsymbol{W}_{b} \times \boldsymbol{M}_{b}\right) + \left(\boldsymbol{W}_{a} \times \boldsymbol{M}_{a}\right)}{100} + \boldsymbol{H}_{2}\boldsymbol{O}$$

$$[10-11]$$

where:

 M_m = Percent moisture of the compost mixture (wet basis), eq. 10–10

 W_w = Wet weight of waste (lb)

 $M_{w}^{"}$ = Percent moisture content of waste (wet basis), eq. 10-10

 W_b = Wet weight of bulking agent (lb) M_b = Percent moisture content of bulking agent (wet basis), eq. 10–10

 W_a = Wet weight of amendment (lb) M_a = Moisture content of amendment (wet basis) H_2O = Weight of water added (lb) = G x 8.36, where G = Gallons of water

 W_m = Weight of the compost mix (lb) including wet weight of waste, bulking agent, amendments, and added water.

Step 2 (continued): Determine the amount of amendment to add, if any, to the compost mix that will result in a final moisture content that is between 40 and 60 percent. If the moisture content of the compost mix is less than 40 percent, adding an amendment is necessary to raise the moisture content to an acceptable level. Water is the amendment that is generally added to raise the moisture content, but an amendment that has a higher moisture content than the desired moisture content of the compost mix is acceptable. It is generally best to begin the composting process when the moisture content is closer to 60 percent because the process of composting elevates the temperature and reduces moisture.

If the moisture content of the compost mix is above 60 percent, the addition of an amendment is necessary to lower the moisture content at or below 60 percent. Straw, sawdust, wood chips, and leaves are commonly used.

Equation 10–12 can be used to determine the amount of amendment to add to lower or raise the moisture content of the compost mix.

$$W_{aa} = \frac{W_{mb} \times (M_{mb} - M_d)}{M_d - M_{aa}}$$
 [10-12]

where:

 W_{aa} = Wet weight of amendment to be added

 W_{mb}^{aa} = Wet weight of mix before adding in amend-

 M_{mh} = Percent moisture of mix before adding amendment

 M_d = Desired percent moisture content of mix (wet bases)

 M_{aa} = Moisture content of amendment added

Note: Equation 10–12 can be used for the addition of water by using:

$$M_{aa} = 100\%$$
 for water.

Step 3: Calculate the C:N ratio. The C:N ratio for the compost mix is calculated from the C:N ratios of the waste, bulking agents, and amendments. Typical values for various selected agricultural wastes are shown in chapter 4 of this handbook. The C:N ratios for various waste products and amendments are also shown in table 10–6. The C:N ratios not reported in the literature can be estimated from the amount of fixed solids (amount of ash left after organic matter is burned off) or the volatile solids and the nitrogen content. Equations 10-13 and 10-14 are used to estimate the C:N ratio from the fixed or volatile solids.

$$%C = \frac{100 - %FS}{1.8}$$
 [10-13a]

$$W_c = \frac{VS}{1.8}$$
 [10–13b]

$$C:N = \frac{\%C}{\%N} = \frac{W_c}{W_n}$$
 [10-14]

where:

%C = Percent carbon (dry basis)

%FS = Percent fixed solids (dry basis)

= Dry weight of carbon = Weight of volatile solids C:N = Carbon to nitrogen ratio

%N = Percent total nitrogen (dry basis)

 W_n = Dry weight of nitrogen

Typical values for nitrogen content of manure are reported in chapter 4 of this handbook, and typical values for percent nitrogen (dry basis) for many agricultural crops are reported in chapter 6. The C:N ratio and nitrogen content of manure and of other amendments are highly variable. Using local values for C:N ratios and nitrogen or testing of the compost constituents is highly recommended. The general equation for estimating the C:N ratio of the compost mix is given by equation 10-15.

$$R_m = \frac{W_{cw} + W_{cb} + W_{ca}}{W_{nw} + W_{nb} + W_{na}}$$
 [10-15]

where:

 $R_m = C:N$ ratio of compost mix

 W_{cw} = Weight of carbon in waste (lb)

 W_{ch}^{-} = Weight of carbon in bulking agent (lb)

 W_{ca}^{cb} = Weight of carbon in amendment (lb) W_{nw} = Weight of nitrogen in waste (lb)

 W_{nb}^{nw} = Weight of nitrogen in bulking agent (lb) W_{na} = Weight of nitrogen in amendment (lb)

The weight of carbon and nitrogen in each ingredient can be estimated using the following equations:

$$W_w = \%N \times W_{drv}$$
 [10–16a]

$$W_n = \frac{W_c}{C \cdot N}$$
 [10–16b]

$$W_c = \%C \times W_{dry} \qquad [10-17a]$$

$$W_c = C: N \times W_n$$
 [10-17b]

 W_{drv} = Dry weight of material in question

The dry weight of material can be calculated using equation 10-18.

$$W_{dry} = W_{wet} \times \frac{100 - M_{wet}}{100}$$
 [10-18]

where:

 W_{wet} = Wet weight of material in question

 M_{wet} = Percent moisture content of material (wet

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Step 3 (continued): Determine the amount of amendment, if any, to add to the compost mix that will result in an initial C:N ratio that is **between 25 and 40.** If the C:N ratio calculated in step 3a is less than 25 or more than 40, the type and amount of amendment to add to the compost mix must be determined. For a compost mix that has a C:N ratio below 25. an amendment should be added that has a C:N ratio higher than the desired C:N ratio. For a compost mix that has a C:N ratio of more than 40, an amendment must be added that has a C:N ratio that is less than the desired C:N ratio.

Equation 10-19 or 10-20 can be used to calculate the weight of amendment to add to achieve a desired C:N ratio.

$$W_{aa} = \frac{W_{nm} \times (R_d - R_{mb}) + 10,000}{N_{aa} \times (100 - M_{aa}) \times (R_{aa} - R_d)}$$
[10-19]

$$W_{aa} = \frac{N_m W_{mb} \times (100 - M_{mb}) \times (R_d - R_{mb})}{N_{aa} \times (100 - M_{aa}) \times (R_{aa} - R_d)} [10-20]$$

where:

 W_{nm} = Weight of nitrogen in compost mix (lb) R_d = Desired C:N ratio R_{mb} = C:N ratio of the compost mix before adding

amendment

 N_{aa} = Percent nitrogen in amendment to be added

 R_{aa} = C:N ratio of compost amendment to be added

 N_m = Percent nitrogen in compost mix (dry basis) M_{mb} = Percent moisture of compost mix before

adding amendment (wet basis), equation

10 - 10

For a compost mix that has a C:N ratio of more than 40, a carbonless amendment, such as fertilizer, can be added to lower the C:N ratio to within the acceptable range. In this special case, the following equation can be used to estimate the dry weight of nitrogen to add to the mix:

$$W_{nd} = \frac{W_{cw} + W_{cb} + W_{ca}}{R_d} - \left(W_{nw} + W_{nb} + W_{na}\right) [10-21]$$

where:

 W_{nd} = Dry weight of nitrogen to add to mix

After the amount of an amendment to add has been determined to correct the C:N ratio, the design process then returns to step 2. If no change is necessary in steps 2 and 3, the compost mix design process is complete.

(6) Design example 10–7—Compost mix

A dairy farmer wishes to compost the waste generated from the herd in the barn. The waste is scraped daily from the barn and contains straw as a bedding material, but no extra water is added. Straw is the cheapest and most abundant source of a high C:N ratio amendment on the farm. The 100 cow herd is in the barn for an average of 6 hours. The average weight of a cow is 1,200 pounds. Ten 60-pound bales of straw (chopped) are added daily for bedding. It has been determined that in this case no bulking agent is necessary to improve the compost porosity or structure. Determine the design mix for the compost operation on a daily basis.

Given:

Wheat straw:

Moisture content = 15% (estimated)

= 80 (from table 10-6) C:N ratio

Percent N = 0.67% (from chapter 6 of this

handbook)

Manure:

Nitrogen production

Number of cows = 100Size of cows = 1,200 lb

Number of animal units (AU) = $100 \times 1,200/1,000 =$

Moisture content = 87.5% (from chapter

4 of this handbook)

Manure production = 80 lb/day/1000 lb

> (from chapter 4 of this handbook)

Fraction in barn = 6 hrs/24 hrs = 0.25

> = 0.45 lb/1000 lb/day(from chapter 4 of

this handbook)

Volatile solids = 8.5 lb/1000 lb/day

> (from chapter 4 of this handbook)

Step 1: Bulking agent. A sample of the manure was stacked, and the manure appeared to have sufficient porosity to allow air movement and had the ability to support itself. Therefore, the addition of a bulking agent is not necessary.

Step 2a: Determine the moisture content of the waste. To determine the quantity of waste:

Manure in barn:

$$120 \text{AU} \times 80 \text{ lb} / \text{day} \times 0.25 = 2,400 \text{ lb}$$

Weight of straw added daily:

10 bales
$$\times$$
 60 lb = 600 lb

Weight of manure and straw (W_m):

10 bales
$$\times$$
 60 lb = 600 lb

Using equation 10–11, determine the moisture content of manure plus straw.

$$M_n = \frac{(2,400 \times 87.5) + (600 \times 15)}{100} \times 100 = 73\%$$
3.000 lb

Step 2 (continued): Using equation 10–12, determine the amount of straw to add to bring the moisture content of the compost mix to 60 percent.

$$W_{aa} = \frac{3,000 \text{ lb} \times (73\% - 60\%)}{60\% - 15\%} = 867 \text{ lb}$$

$$W_{\rm m} = 3,000 \text{ lb} + 867 = 3,867 \text{ lb}$$

New weight of compost mix:

Step 3: Determine the C:N ratio of the compost mix. Determine the carbon and nitrogen content of the straw:

Total weight of straw:

$$600 \text{ lb} + 867 \text{ lb} = 1,467 \text{ lb}$$

Straw dry weight (equation 10–18):

1,467 lb +
$$\frac{(10-15)}{100}$$
 = 1,247.9 lb

Weight of nitrogen in straw:

$$W_{na} = \frac{(0.67 \times 1,247.9 \text{ lb})}{100} = 8.4 \text{ lb}$$

Weight of carbon in straw (equation 10-17b):

$$W_{ca} = 8.4 \times 80 = 672.0$$
 lb

Determine the carbon and nitrogen content in manure:

Volume of volatile solids in barn:

$$120 \text{AU} \times 8.5 \text{ lb} / \text{day} / \text{AU} \times 0.25 = 2.55 \text{ lb}$$

Weight of carbon in manure (using equation 10–13b):

$$\frac{255 \text{ lb}}{1.8}$$
 = 141.7 lb

Weight of nitrogen in manure:

$$120$$
AU $\times 0.45 \times 0.25 = 13.5$ lb

C:N ratio of manure:

$$\frac{141.7}{13.5} = 10.5$$

Determine C:N ratio of mixture (equation 10–15):

$$C:N = \frac{141.7 \text{ lb} + 672.0 \text{ lb}}{13.5 \text{ lb} + 8.4 \text{ lb}} = 37.2$$

A compost mix that has a C:N ratio of 37.2 is in the acceptable range, but for purposes of this example, continue step 3.

Step 3 (continued): Determine the type and amount of amendment to add to bring the C:N ratio of the mix to 30:1. To lower the C:N ratio, an amendment with a C:N ratio that is less than the desired final C:N ratio is necessary. Fresh manure that has a C:N ratio of 10.5 could be collected outside the barn, or fertilizer could be added to the mix. The farmer would like to see both alternatives.

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Weight of nitrogen in current compost mix:

$$(13.5 \text{ lb} + 8.4 \text{ lb}) = 21.9 \text{ lb}$$

Dry weight of manure (equation 10-18):

$$2,400 \times \frac{(100 - 87.5)}{100} = 300 \text{ lb}$$

Percent nitrogen in manure:

$$\frac{13.5}{300} \times 100 = 4.5\%$$

Pounds of manure to add to bring mix to 30:1 (using equation 10–19):

$$W_{aa} \frac{21.9 \times (30.0 - 37.2) \times 10,000}{4.5 \times (100 - 87.5) \times (10.5 - 30)} = 1,437 \text{ lb}$$

Pounds of nitrogen to add to bring compost mix to 30:1 (using equation 10–21)

$$W_{nd} \frac{141.7 + 672}{31} - (13.5 + 8.4) = 5.2 \text{ lb}$$

Adding 5.2 pounds of nitrogen is easier than adding 1,437 pounds of manure, so the obvious choice is to add nitrogen. If the farmer chooses to add nitrogen, no further calculations are necessary, because the moisture content of the mix is not changed with the addition of nitrogen. The design process would continue with step 2 if another type of amendment was added that resulted in a change in the moisture content of the manure.

The final compost mix consists of the following:

Waste scraped from the barn — 3,000 lb Additional straw to correct moisture — 867 lb Nitrogen added to lower C:N ratio — 5.2 lb

(7) Design example 10-8

A grass seed farmer wishes to compost straw from rye grass seed harvest. A nearby dairy operation has agreed to furnish fresh manure for 2 weeks. Determine the compost mixture design.

Given:

Rye grass straw:

Amount = 600 tons
Moisture content = 7%
N per ton = 6 lb
C:N ratio = 100:1

Manure:

Number of cows = 400 Size of cows = 1,400 lb

 $\begin{array}{lll} \mbox{Number of animal units (AU)} &= 400 \ x \ 1400/1000 = 560 \\ \mbox{Manure production} &= 80 \ lb/day/1000 \ lb \\ \mbox{Nitrogen production} &= 0.43 \ lb/day/1000 \ lb \\ \mbox{Fixed solids} &= 1.5 \ lb/day/1000 \ lb \\ \end{array}$

Percent moisture = 87.5%

Step 1: No bulking agent is needed to improve structure or porosity.

Step 2: Determine moisture content of rye grass straw and manure mixture:

Straw weight:

$$600 \text{ tons} \times 2000 \text{ lb} / \text{ton} = 1,200,000 \text{ lb}$$

Manure weight:

$$560 \text{ AU} \times 80 \text{ lb / day / AU} \times 14 \text{ days} = 67,200 \text{ lb}$$

Moisture content (M_m) of straw and manure (equation 10–11):

$$\frac{\left(1,200,000\times7\right) + \left(627,000\times87.5\right)}{100} \times 100 = 34.6$$

The 34.6 percent moisture content of the mix is less than 40 percent; therefore, water needs to be added to bring the moisture content to 50 percent.

Step 2 (continued): Using equation 10–12, determine the amount of water to add to bring the moisture content to 50 percent (W_{aa}) .

$$\frac{\left(1,200,000 \times 627,200\right) \times \left(34.6 - 50\right)}{50 - 100} = 562,778 \text{ lb}$$
$$\frac{562,778}{8.33 \text{ lb / gal}} = 67,560 \text{ gal}$$

Step 3: Determine C:N ratio of the straw and manure mix. Determine the amount of carbon and nitrogen in the rye straw:

Nitrogen in straw:

$$W_{na} = 600 \text{ ton} \times 6 \text{ lb} / \text{ton} = 3,600 \text{ lb}$$

Carbon in straw (equation 10-17b):

$$W_{ca} = 3,600 \text{ lb} - (\text{N}) \times 100 = 360,000 \text{ lb}$$

Determine the amount of carbon and nitrogen in the manure:

Nitrogen in manure (use chapter 4 values for N):

$$560 \text{AU} \times 0.45 \times 14 \text{ days} - 3,528 \text{ lb}$$

Assume a 20 percent loss of nitrogen in handling manure. Nitrogen left in manure:

$$W_{nw} = 3,528 \times \frac{100 - 20}{100} = 2,822 \text{ lb}$$

Volume of solids in manure (use chapter 4 values):

$$560 \text{AU} \times 8.5 \times 14 \text{ days} - 66,640 \text{ lb}$$

Carbon in manure (using equation 10–13b):

$$W_{cw} = \frac{66,640 \text{ lb}}{1.8} = 37,022 \text{ lb}$$

C:N ratio of straw and manure mix (equation 10-15):

$$\frac{360,000 + 37,022}{3,600 + 2,822} = 62:1$$

A C:N ratio of 62:1 is more than the maximum recommended of 40:1. The compost mix needs more nitrogen.

Step 3 (continued): Determine the amount of commercial nitrogen to add to the mix to bring the C:N ratio to 40:1.

Amount of nitrogen to add (equation 10-21):

$$N_a = \frac{36,000 + 37,022}{40} - (3,600 + 2,822)$$

= 3,504 lb

The final design mix is:

Rye grass straw = 600 tons Manure (14 days) = 313.6 tons Commercial nitrogen = 3,504 lb

(8) Composting operational considerations

The landowner/operator should be provided a written set of instructions as a part of the waste management plan. These instructions should detail the operation and maintenance requirements necessary for successful composting operation. They should include the compost mix design (recipe), method or schedule of turning or aerating, and instructions on monitoring the compost process and on long-term storage compost. The final use of the compost should be detailed in the Waste Utilization Plan.

(i) Composting time—One of the primary composting considerations is the amount of time it takes to perform the composting operation. Composting time varies with C:N ratio, moisture content, climate, type of operation, management, and the types of wastes and amendments being composted. For a well managed windrow or static pile composting operation, the composting time during the summer months ranges from 14 days to a month. Sophisticated in-vessel methods may take as little as 7 days to complete the composting operation. In addition to the actual composting time, the amount of time necessary for compost curing and storage should be considered.

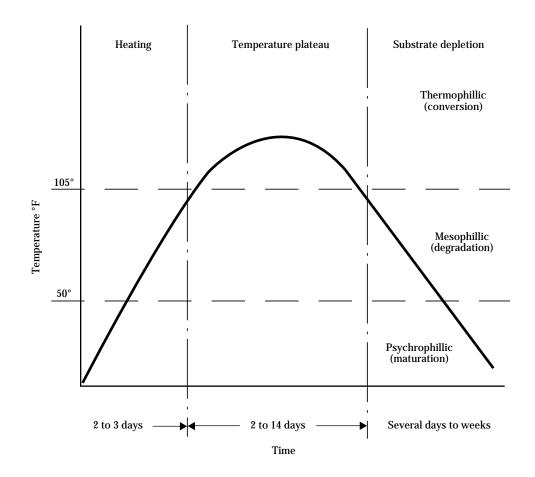
(ii) Temperature—Consideration should be given to how the compost temperature is going to be monitored. The temperature probe should be long enough to penetrate a third of the distance from the outside of the pile to the center of mass. The compost temperature should be monitored on a daily basis if possible. The temperature is an indicator of the level of microbial activity within the compost. Failure to achieve the desired temperatures may result in the incomplete destruction of pathogens and weed seeds and can cause fly and odor problems.

Initially, the compost mass is at ambient temperature; however, as the micro-organisms multiply, the temperature rises rapidly.

The composting process is commonly grouped into three phases based on the prominent type of bacteria present in the compost mix. Figure 10-34 illustrates the relationship between time, temperature, and compost phase. If the temperature is less than 50 °F, the compost is said to be in the psychrophillic stage. If it is in the range of 50 °F to 105 °F, the compost is in the mesophillic stage. If the compost temperature exceeds 105 °F, the compost is in the thermophillic stage. For complete pathogen destruction, the compost temperature must exceed 135 °F.

The compost temperature will decline if moisture or oxygen is insufficient or if the food source is exhausted. In compost methods where turning is the

Figure 10-34 Composting temperature



method of aerating, a temperature rhythm often develops with the turning of the compost pile (fig. 10-35).

(iii) **Moisture**—The moisture content of the compost mixture should be monitored periodically during the process. A low or high moisture content can slow or stop the compost process. A high moisture content generally results in the process turning anaerobic and foul odors developing. A high temperature drives off significant amounts of moisture, and the compost mix may become too dry, resulting in a need to add water.

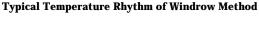
(iv) **Odor**—The odor given off by the composting operation is a good indicator of how the compost operation is proceeding. Foul odors may mean that the process has turned from aerobic to anaerobic. Anaerobic conditions are the result of insufficient oxygen in the compost. This may be caused by excessive moisture in the compost or the need for turning or aerating of the compost.

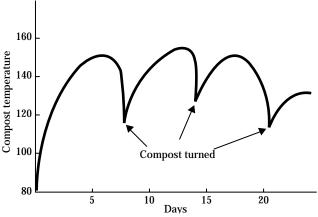
(9) Compost process steps

The composting operation generally follows these steps (fig. 10–36):

(i) Preconditioning of materials (as needed)— Grinding or shredding of the raw material may be necessary to increase the exposed surface area of the compost mixture to enhance decomposition by microorganisms.

Figure 10–35 Typical temperature rhythm of windrow method





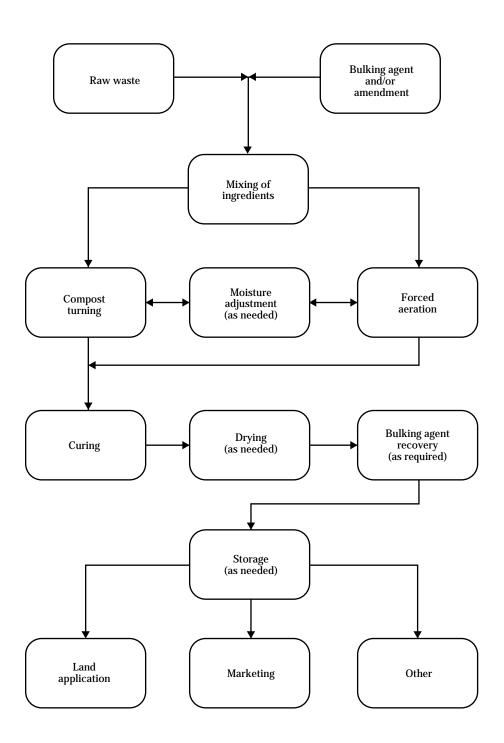
(ii) Mixing of the waste with a bulking agent or **amendment**—A typical agricultural composting operation involves mixing the raw waste with a bulking agent or amendment, or both, according to a prescribed mix or design. The prescribed mix should detail the quantities of raw waste, amendments, and bulking agents to be mixed. The mixing operation is generally done with a front-end loader on a tractor, but other more sophisticated methods can be used.

(iii) Aeration by forced air or mechanical turning—Once the materials are mixed, the composting process begins. Bacteria begin to multiply and consume carbon and free oxygen. To sustain microbial activity, air must be added to the mix to re-supply the oxygen to the compost pile. Air can be added by simply remixing or turning the compost pile. With more sophisticated methods, such as an aerated static pile, air is forced or sucked through the compost mix using a blower. The pounds of air per pound of volatile matter per day generally range from 5 to 9. Given in percentage, the optimum oxygen concentration of the compost mixture ranges from 5 to 15 percent, by volume. An increase of oxygen beyond 15 percent generally results in a decrease in temperature because of greater air flow. Low oxygen concentrations generally result in anaerobic conditions and slow processing times. Inadequate aeration results in anaerobic conditions and increased odors. Odor is an excellent indicator of when to turn and aerate a compost pile.

(iv) Moisture adjustment (as needed)—Water should be added with caution because too much moisture can easily be added. A compost mix that has excessive moisture problems does not compost properly, appears soggy and compacted, and is not loose and friable. Leachate from the compost mixture is another sign of excessive moisture conditions.

(v) Curing (optional)—Once the compost operation is completed, it can be applied directly to the field or stored and allowed to cure for a period of months. During the curing process, the compost temperature returns to ambient conditions and the biological activity slows down. During the curing phase, the compost nutrients are further stabilized. The typical curing time ranges from 30 to 90 days, depending on the type of raw material and end use.

Figure 10-36 Agricultural composting process flow



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(vi) Drying (optional)—Further drying of the compost to reduce weight may be necessary if the finished compost is to be marketed, hauled long distances, or used as bedding. Drying can be accomplished by spreading the compost out in warm, dry weather or under a roofed structure until a sufficient quantity of moisture evaporates.

(vii) Bulking agent recovery (as needed or required)—If such bulking agents as shredded tires or large wood chips are used in the compost mixture, they can be recovered from the finished compost by screening. The recovered bulking agents are then reused in the next compost mix.

(viii) Storage (as needed)—Finished compost may need to be stored for a period of time during frozen or snow-covered conditions or until the compost product can be marketed. If possible, finished compost should be covered to prevent leaching or runoff.

(10) Dead animal composting

The disposal of dead animals is a major environmental concern. Composting can be an economical and environmentally acceptable method of handling dead animals. This process produces little odor and destroys harmful pathogens. Composting of dead poultry is the most common process. The process does apply equally well to other animals. Some operators have composted dead animals weighing as much as 100 pounds by grinding or cutting them into smaller pieces.

Composting of dead animals should be considered when—

- A preferred use, such as rendering, is not available.
- The mortality rate as a result normal animal production is predictable.
- Sufficient land is available for nutrient utilization.
- State or local regulations permit dead animal composting.
- Other disposal methods are not permitted or desired.
- Marketing of finished compost is feasible.

(i) Special planning considerations—Because composting of dead animals is similar in many ways to other methods of composting, the same siting and planning considerations apply. These considerations will not be repeated here. Composting of dead animals does, however, have unique problems that require special attention.

Many States and localities regulate the disposal of dead animals. A construction permit may be required before installation of the facility begins, and an operating permit may be necessary to operate the facility. The animal producer is responsible for procuring all necessary permits to install and operate the facility.

The size of the animals to be composted should be considered when planning a compost facility. Larger animals require additional equipment, labor, and handling to cut the animals into smaller pieces to facilitate rapid composting.

Dead animal composting facilities should be roofed to prevent rainfall from interfering with the compost operation. Dead animal composting must reach a temperature in excess of 130 °F to destroy pathogens. The addition of rainfall can elevate the moisture content and result in a compost mix that is anaerobic. Anaerobic composting takes much longer and creates odor problems.

(ii) Sizing dead animal composting facilities—A typical dead animal composting facility consists of two stages. The first stage, also called the primary composter, is made up of equally sized bins in which the dead animals and amendments are initially added and allowed to compost. The mixture is moved from the first stage to the second stage, or secondary digester, when the compost temperature begins to decline. The second stage can also consist of a number of bins, but it is most often one bin or concrete area or alley that allows compost to be stacked with a volume equal to or greater than the sum of the first stage bins.

The design volume for each stage should be based on peak disposal requirements for the animal operation. The peak disposal period normally occurs when the animals are close to their market weight. The volume for each stage is calculated by multiplying the weight of dead animals at maturity times a volume factor. The volume factor (VF) can vary from 1.0 to 2.5 cubic feet

per pound, depending on the type of composter, local conditions, and experences. Equation 10–22 can be used to calculate the volume for each stage in the compost facility.

$$Vol = B \times \frac{M}{T} \times W \times \frac{VF}{100}$$
 [10-22]

where:

Vol = Volume required for each stage (ft³)

B = Number of animals

M = Percent normal mortality of animals for the entire life cycle expressed as percent

T = Number of days for animal to reach market weight (days)

W = Market weight of animals (lb)

VF = Volume factor

Note: M/T is used to estimate the percentage of dead animals to be composted at maturity. Other estimators or field experience may be more accurate.

The number of bins required for the first and second stages can be estimated to the nearest whole number by dividing the total volume required by the volume of each bin (equation 10–23).

Bins =
$$\frac{\text{Total 1st stage volume } \left(\text{ft}^3\right)}{\text{volume of } sin \text{gle bin } \left(\text{ft}^3\right)}$$
 [10-23]

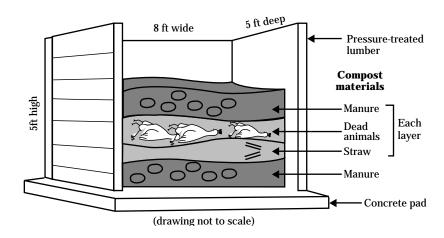
Bins are typically 5 feet high, 5 feet deep, and 8 feet across the front. The width across the front should be sized to accommodate the equipment used to load and unload the facility. To prevent spontaneous combustion and to allow for ease of monitoring, a bin height of no more than 6 feet is recommended. The depth should also be sized to accommodate the equipment used.

A high volume to surface area ratio is important to insulate the compost and allow the internal temperature to rise. The bin height and depth should be no less than one-half the width. Shallow bins are easier to unload and load; therefore the bin depth should be no more than the width. Figure 10–37 is an example of a dead animal composting bin.

Mortality rates vary considerably because of climate and among varieties, species, and types of operation. Information provided by the animal producer/operator should be used whenever possible. Table 10–7 gives typical mortality rates, flock life, and market weights for poultry.

(iii) Mix requirements—Rapid composting of dead animals occurs when the C:N ratio of the compost mix is maintained between 10 and 20. This is considerably lower than what is normally recommended for other types of composting. Much of the nitrogen in the dead

Figure 10-37 Dead animal composting bin



animal mass is not exposed on the surface; therefore, a lower C:N ratio is necessary to ensure rapid composting with elevated temperatures. If the dead animals are shredded or ground up, a higher C:N ratio of 25:1 would be more appropriate. The initial compost mix should have a C:N ratio that is between 13 and 15. As composting proceeds, nitrogen, carbon, and moisture are lost. Once composting is complete, the C:N ratio should be between 20 and 25. A C:N ratio of more than 30 in the initial compost mixture is not recommended because excessive composting time and failure to achieve the temperature necessary to destroy pathogens may result .

The moisture content of the initial compost mixture should be between 45 and 55 percent, by weight, to facilitate rapid decomposition. An initial moisture content of more than 60 percent would be excessively moist and would retard the compost process. The most common problem in dead animal composting is the addition of too much water. Depending on the mass of dead animals and the moisture content of the amendments, water may not need to be added to the

Table 10-7	Poultry 1	mortality r	ates	
Poultry type	Loss rate	Flock life (days)	Cycles per year	Market weight (lb)
Broiler	4.5-5.5	42-49	5.5-6.0	4.2
Roaster				
females	3	42	4	4.0
males	8	70	4	7.5
Laying hens	14	440	0.9	4.5
Breeding hens	10-12	440	0.9	7–8
Breeder males	20-25	300	1.1	10–12
Turkey females	5-6	95	3	14
Turkey male	9	112	3	24
Turkey feather pro	od. 12	126	2.5	30

initial mix. Because water is relatively dense compared to the compost mix, the addition of a little water can raise the moisture content of the mix considerably. Even though water may not need to be added to the initial mix, it is advisable to have a source of water available at the compost site for temperature control.

Composting of dead animals should remain aerobic at all times throughout the process. Anaerobic conditions result in putrid odors and may not achieve temperatures necessary to destroy pathogens. Foul odor during the compost process indicates that the compost process has turned anaerobic and that corrective action is needed. These actions will be discussed later. To prevent the compost process from going anaerobic, the initial mix should have enough porosity to allow air movement into and out of the compost mix. This can be accomplished by layering dead animals and amendments in the mix. For example, a dead poultry compost mix would be layered with straw, dead birds, and manure or waste cake from the poultry houses. Layers of such high porosity material as straw, wood chips, peanut hulls, and bark allow lateral movement of air in the compost mix. Figure 10-38 is an example of commonly recommended layering of manure, straw, and dead poultry.

Table 10–8 is a typical recipe for composting dead birds. The ingredients are presented by volume as well as weight.

Research and evaluation on composting dead animals other than poultry is limited. The differences between livestock and poultry as related to composting are insignificant except for the size of the animal to be

Table 10-8	Broiler compost mix				
Ingredient	Volumes (parts)	Weights (parts)			
Straw	1.0	0.1			
Broiler	2.0	1.0			
Manure	2.0	1.5			
Water*	0.5	0.75			

^{*} More or less water may be necessary depending on the moisture content of the straw and manure.

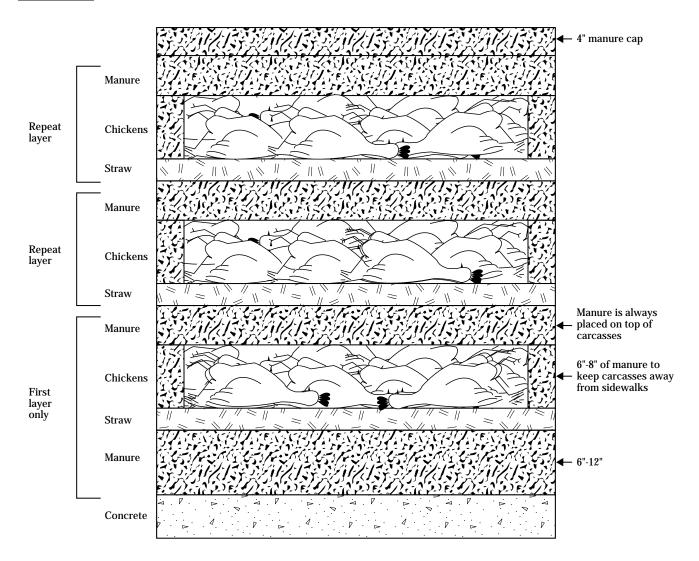
composted and the density of skeletal material. Large birds, such as turkeys, have been successfully composted. If large animals are to be composted, they should be cut into no larger than 15-pound pieces and be cut in a manner to maximize surface exposure. Large animal composting is a promising technology, but it is not well documented. Caution is advised.

(iv) Operational considerations—Efficient and rapid composting requires careful control of the C:N ratio, percent moisture and aerobic conditions, and the internal temperature of the compost mix. A deficiency in any of these three areas retards and possibly

inhibits the composting process achieving temperatures too low for pathogen destruction. Careful planning and monitoring is required to ensure that the process is proceeding as expected.

The landowner/operator should be provided a written set of instructions as a part of the waste management plan that detail the operation and maintenance requirements necessary for successful dead animal composting. The instructions should include compost mix design (recipe), method or schedule of when to unload the primary digester (first stage) and load the secondary digester (second stage), methods to moni-

Figure 10-38 Recommended layering for dead bird composting



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tor the compost process, and information on long-term compost storage. The final utilization of the compost should be detailed in the Waste Utilization Plan.

Temperature is an important gauge of the progress of the composting operation. After initial loading into the first stage, the compost temperature should peak between 130 and 140 degrees in 5 to 7 days. The same is true for when the compost is moved and stacked in the second stage. Elevated temperatures are necessary to destroy the fly larvae, pathogenic bacteria, and viruses. The two-stage process maximizes the destruction of these elements.

When the compost is initially loaded into the compost bin, the internal temperature begins to rise as a result of bacterial activity. Maximum internal temperatures within the first stage should exceed 130 °F within a few days. Although internal compost temperatures rise to a level necessary for the destruction of pathogenic organisms and fly larvae, the temperatures near the edge of the compost pile will not be sufficient to destroy these elements. The edge of the compost stack in the first stage may remain an incubation area for fly larvae and allow the survival of the more heat-resistant pathogens.

Removing the compost from the first stage and restacking in the second stage mixes and aerates the compost. The compost that was on the edge of the compost pile is mixed with the internal compost material, and subsequently is exposed to temperatures in excess of 130 $^{\circ}$ F in the second stage stack.

The internal temperature of the compost in the first and second stages should be monitored on a daily basis. The compost should be moved from the first stage to the second stage when the internal temperature of the first stage compost begins to decline. This generally occurs after 5 to 7 days.

If internal temperatures fail to exceed $130\,^{\circ}F$ in the first or second stages of the composter, the compost material should immediately be incorporated if land applied or remixed and composted a second time.

Excessively high temperatures are also a danger in dead animal composting because spontaneous combustion of the compost material can occur when the compost temperature exceeds 170 $^{\circ}$ F. If the temperature exceeds 170 $^{\circ}$ F, the compost should be removed

from the bin and spread out in a uniform layer no more than 6 inches deep. Water should be used, if necessary, to further cool the compost. Once the temperature has fallen to a safe level, the compost can be restacked. Adding moisture to the compost should retard the biological growth and reduce the temperature. Excessive applications of water stops the process and can cause anaerobic conditions to develop. The compost mix should be rehydrated to a moisture content of 55 to 65 percent, by weight, to reduce excessive temperatures.

Anaerobic conditions may develop if the initial porosity of the compost mix is too low, excessive amounts of water are added to the mix, or the C:N ratio is excessively low. Odor generally is a good indicator of anaerobic conditions. If foul odors develop, the reason for the odor problem must be identified before corrective action can be taken. Anaerobic conditions may be the result of any one or a combination of excessive moisture, low porosity, or low C:N ratio.

(g) Mechanical separation

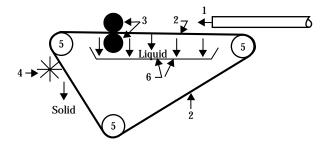
Animal manure contains material that can often be reclaimed. Much of the partly digested feed grain can be recovered from manure of poultry and livestock fed high grain rations. This material can be used as a feed ingredient for other animals. Solids in dairy manure from animals fed a high roughage diet can be removed and processed for use as good quality bedding. Some form of separation must be used to recover these solids. Typically, a mechanical separator is employed. Separators are also used to reduce solids content and required storage volumes.

Separators also facilitate handling of manure. For example, solid separation can allow the use of conventional irrigation equipment for land application of the liquids. Separation eliminates many of the problems associated with the introduction of solids into waste storage ponds and treatment lagoons. For example, it eliminates the accelerated filling of storage volumes with solids and also minimizes agitation requirements.

Several kinds of mechanical separators can be used to remove by-products from manure (fig. 10–39). One kind commonly used is a screen. Screens are statically inclined or in continuous motion to aid in separation. The most common type of continuous motion screen is a vibrating screen. The TS concentration of manure

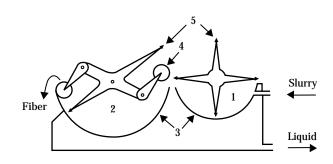
Figure 10-39 Schematic of mechanical solid-liquid separators

Flat belt separator



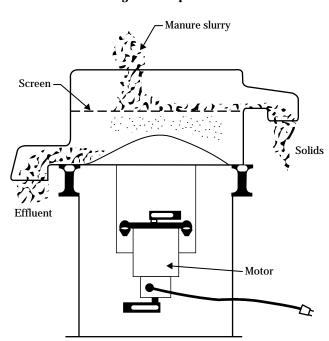
- Slurry input
 Polyester mesh belt
 Press rollers
- 4 Rotary brush
- 5 Belt guide rollers
- 6 Liquid collection trough

Roller-press separator

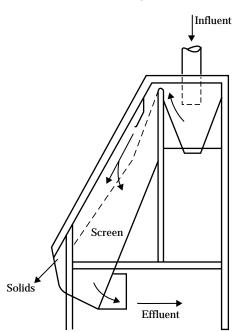


- Screening stage
- Roller pressing stage
- Screens
- Spring loaded press roller
- 5 Brushes

Vibrating screen separator



Stationary inclined screen separator



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to be processed by a screen should be reduced to less than 5 percent. Higher TS concentrations reduce the effectiveness of the separator.

A centrifuge separator uses centrifugal force to remove the solids, which are eliminated from the machine at a different point than the liquids. In addition, various types of presses can be used to force the liquid part of the waste from the solid part.

Several design factors should be considered when selecting a mechanical separator. One factor is the amount of liquid waste that the machine can process in a given amount of time. This is referred to as the "throughput" of the unit. Some units have a relatively low throughput and must be operated for a long time. Another very important factor is the TS content required by the given machine. Centrifuges and presses can operate at a higher TS level than can static screens.

Consideration should be given to handling the separated materials. Liquid can be collected in a reception pit and later pumped to storage or treatment. The separated solids will have a TS concentration of 15 to 40 percent. While a substantial amount of nutrients are removed with the solids, the majority of the nutrients and salt remain in the liquid fraction. In many cases water drains freely from piles of separated solids. This liquid needs to be transferred to storage to reduce odors and fly breeding.

Typically, solids must still be processed before they can be used. If they are intended for bedding, the material should be composted or dried. If the solids are intended for animal feed, they may need to be mixed with other feed ingredients and ensiled before feeding to prevent bacteriological disease transmission. A feed ration using manure must be proportioned by an animal nutritionist so that it is both nutritious and palatable.

A planner/designer needs to know the performance characteristics of the separator being considered for the type of waste to be separated. The best data, if available, would be that provided by the separator manufacturer. If that data is not available, the manufacturer or supplier may agree to demonstrate the separator with waste material to be separated. This can also provide insight as to the effectiveness of the equipment.

If specific data on the separator is not available, tables 10–9 and 10–10 can be used to estimate performance characteristics. Table 10–9 gives data for separating different wastes using different separators, and table 10–10 presents general operational characteristics of mechanical separators.

(h) Settling basins

In many situations, removing manure solids, soil, and other material from runoff from livestock operations is beneficial. The most common device to accomplish this is the settling or solids separation basin. A settling basin used in association with livestock operations is a shallow basin or pond that is designed for low velocities and the accumulation of settled materials. It is positioned between the waste source and the waste storage or treatment facilities. Most readily settleable solids will settle from the flow if the velocity of the liquid is below 1.5 feet per second.

The basins should be planned and designed in accordance with SCS Conservation Practice Standard, Sediment Basin, Code 350 (USDA 1978). Settling basins should have access ramps that facilitate removal of settled material. Outlets from settling basins should be located so that sediment removal is not restricted.

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 Table 10-9
 Operational data for solid/liquid separators

Waste type	Separator	TS Raw	TS concentration (%) v Separated		% Retained in separated solids				
туре		waste	liquids	solids	TS	VS	COD	N	P
Dairy	Vibrating screen								
v	16 mesh	5.8	5.2	12.1	56	_			_
	24 mesh	1.9	1.5	7.5	70	_			_
	Decanter centrifuge								
	16-30 gpm Static inclined screen	6–8	4.9-6.5	13–33	35–40	_	_	_	
	12 mesh	4.6	1.6	12.2	49				
	32 mesh	2.8	1.0	6.0	68				
- 0		۵.0	1.1	0.0	00				
Beef	Static inclined			40.0					
	screen	4.4	3.8	13.3	15	_			_
	Vibrating screen	1–2	_	_	40–50		_	_	_
Swine	Decanter centrifuge								
	3 gpm Vibrating screen 22 gpm/ft²	7.6	2.6	37	14	_	_	_	_
	18 mesh	4.6	3.6	10.6	35	39	39	22	26
	30 mesh	5.4	3.5	9.5	52	56	49	33	34

 Table 10-10
 Characteristics of solid/liquid separators (Barker 1986)

Characteristic	Decanter centrifuge	Vibrating screen	Stationary inclined screen
Typical screen opening	_	20 mesh	10-20 mesh
Maximum waste TS concentration	8%	5%	5%
Separated solids TS concentration	to 35%	to 15%	to 10%
TS reduction*	to 45%	to 30%	to 30%
COD reduction*	to 70%	to 25%	to 45%
N reduction*	to 20%	to 15%	to 30%
P reduction*	to 25%	_	_
Throughput (gpm)	to 30	to 300	to 1,000

^{*} Removed in separated solids

(i) Dilution

Dilution is often used to prepare the waste to facilitate another function. This involves adding clean water or another waste that has less total solids to the waste, resulting in a waste that has a desired percentage of total solids. A common use of dilution is to prepare the waste to facilitate utilization by land application using a sprinkler system. Figure 10–40 is a design aid for determining the amount of clean dilution water required to lower the TS concentration.

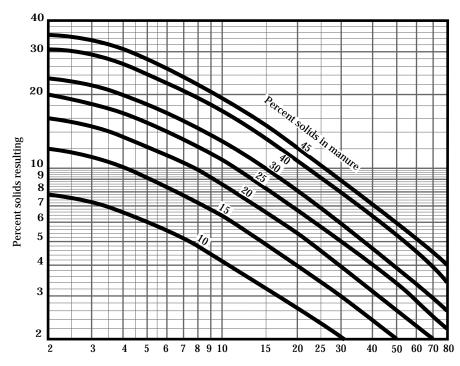
(j) Vegetative filters

A vegetative filter can be a shallow channel or a wide, flat area of vegetation used for removing suspended solids and nutrients from concentrated livestock area runoff and other liquid wastes. The filters are designed with adequate length and limited flow velocities to promote filtration, deposition, infiltration, absorption, adsorption, decomposition, and volatilization of contaminants. Consideration must be given to hydraulic as well as contaminant loading.

Vegetative filters rely on infiltration to remove nitrates and micro-organisms that are in solution because these waste constituents are very mobile in water. Provision for rest periods between loadings is recommended. In cases where a large volume of solids is expected, settling basins are needed above the filter area or channel. "Clean" water must be diverted from the filter. Installation and maintenance are critical.

Vegetative filters are planned and designed according to Conservation Practice Standard, Filter Strip, Code 393 (USDA 1982), which gives more detailed planning considerations and design criteria. See section 651.0605(c) for additional information. If State or local government has restrictions on the use of vegetative filters, the requirements must be met before design and construction. This is especially true if the outflow from the vegetative filter will flow into a stream or waterway. Unless permitted by State regulations, wastewater treatment by vegetative filters is not sufficient to allow discharge to surface water.

Figure 10-40 Design aid to determine quantity of water to add to achieve a desired TS concentration (USDA 1975)



Gallons of water to add per cubic foot of material

651.1005 Transfer

Manure collected from within a barn or confinement area must be transferred to the storage or treatment facility. In the simplest system, the transfer component is an extension of the collection method. More typically, transfer methods must be designed to overcome distance and elevation changes between the collection and storage facilities. In some cases gravity can be used to move the manure. In many cases, however, mechanical equipment is needed to move the manure. Transfer also involves movement of the waste from storage or treatment to the point of utilization. This may involve pumps, pipelines, and tank wagons.

(a) Reception pits

Slurry and liquid manure collected by scraping, gravity flow, or flushing are often accumulated in a reception pit (fig. 10–41). Feedlot runoff can also be accumulated. These pits can be sized to hold all the waste produced for several days to improve pump efficiency or to add flexibility in management. Additional capacity might be needed for extra liquids, such as milk

parlor water or runoff from precipitation. For example, if the daily production of manure and parlor cleanup water for a dairy is estimated at 2,500 gallons and 7 days of storage is desired, then a reception pit that has a capacity of 17,500 gallons (2,500 gallons/day x 7 days) is the minimum required. Additional volume should be allowed for freeboard emergency storage.

Reception pits are rectangular or circular and are often constructed of cast-in-place reinforced concrete or reinforced concrete block. Reinforcing steel must be added so that the walls withstand internal and external loads.

Waste can be removed with pumps or by gravity. Centrifugal pumps can be used for agitating and mixing the manure before transferring the material. Both submersible pumps and vertical shaft pumps that have the motor located above the manure can be used. Diluted manure can be pumped using submersible pumps, often operated with float switches. The entrance to reception pits should be restricted by guard rails or covers.

Debris, such as pieces of metal and wood and rocks, must sometimes be removed from the bottom of a reception pit. Most debris must be removed manually,

Figure 10-41 Reception pit for dairy freestall barn

Earth storage basin

Check valve

Manual valves

Centrifugal pump

Agitation

notzle

Reception pit for dairy freestall barn

but if possible, this should be done remotely from outside the pit. The pit should be well ventilated before entering. If waste is in the pit, a self-contained breathing apparatus must be used. Short baffles spaced around the pump intake can effectively guard against debris clogging the pump.

In cold climates, reception pits need to be protected from freezing. This can be accomplished by covering or enclosing it in a building. Adequate ventilation must be provided in all installations. In some installations, hoppers and either piston pumps or compressed air pumps are used instead of reception pits and centrifugal pumps. These systems are used with semisolid manure that does not flow readily or cannot be handled using centrifugal pumps.

(b) Gravity flow pipes

Liquid and slurry manure can be moved by gravity if sufficient elevation differences are available or can be established. For slurry manure, a minimum of 4 feet of elevation head should exist between the top of the collection pit or hopper and the surface of the material in storage when storage is at maximum design depth.

Gravity flow slurry manure systems typically use 18- to 36-inch diameter pipe. In some parts of the country 4- to 8-inch diameter pipe is used for the gravity transport of low (<3%) TS concentration waste. The planner/designer should exercise caution when specifying the 4- to 8-inch pipe. Smooth steel, plastic, concrete, and corrugated metal pipe are used. Metal pipes should be coated with asphalt or plastic to retard corrosion, depending upon the type of metal. All joints must be sealed so that the pipe is water tight.

Gravity flow pipes should be designed to minimize changes in grade or direction over the entire length. Pipe slopes that range from 4 to 15 percent will work satisfactorily, but 7 to 8 percent slope is preferable. Excessive slopes allow separation of liquids and solids and increase the chance of plugging. The type and quantity of bedding and the amount of milkhouse waste and wash water added have an effect on the flow characteristics and the slope needed in a particular situation. Straw bedding should be discouraged, especially if it is not chopped. Smooth, rounded transition from reception pit to pipe and the inclusion of an air vent in the pipeline aid the flow and prevent plugging.

Figure 10–42 illustrates the use of gravity flow for manure transfer. At least two valves should be located in an unloading pipe. Proper construction and operation of gravity unloading waste storage structures are extremely important. Containment berms should be considered if the contamination risk is high downslope of the unloading facility.

(c) Push-off ramps

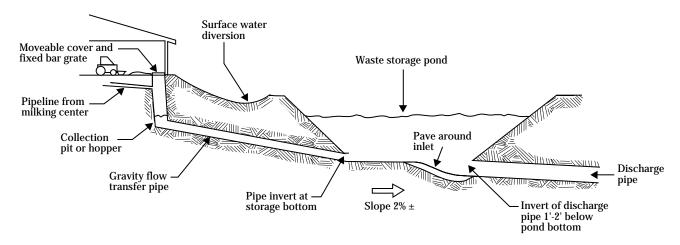
Manure that is scraped from open lots can be loaded into manure spreaders or storage and treatment facilities using push-off ramps (fig. 10–43) or docks. A ramp is a paved structure leading to a manure storage facility. It can be level or inclined and usually includes a retaining wall. A dock is a level ramp that projects into the storage or treatment facility. Runoff should be directed away from ramps and docks unless it is needed for waste dilution. Ramp slopes should not exceed 5 percent. Push-off ramps and docks should have restraints at each end to prevent the scraping tractors from accidentally going off the end.

(d) Picket dams

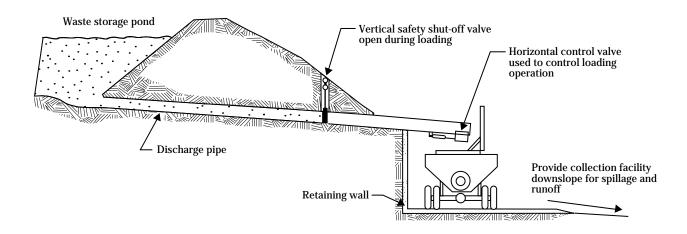
Manure that has considerable bedding added can be stored as a solid or semi-solid. If the manure is stored uncovered, precipitation can accumulate in the storage area. Picket dams can be used to drain runoff from the storage area while retaining the solid manure and bedding within the storage area. Any water drained should be channeled to a waste storage pond. The amount of water that drains from the manure depends on the amount of precipitation and the amount of bedding in the manure. Water will not drain from manure once the manure and water are thoroughly mixed. Picket dams will not dewater liquid manure.

The picket dam should be near the unloading ramp to collect runoff and keep the access as dry as possible. It should also be on the side of the storage area opposite the loading ramp. Water should always have a clear drainage path from the face (leading edge) of the manure pile to the picket dam.

Figure 10-42 Examples of gravity flow transfer



Gravity flow transfer



Gravity flow from storage

The floor of the storage area using a picket dam should have slope of no more than 2 percent toward the dam. Picket dams should be made of pressure-treated timbers that have corrosion resistant fasteners. The openings in the dam should be about 0.75 inch wide vertical slots. Figure 10–44 shows different aspects of picket dam design.

(e) Pumps

Most liquid manure handling systems require one or more pumps to either transport or agitate manure. Pumps are in two broad classifications—displacement and centrifugal. The displacement group are piston, air pressure transfer, diaphragm, and progressive cavity pumps. The first two are used only for transferring manure; however, diaphragm and progressive cavity pumps can be used for transferring, agitating, and irrigating manure.

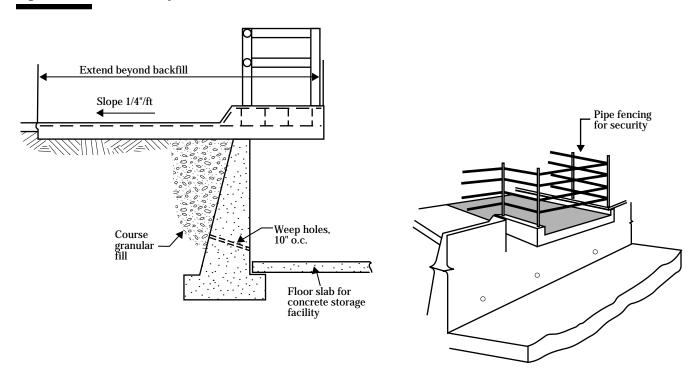
The centrifugal group is vertical shaft, horizontal shaft, and submersible pumps. They can be used for agitation and transfer of liquid manure; however, only vertical and horizontal shaft pumps are used for irrigation because of the head that they can develop.

Pump selection is based on the consistency of the material to be handled, the total head to be overcome, and the desired capacity (pumping rate). Pump manufacturers and suppliers can provide rating curves for a variety of pumps.

(f) Equipment

Other equipment used in the transfer of agricultural wastes include a variety of pumps including chopper/agitator, centrifugal, ram, and screw types. Elevators, pipelines, and hauling equipment are also used. See chapter 12 for information about specific equipment.

Figure 10-43 Push-off ramp



651.1006 Utilization

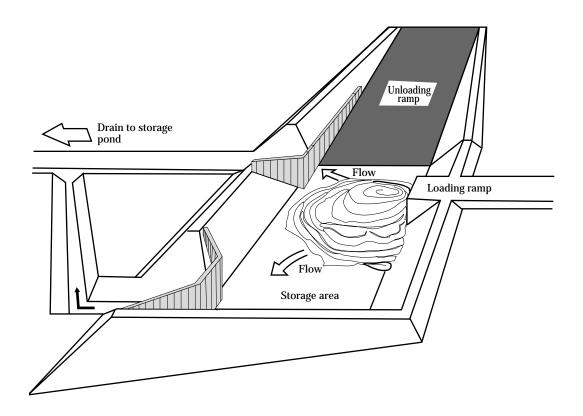
Utilization is a function in a waste management system employed for a beneficial purpose. The typical method is to apply the waste to the land as a source of nutrients for plant growth and of organic matter to improve soil tilth and water holding capacity and to help control erosion. The vast majority of animal waste produced in the United States is applied to cropland, pasture, and hayland. Manure properly managed and applied at the appropriate rates and times can significantly reduce the amount of commercial fertilizer needed for crop production. An anaerobic digester used for biogas production is considered a utilization function component because the waste is being managed for use even though further management of the digester effluent is required.

(a) Nutrient management

Manure should be applied at rates where the nutrient requirements of the crop to be grown are met. Concentration of nutrients in the manure should be known, and records on manure application rates should be maintained.

Between the time of manure production and the time of application, nutrient concentrations can vary widely because of storage, dilution, volatilization, settling, drying, or treatment. To accurately use manure, representative samples of the material to be land applied should be analyzed for nutrient content. Before application rates can be computed, the soil in the fields where manure will be applied should be analyzed and nutrient recommendations obtained. This information should indicate the amount of nutrients to be applied for a given crop yield.

Figure 10-44 Solid manure storage with picket dam



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Scheduling land application of wastes is critical. Several factors must be considered:

- Amount of available manure storage
- Major agronomic activities, such as planting and harvesting
- Weather and soil conditions
- Availability of land and equipment
- Stage of crop growth

A schedule of manure application should be prepared in advance. It should consider the most likely periods when application is not possible. This can help in determining the amount of storage, equipment, and labor needed to make application at desired times.

(b) Land application equipment

Animal waste is land applied using a variety of equipment. The kind of equipment used depends on the TS concentration of the waste. If the manure handles as a solid, a box spreader or flail spreader is used. Solids spreaders are used for manure from solid manure structures and for the settled solids in sediment basins.

Slurry wastes are applied using tank wagons or flail spreaders. Some tank wagons can be used to inject the waste directly into the soil. Slurry spreaders are typically used for waste that is stored in above or below ground storage structures, earthen storage structures, and sometimes lagoons.

Waste that has a TS concentration of less than 5 percent can be applied using tank wagons, or it can be irrigated using large diameter nozzles. Irrigation is used primarily for land application of liquids from lagoons, storage ponds, and tanks. Irrigation systems must be designed on a hydraulic loading rate as well as on nutrient utilization.

Custom hauling and application of manure are becoming popular in some locations. This method of utilization reduces the amount of specialized equipment needed by the owner/operator.

(c) Land application of municipal sludge

Municipalities in the United States treat wastewater biologically using either anaerobic or aerobic processes. These processes generate sludge that has agronomic value as a nutrient source and soil amendment. Land application of sludge is currently recognized as acceptable technology; however, strict regulations and practices must be followed.

(d) Biogas production

Some of this material was taken directly from "Tentative guidelines for methane production by anaerobic digestion of manure" (Fogg 1981).

Liquid manure confined in an air-tight vessel decomposes and produces methane, carbon dioxide, hydrogen sulfide, and water vapor as gaseous by-products. This process is known as anaerobic digestion. Many municipalities use this technique to treat sludge generated in wastewater treatment. Many livestock and poultry producers have become interested in the process because of the potential for onsite energy production.

Biogas, the product of anaerobic digestion, is typically made up of 55 to 65 percent methane (CH $_4$), 35 to 45 percent carbon dioxide (CO $_2$), and traces of ammonia (NH $_3$) and hydrogen sulfide (H $_2$ S). Pure methane is a highly combustible gas that has an approximate heating value of 994 BTU/ft 3 . Biogas can be burned in boilers to produce hot water, in engines to power electrical generators, and in absorption coolers to produce refrigeration.

The most frequent problem with anaerobic digestion systems is related to the economical use of the biogas. The biogas production rate from a biologically stable anaerobic digester is reasonably constant; however, most onfarm energy use rates vary substantially. Because compression and storage of biogas is expensive, economical use of biogas as an onfarm energy source requires that farm use must closely match the energy production from the anaerobic digester.

Because of the presence of hydrogen sulfide, biogas may have an odor similar to that of rotten eggs. Hydrogen sulfide mixed with water vapor can form sulfuric acid, which is highly corrosive. It can be removed from biogas by passing the gas through a column of iron-impregnated wood chips. Water vapor can be removed by condensers or condensate traps. Carbon dioxide can be removed by passing biogas through lime water under high pressure.

Biogas can be used to heat the slurry manure in the digester. From 25 to 50 percent of the biogas is required to maintain a working digester temperature of 95 °F, depending on the climate and the amount of insulation used. Below ground digesters require less insulation than those above ground. Engines can burn biogas directly from digesters; however, removal of hydrogen sulfide and water vapor is recommended.

If digested solids are separated from digester effluent and dried, they make an excellent bedding material. A brief period of composting may be necessary before it is used.

Anaerobic digestion in itself is not a pollution control practice. Digester effluent must be managed similarly to undigested manure by storing in waste storage ponds or treating in lagoons. Initial start-up of a digester is critical. The digester should be partly filled with water (50 to 75 percent full) and brought to temperature using an auxiliary heater. Feeding of the digester with manure should increase over a period of

3 to 6 weeks starting with a feeding rate of about 25 percent of full feed (normal operation).

Biogas production rates can be measured using specially designed corrosion resistant gas meters. These rates and carbon dioxide levels are good indicators of digester health during start-up. Several simple tests can be used in the field to determine carbon dioxide.

(1) Design procedure

Because of the safety issues and economic and operational complexities involved, SCS assistance on biogas production is generally limited to planning and feasibility. The information presented here is intended for that type of assistance. Interested farmers and ranchers should be advised to obtain other assistance in the detailed design of the facility.

The guidelines presented here are based on digestion of manure in the mesophillic temperature range (about 95 °F) and may be subject to change as a result of additional research and experience. They provide a basis for considering biogas production facilities based on current knowledge as part of a waste management system.

Several digester types are used (figs. 10–45, 10–46, 10–47). The mixed tank is a concrete or metal cylindrical vessel constructed aboveground. If the manure is

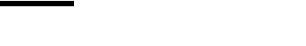
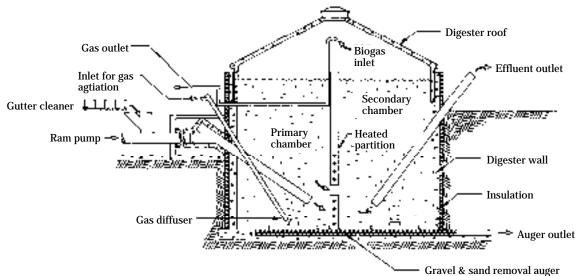


Figure 10–45 Two stage, mixed tank anaerobic digester



highly liquid (low TS), the digester must be periodically mixed to get good digestion. This can be done mechanically using a mechanical mixer, recirculating digestion liquid, or pumping biogas into the bottom sludge to remix the contents of the digester.

Another digester, known as the plug flow, is used for relatively thick manure (12 to 14 percent TS), such as dairy manure. The manure is introduced at one end and theoretically moves as a "plug" to the other end. However, if the TS content of the influent manure is too low, the manure will "channel," the actual retention time will be reduced, and the biogas yield will diminish.

For any digester, the influent must be managed for consistency in frequency of feeding as well as in the VS concentration. For this to happen the rations fed and manure management must be consistent. Some manure requires preprocessing before it enters the digester. For example, poultry manure must be diluted to about 6 percent TS to allow grit to settle before the manure is pumped into the digester. Grit material is very difficult to remove from digesters. All digesters must be periodically cleaned. The frequency of cleaning can vary from 1 to 4 years.

(i) **Determine manure production**—Manure production can be based on the tables in chapter 4 or on reliable local data. The following data will be needed:

Volume of manure produced	=	ft3/day
Wet weight of manure produce	=	lb/day
Total solids (TS)	=	lb/day
Volatile solids (VS)	=	lb/day
Percent solids (TS/wet weight) =	=	percent

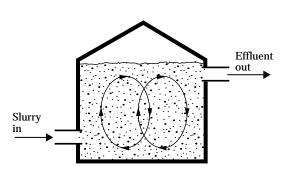
Fresh manure is desirable for digestion. Characteristics of beef feedlot manure must be determined for each operation.

(ii) Establish TS concentration for digester feed—TS concentrations considered desirable as input to the digester can range from about 6 to 12 percent. The following are guidelines:

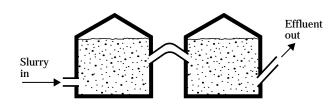
Dairy manure	10 to 12 %
Confined beef manure	10 to 12 %
Beef feedlot manure	8 to 10 %
(after settling grit)	
Swine manure	8 to 10 %
Chicken manure	7 to 9 %

Figure 10-46 Typical anaerobic digester types

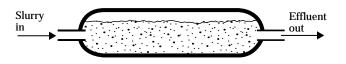
Mixed type digester



Two stage digester



Plug flow digester



These percentages may need to be adjusted to eliminate scum formation and promote natural mixing by the gas produced within the mass. If scum forms, a small increase in percent solids may be desirable. This increase may be limited by pumping characteristics and should seldom go above 12 percent solids.

(iii) Determine effective digester volume—A hydraulic detention time of 20 days is suggested. This time appears to be about optimum for efficient biogas production. The daily digester inflow in cubic feet per day can be determined using equation 10–24.

$$DMI = \frac{TMTS \times 100}{DDFSC \times 62.4}$$
 [10-24]

where:

DMI = Daily manure inflow, ft³

TMTS = Total manure total solids production,

ft3/day

DDSFC = Desired digester input total solids concentration, %

The necessary digester volume in cubic feet can be determined using equation 10–25.

$$DEV = DMI \times 20$$
 [10-25]

where:

DEV = Digester effective volume, ft³

20 = Recommended detention time, days

(iv) Select digester dimensions—Optimum dimensions of the liquid part of the digester volume have not been established. The digester should be longer than it is wide to allow raw manure to enter one end and digested slurry to be withdrawn at the other. An effectively operating digester has much mixing by heat convection and gas bubbles. True plug flow will not occur.

Sufficient depth should be provided to preclude excessive delay at start-up because of the oxygen interchange at the surface. A combination of width equal to about two times the depth and length equal to about four times the depth is a realistic approach. Other

proportions of width and length should work equally well. For the purpose of discussion assume:

$$H = \left(\frac{DEV}{8}\right)^{0.33}$$

$$WI = 2 \times H$$

$$L = 4 \times H$$

where:

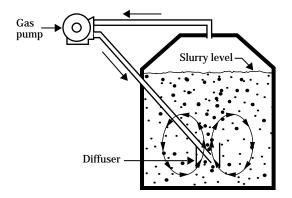
H = height, ft WI = width, ft L = length, ft

Dimensions should be adjusted to round numbers to fit the site and provide economical construction.

(v) Estimate biogas production—Biogas production is dependent on VS destruction within the digester. An efficient digester that has a 20-day retention should reduce VS by 50 percent. Some research indicates a reduction of 55 percent of VS in swine manure and 60 to 65 percent in poultry manure. Biogas production from poultry manure may vary significantly from the estimates presented below. Animals fed a high roughage ration produce less biogas than those fed a high concentrate ration. Estimated VS reductions are:

Dairy and beef	50 %
Swine	55 %
Poultry	60%

Figure 10-47 Gas agitation in an anaerobic digester



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Estimated daily biogas production rates are:

Dairy	12 ft ³ /lb VS destroyed
Beef	10 ft ³ /lb VS destroyed
Swine	13 ft ³ /lb VS destroyed
Poultry	13 ft ³ /lb VS destroyed

Biogas production per day is estimated by multiplying the percent volatile solids reduction times the estimated daily biogas production rate times the daily volatile solids input. Biogas production in cubic feet per day would be:

Dairy	6 x daily VS input
Beef	5 x daily VS input
Swine	7.2 x daily VS input
Poultry	

Initial start-up of a digester requires a period of time for anaerobic bacteria to become acclimated and multiply to the level required for optimum methane production. If available, sludge from a municipal anaerobic digester or another anaerobic manure digester can be introduced to speedup the start-up process. The digester contents must be maintained at about 95 °F for continuous and uniform biogas production. Hot water tubes within the digester can serve this purpose.

(2) Other considerations

Biogas is difficult to store because it can't be compressed at normal pressures and temperatures. Storage pressures above 250 psi are rarely used. Because of these reasons, biogas usage is generally planned to match production, and thus eliminate the need for storage.

The most common use of biogas is the production of electricity using an engine-generator set. The thermal conversion efficiency is about 25 percent for this type of equipment. The remainder of the energy is lost as heat. Heat exchangers can be used to capture as much as 50 percent of the initial thermal energy of the biogas from the engine exhaust gases and the engine cooling water. This captured heat can sometimes be used onsite for heating. Some of it must be used to maintain the digester temperature.

Effluent from anaerobic digesters has essentially the same amount of nutrients as the influent. Some of the organic nitrogen will be converted to ammonia, making it more plant available but more susceptible to volatilization unless the liquid is injected. Only a little volume is lost by processing the manure through an anaerobic digester. For manure requiring dilution before digestion, the amount of liquid to be stored and handled actually increases as compared to the original amount of manure.

(3) Design example 10–9—Biogas digester Mr. Joe Sims of Hamburg, Pennsylvania, has requested assistance on development of an agricultural waste management system for his 100 Guernsey milk cows that weigh an average of 1,200 pounds. He has requested that an alternative be developed that includes an anaerobic digester to produce methane gas. Determine the approximate size of the digester using worksheet 10A–5.

Completed worksheet for Design example 10-9

Decisionmaker: Joe Sims	Date: 6/13/89
H amburg, PA	,
Animal units	
1. Animal type	3. Number of animals (N)
2. Animal weight, lbs (W)1200	4. Animal units, AU = $\frac{W \times N}{1000}$ =
Manure volume 5. Daily volume of daily manure production per AU, ft ³ /AU/day (DVM)= 1.30	7. Total daily manure production volume, ft³/day(TMP)
6. Total volume of daily manure production for animal MPD = AU x DVM $\frac{156}{}$	ıl type, ft ³ /day
Manure total solids 8. Daily manure total solids production, lbs/AU/day (9. Daily manure total solids production for animal tyl MTSD = MTS	pe, lb/day 1200 solids production,
Manure volatile solids 11. Daily manure volatile solids production per AU, Ik	SCIALLIDAN (AANS) 8.5
	1000
13. Total manure volatile solids production, lbs/day ((IMVS)
	Digester feed solid concentration 15. Desired digester feed solids concentration, % (DDFSC) =
13. Total manure volatile solids production, lbs/day (Percent solids 14. Percent solids, % (PS)	Digester feed solid concentration 15. Desired digester feed solids concentration, % (DDFSC) = Digester effective volume 17. Digester effective volume, ft ³
13. Total manure volatile solids production, lbs/day (Percent solids 14. Percent solids, % (PS) $PS = \frac{TMTS \times 100}{TMP \times 62.4} = \frac{(1200) \times 1}{(156) \times 62}$ Daily manure inflow 16. Daily manure inflow, ft ³ $DMI = \frac{TMTS \times 100}{DDFSC \times 62.4} = \frac{(1200) \times 1}{(12) \times 62}$ Digester dimensions	(TMVS) Digester feed solid concentration 15. Desired digester feed solids concentration, % (DDFSC) = Digester effective volume 17. Digester effective volume, ft ³ DEV = DMI x 20 = ($\frac{160.2}{160.2}$) x 20 =
13. Total manure volatile solids production, lbs/day (Percent solids 14. Percent solids, % (PS) $PS = \frac{TMTS \times 100}{TMP \times 62.4} = \frac{(1200) \times 1}{(156) \times 62}$ Daily manure inflow 16. Daily manure inflow, ft ³ $DMI = \frac{TMTS \times 100}{DDFSC \times 62.4} = \frac{(1200) \times 1}{(12) \times 62}$ Digester dimensions	(TMVS) Digester feed solid concentration 15. Desired digester feed solids concentration, % (DDFSC) = Digester effective volume 17. Digester effective volume, ft ³ DEV = DMI x 20 = (160.2) x 20 =

651.1007 Ancillary components

(a) Fences

Fences are an important component in some agricultural waste management systems. They are planned and designed in accordance with Conservation Practice Standard, Fencing, Code 382 (USDA-SCS 1980). As they apply to agricultural waste management, fences are used to:

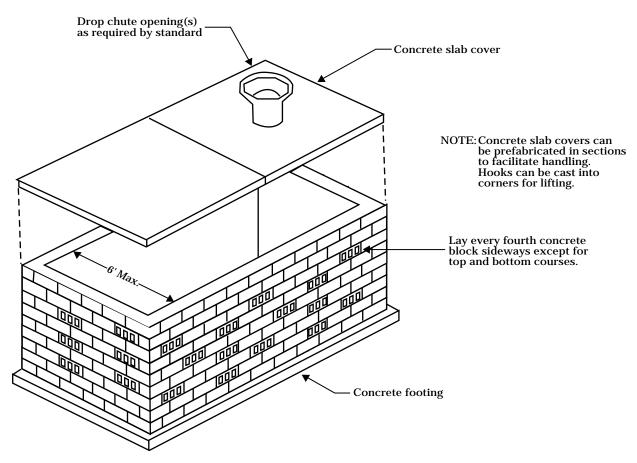
- Confine livestock so that manure can be more efficiently collected.
- Exclude livestock from surface water to prevent direct contamination.

- Provide the necessary distance between the fence and surface water to be protected for the interception of lot runoff in a channel, basin, or other collection or storage facility located above the lot.
- Reduce the lot area and thus reduce the volume of lot runoff to be collected or stored.
- Exclude livestock from hazardous areas, such as waste storage ponds.
- Allow management of livestock for waste utilization purposes.
- Protect vegetative filters from degradation by livestock.

(b) Dead animal disposition

Every livestock and poultry facility experiences loss of animals by death. Regardless of the method used, the disposition of dead animals should be accomplished in

Figure 10-48 Poultry and suckling pig disposal pit constructed with 8" x 8" x 16" concrete blocks



a sanitary manner and in accordance with all State and local laws.

Utilization of the energy contained in the dead animals should be given first consideration. Rendering and composting of dead animals both result in byproducts that can used. Refer to 651.1004(g) for discussion on composting animal carcasses. If utilization is not viable, consideration can be given to disposal by incineration or burial. Incineration can cause odor problems unless an afterburner or excess air system is used.

A common method for onsite dead animal disposal is burial. The burial sites need to be at least 150 feet downgradient from any ground water supply source. Sites that have highly permeable soils, fractured or cavernous bedrock, and a seasonal high-water table are not suitable and should be avoided. In no case should the bottom of the burial pit be closer than 5 feet from the ground water table. Surface water should be diverted from the pit.

For large animals (cattle and mature swine), individual pits should be opened for each occasion of burial. The pits should be closed and marked after burial. For small animals (poultry and small pigs), pits can be constructed for use over a period of time.

Typical pit sizes for small animals are 4 to 6 feet wide, 4 to 12 feet long, and 4 to 6 feet deep. The sides of the pit should be constructed of concrete block, treated timber, or pre-cast concrete. The side walls must have some openings to allow for pressure equalization. The bottom of small animal pits is not lined. The top should be airtight with a single capped opening to allow for adding dead animals. Figure 10–48 illustrates one possible disposal pit configuration.

Disposal pits should have adequate capacity. The recommended capacity for broilers is 100 cubic feet per 10,000 broilers. For small pigs, the capacity is 1 cubic foot per sow. The pit size for layers and turkeys can be determined using figure 10–49.

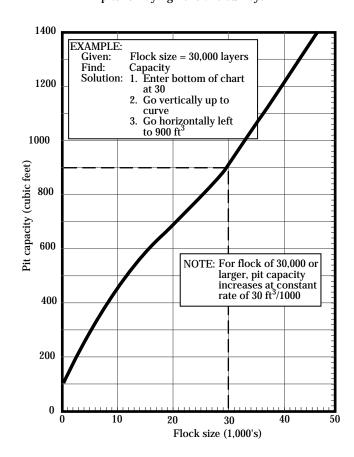
(c) Human waste management

If at all possible, human waste should be treated in municipal facilities designed to provide proper treatment. However, in many rural areas this is not possible. Septic tank systems designed for specific soil conditions are typically used for treating human waste in areas not served by municipal treatment facilities.

Most home sewage systems rely on anaerobic decomposition in septic tanks with the resulting effluent being discharged into a leaching field. Some conditions, such as a high water table, require that the septic system be constructed above ground in mounds. Human waste is not to be stored or processed in animal waste management facilities because of the potential for disease transmission.

Landowners should contact local health authorities for design requirements and permit information before installing treatment systems for human waste. SCS does not design human waste management systems, but some States have extension specialists or environmental engineers that can assist in designing suitable systems.

Figure 10-49 Capacity requirements for poultry disposal pits for laying hens and turkeys



651.1008 Safety

Much of this material was taken from the publication, "Safety and Liquid Manure Handling" (White and Young 1980).

Safety must be a primary consideration in managing animal waste. It must be considered during planning and designing of waste management system components as well as during the actual operation of handling wastes. The operator must be made aware of safety aspects of any waste management system components under consideration. Accidents involving waste management may be the result of:

- Poor design or construction
- Lack of knowledge or training about components and their characteristics
- Poor judgement, carelessness, or lack of maintenance
- Lack of adequate safety devices, such as shields, guard rails, fences, or warning signs

The potential for an accident with waste management components is always present. However, accidents do not have to happen if components are properly designed, constructed, and maintained and if all persons involved with the components are adequately trained and supervised.

First aid equipment should be near storage units and lagoons. A special, easily accessible area should be provided for storing the equipment. The area should be inspected periodically to ensure that all equipment is available and in proper working condition. The telephone numbers of the local fire department and/or rescue squad should be posted near the safety equipment and near all telephones.

(a) Confined areas

Manure gases can accumulate when manure is stored in environments that do not have adequate ventilation, such as underground covered waste storage tanks. These gases can reach toxic concentrations, and displace oxygen. The four main gases are ammonia (NH_3) , carbon dioxide (CO_2) , hydrogen sulfide (H_2S) ,

and methane ($\mathrm{CH_4}$). The gases produced under anaerobic conditions and the requirements for safety because of these deadly gases are described in chapter 3. Because of the importance of safety considerations, the following repeats and elaborates on these safety requirements.

Ammonia is an irritant at concentrations below 20 ppm. At higher levels it can be an asphyxiant.

Carbon dioxide is released from liquid or slurry manure. The rate of release is increased with agitation of the manure. High concentrations of carbon dioxide can cause headaches and drowsiness and even death by asphyxiation.

Hydrogen sulfide is the most dangerous of the manure gases and can cause discomfort, headaches, nausea, and dizziness. These symptoms become severe at concentrations of 800 ppm for exposures over 30 minutes. Hydrogen sulfide concentrations above 800 ppm can lead to unconsciousness and death through paralysis of the respiratory system.

Methane is also an asphyxiant; however, it's most dangerous characteristic is that it is explosive.

Several rules should be followed when dealing with manure stored in poorly ventilated environments:

Safety equipment can include air packs and face masks, nylon line with snap buckles, safety harness, first-aid kits, flotation devices, safety signs, and hazardous atmosphere testing kits or monitors. All family members and employees should be trained in first-aid, CPR techniques, and safety procedures and policies. The following material discusses specific safety considerations.

Do not enter a manure pit unless absolutely necessary and then only if (1) the pit is first ventilated, (2) you have air supplied to a mask or a self contained breathing apparatus, and (3) you have on a safety harness and attached rope and have two people standing by.

If at all feasible, construct lids for manure pits or tanks and keep access covers in place. If an open, ground level pit or tank is necessary, put a fence around it and post "Keep Out" signs.

Do not attempt without assistance to rescue humans or livestock that have fallen into a manure storage structure or reception pit.

Move all the animals out of the building if possible when agitating manure stored beneath that building. If the animals cannot be removed, the following steps should be taken:

- If the building is mechanically ventilated, turn fans on full capacity when beginning to agitate, even in the winter.
- If the building is naturally ventilated, do not agitate unless there is a brisk breeze blowing. The animals should be watched when agitation begins, and at the first sign of trouble, the pump should be turned off. The critical area of the building is where the pumped manure breaks the liquid surface in the pit. If an animal drops over because of asphyxiation, do not try to rescue it, or you might also become a victim. Turn off the pump and allow time for the gases to escape before entering the building.

Do not smoke, weld, or use an open flame in confined, poorly ventilated areas where methane can accumulate.

Keep electric motors, fixtures, and wiring near manure storage structures in good condition.

(b) Aboveground tanks

Aboveground tanks can be dangerous if access is not restricted. Uncontrolled access can lead to injury or death from falls from ladders and to death from drowning if someone falls into the storage tank. The following rules should be enforced:

Permanent ladders on the outside of aboveground tanks should have entry guards locked in place or the ladder should be terminated above the reach of individuals.

A ladder must never be left standing against an aboveground tank.

(c) Lagoons, ponds, and liquid storage structures

Lagoons, ponds, and liquid storage structures present the potential for drowning of animals and humans if access is not restricted. Floating crusts can appear capable of supporting a person's weight and provide a false sense of security. Tractors and equipment can fall or slide into storage ponds or lagoons if they are operated too close to them. The following rules should be obeyed:

Rails should be built along all walkways or ramps of open manure storage structures.

Fence around storage ponds and lagoons, and post signs "Caution Manure Storage (or Lagoon)." The fence keeps livestock and children away from the structure. Additional precautions include a minimum of one lifesaving station equipped with a reaching pole and a ring buoy on a line.

Place a barrier strong enough to stop a slowmoving tractor on all push-off platforms or ramps.

If manure storage is outside the livestock building, use a water trap or other device to prevent gases in the storage structure from entering the building, especially during agitation.

(d) Equipment

All equipment associated with waste management, such as spreaders, pumps, conveyors, and tractors, can be dangerous if improperly maintained or operated. Operators should be thoroughly familiar with the operator's manual for each piece of equipment. Equipment should be inspected frequently and serviced as required. All guards and safety shields must be kept in place on pumps, around pump hoppers, and on manure spreaders, tank wagons, and power units.

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651.1050 Appendix 10A—Blank worksheets

Worksheet 10A-1—Waste storage structure capacity design

Decisionmaker:	Date:
Site:	
Animal units	
1. Animal type	3. Number of animals (N)
2. Animal weight, lbs (W)	4. Animal units, $AU = \frac{W \times N}{1000} = \frac{1}{1000} = 1$
Manure volume	
Daily volume of daily manure production per AU, ft ³ /AU/day (DVM)=	VMD = AU x DVM x D =
6. Storage period, days (D) =	8. Total manure production for storage period, ft ³ (TVM)
Wastewater volume	
9. Daily wastewater volume per AU, ft ³ /AU/day (DWW) =	11. Total wastewater volume for storage period, ft ³ (TWW)
10. Total wastewater volume for animal description for storage period, ft ³ WWD = DWW x AU x D =	_
Bedding volume	
12. Amount of bedding used daily for animal type, lbs/AU/day (WB) =	14. Bedding volume for animal type for storage period, ft ³ BV =
13. Bedding unit weight, lbs/fb ³ (BUW) =	0.5 x WB x AU x D
	15. Total bedding volume for storage period, ft ³ (TBV) =
Minimum waste storage volume requirement	
16. Waste storage volume, ft ³ (WV) = TVM + TWW + TBV =	+ + =
Waste stacking structure sizing	
17. Structure length, ft L = WV = — — — — — — — — — — — — — — — — — —	19. Structure height, ft H = WV L x WI
18. Structure width, ft WI =WV =	-
Notes for waste stacking structure:	
1. The volume determined (WV) does not include any volume for freeboard. It is recommended that a minimum of 1 foot of freeboard be provided for a waste stacking structure.	The equations for L, WI, and H assume manure is stacked to average height equal to the sidewall height. Available storage volume must be adjusted to account for these types of variations.
Tank sizing	22. Rectangular tank dimensions
20. Effective depth, ft. (EH)	Total height, ft (H) = —————Selected width, ft (WI) =
Total height (or depth) of tank desired, ft (H)	
Less precipitation for storage period, ft.	Length, ft L = $\frac{SA}{WI}$ = $\frac{SA}{WI}$
(uncovered tanks only)	23. Circular tank dimensions
Less depth allowance for accumulated solids, ft –	Total height, ft H =
(0.5 ft. minimum)	DI
Less depth for freeboard (0.5 ft. recommended), ft –	Diameter, ft DIA = (1.273 x SA)0 ^{.5} =
Effective depth, ft (EH) =	 Final dimensions may be rounded up to whole numbers or to use increments on standard drawings.
EH	• • • • • • • • • • • • • • • • • • • •

Worksheet 10A-2—Waste storage pond design

Site:		
Animal units		
1. Animal type	3. Number of animals (N)	
2. Animal weight, lbs (W)	4. Animal units, $AU = \frac{W \times N}{1000}$	=
Manure volume		
5. Daily volume of manure production per AU, ft ³ /AU/day (DVM)=	7. Total volume of manure produ animal type for storage period VMD = AU x DVM x D	d, ft ³
6. Storage period, days (D) =		storage period, ft ³ (TVM)
Wastewater volume		
9. Daily wastewater volume per AU, ft ³ /AU/day (DWW) =	11. Total wastewater volume for storage period, ft ³ (TWW) -	·····
10. Total wastewater volume for animal description for storage period, ft ³ WWD = DWW x AU x D =		
Clean water volume	Runoff volume	
12. Clean water added during storage period, ft ³ (CW)	Includes the volume of runoff to	
Solids accumulation	due to normal runoff for the strunoff volume from the 25-year	
14. Volume of solids accumulation, ft ³ (VSA)	runon volume from the 25-yea	ii, 24-nour storm.
Minimum waste storage volume requirement 15. Waste storage volume, ft ³ (WSV) = TVM + TWW + CW + ROV + VSA = +		+ =
Pond sizing		
16. Sizing by trial and error		
Side slope ratio, (Z) = V must be equal to or	greater than WSV =	ft ³
Rectangular pond,	Circular pond,	
$V = \left(\frac{4 \times Z^2 \times d^3}{3}\right) + \left(Z \times BL \times d^2\right) + \left(Z \times BW \times d^2\right) + \left(BW \times BL \times d\right)$	•	$7 \times W \times Z \times d^2$) + (0.79 x W ² x d)
Trial Bottom width Bottom length Depth* Volume no. ft (BW) ft (BL) ft (d) ft ³ (V)		eter Depth* Volume ft (d) ft ³ (V)
* Depth must be adjusted in Step 17. Depth adjustment		
17. Depth adjustment		
Depth, ft (d)		
Add depth of precipitation less evaporation +(For the storage period)	,	num) +
Add depth of 25-year, 24-hour storm+		

Worksheet 10A-3—Anaerobic lagoon design

Decisionmaker:	Date:
Site:	
Animal units	
1. Animal type	3. Number of animals (N)
2. Animal weight, lbs (W)	4. Animal units, AU = <u>W x N</u> =
Manure volume	
5. Daily volume of daily manure production per AU, ft ³ /AU/day (DVM)=	7. Total volume of manure production for animal type for treatment period, ft ³ VMD = AU x DVM x D =
6. Treatment period, days (D) =	8. Total manure production for treatment period, ft ³ (TVM)-
Wastewater volume	
9. Daily wastewater volume per AU, ft ³ /AU/day (DWW) =	11. Total wastewater volume for treatment period, ft ³ (TWW)
10. Total wastewater volume for animal description for treatment period, ft ³ WWD = DWW x AU x D =	
Clean water volume	
12. Clean water added during treatment period, ft ³ (CW)	
Waste volume 13. Waste volume for treatment period, ft ³ WV = TVM + TWW + CW =	+=
Manure total solids	
 14. Daily manure total solids production, lbs/AU/day (MTS) = 15. Daily manure total solids production for animal type, lbs/day 	16. Total manure total solids production,
MTCD MTC ALL	lbs/day (TMTS) =
Manure volatile solids 17. Daily manure volatile solids production per AU, lbs/AU/day (MVS) =	
18. Daily manure volatile solids production for animal type per day, lbs/day	
Wastewater volatile solids	
20. Daily wastewater volatile solids production, lbs/1000 gal (DWVS)	= <u> </u>
21. Total wastewater volatile solids production for animal type, lbs/day	
WVSD = DWVS x DWW x 7.48	=
D x 1,000	
22. Total wastewater volatile solids production, lbs/day (TWVS)	= =
Total volatile solids (manure and wastewater) 23. Total daily volatile solids production, lbs/day TVS = TMVS + TWVS =	+=
Minimum treatment volume	25. Minimum treatment volume, ft ³
24. Selected lagoon VS loading rate, lbs VS/1,000 ft ³ (VSLR) =	*
	VSLR ()
Sludge volume requirement	28. Sludge volume requirement, ft ³
26. Sludge accumulation ratio, ft ³ /lb TS (SAR) =	SV = 365 x TMTS x T x SAR
27 Sludge accumulation period, years (T) =	= 365 x ()()()=
Minimum lagoon volume requirement 29. Minimum lagoon volume requirements, ft ³ (MLVR) = MTV + SV + WV =	+ =

Worksheet 10A-3—Anaerobic lagoon design—Continued

goon sizing 10. Sizing by trial and error $V = \frac{(4 \times Z)^2}{2}$	$(Z \times d^3) + (Z \times BL \times d^2) + (Z \times BW \times d^2)$	$(d^2) + (BW \times BL \times d)$	
Side slope ratio, (Z) =	3	al to or greater than MLVR =	ft³
Trial Bottom width no. ft (BW)	Bottom length ft (BL)	Depth* ft (d)	Volume ft ³ (V)
epth adjustment			
31. Depth adjustment			
Depth, ft (d)			
Add depth of precipitation less evaporation on lago (for the treatment period)	oon surface+		
Add depth of 25-year, 24-hour storm	+ <u></u>		
Add for freeboard (1.0 foot minimum)	+		
Final depth			
2. Compute total volume using final depth, ft ³ (use ec	quation in step 30)		

Worksheet 10A-4—Aerobic lagoon design

Decicionmolere	Date:
Decisionmaker:	Date.
Site:	
Animal units	
1. Animal type	3. Number of animals (N)
2. Animal weight, lbs (W)	4. Animal units, $AU = \frac{W \times N}{1000} = \dots$
Manure volume	
5. Daily volume of daily manure production per AU, ft ³ /AU/day (DVM) =	VMD = AU x DVM x D =
6. Treatment period, days (D) =	8. Total manure production for treatment period, ft ³ (TVM)
Wastewater volume	
9. Daily wastewater volume per AU, ft ³ /AU/day (DWW) =	11. Total wastewater volume for treatment period, ft ³ (TWW)
10. Total wastewater volume for animal description for treatment period, ft ³ WWD = DWW x AU x D =	
Clean water volume	
12. Clean water added during treatment period, ft ³ (CW)	
Waste volume	
13. Waste volume for treatment period, ft ³ WV = TVM + TWW + 0	CW = + =
Manure total solids 14 Daily manure total solids production lbs/ALVday (MTS)	16. Total manure total solids production,
14. Daily manure total solids production, lbs/AU/day (MTS) = 15. Daily manure total solids production for animal type, lb/day	lbs/day (TMTS) =
MTSD = MTS x AU =	
Manure 5-day biochemical oxygen demand 17. Daily manure BOD ₅ production per AU, lbs/AU/day (MBOD) =	
18. Daily manure BOD ₅ production for animal type per day, lbs/day	MBOD = AU x BOD =
19. Total manure production, lbs/day (TMBOD)	
Wastewater 5-day biochemical oxygen demand	d
20. Daily wastewater BOD ₅ production, lbs/1000 gal (DWBOD)	=
21. Total wastewater BOD_5 production for animal type, lbs/day	
$WBOD = \frac{(DWBOD \times TWW \times 7.48)}{D \times 1,000}$	= ———————————————————————————————
22. Total wastewater BOD ₅ production, lbs/day (TWBOD)	= = <u></u>
TOTAL BOD ₅ (manure and wastewater) 23. Total daily production, lbs/day TBOD = TMBOD + TWBOD =	
Minimum treatment surface area	25. Minimum treatment surface area, acres
24. Selected lagoon BOD ₅ loading rate, lbs BOD ₅ /acre (BODLR) =	MTA = TBOD
Sludge volume requirement	28. Sludge volume requirement, ft ³
26. Sludge accumulation ratio, ft³/lb TS (SAR) =	SV = 365 x TMTS x T x SAR
27 Sludge accumulation period, years (T) =	= 365 ()()()=
Minimum lagoon volume requirement 29. Minimum lagoon volume requirements, ft ³ MLVR = SV + WV =	

Worksheet 10A-4—Aerobic Lagoon Design—Continued

	torosio Lagoon Doc	3				
Lagoon sizing 30. Sizing by trial and Side slope ratio, (a	error: Z) =					
V must be equal to	o or greater than MLVR = _	ft ³				
SA must be equal	to or greater than MTA = _	acres				
Rectangular lagoo	n:					
d must be less tha	n 5 feet					
$SA = \frac{(BL + 2Zd)}{43}$)(BW + 2Zd) ,560					
$V = \left(\frac{4 \times Z^2 \times d^3}{3}\right)$	$\left(Z \times BL \times d^2\right) + \left(Z \times BL \times d^2\right)$	$W \times d^2$) + (BW × BL × d)				
Trial no.	Bottom width ft (BW)	Bottom length ft (BL)	Depth* ft (d)	Volume ft ³ (V)	Surface area acres (SA)	
* Depth must be ac	djusted in Step 31					
Depth adjustme	nt					
31. Depth adjustment						
Depth , ft (d)						
Add depth of precip (for the treatme	itation less evaporation on nt period)	lagoon surface +				
Add depth of 25-year	ar, 24-hour storm	+				
Add for freeboard ((1.0 foot minimum)	+				
Final depth						
32. Compute total volur (use equation in ste	me using final depth, ft ³ p 30)					

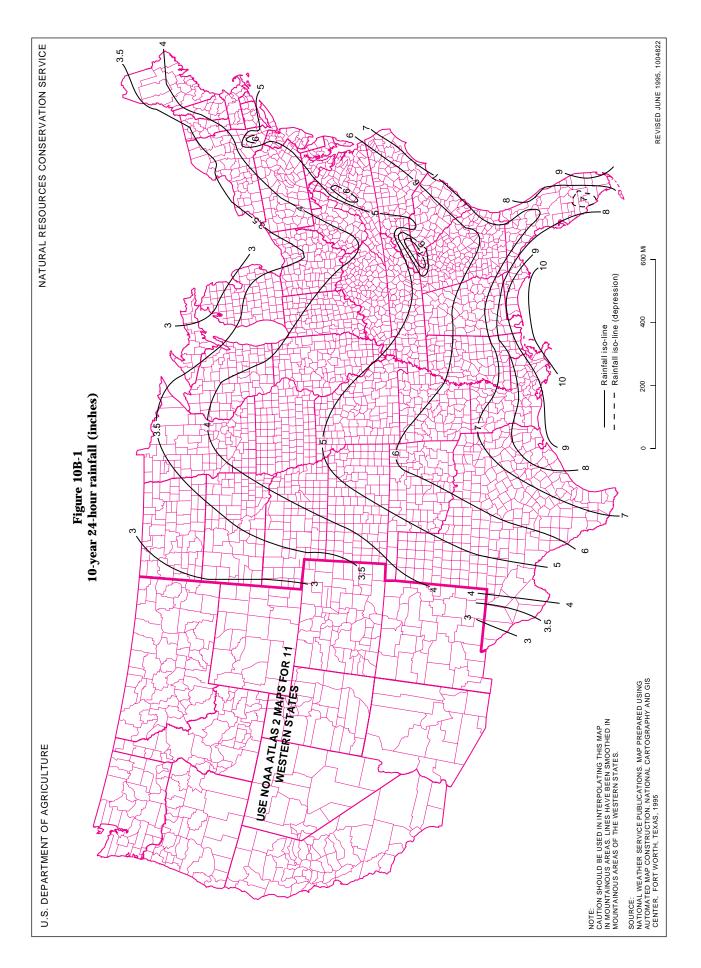
Site:	
Animal units	
1. Animal type	3. Number of animals (N)
2. Animal weight, lbs (W)	4. Animal units,, $AU = \frac{W \times N}{1000} = \frac{W \times N}{1000}$
Manure volume 5. Daily volume of daily manure production per AU, ft ³ /AU/day (DVM)=	7. Total daily manure production volume, ft ³ /day (TMP)
6. Total volume of daily manure production for animal type, ft ³ /day MPD = AU x DVM	
Manure total solids 8. Daily manure total solids production, lbs/AU/day (MTS) =	
Daily manure total solids production for animal type. lb/day	solids production, lbs/day (TMTS) =
	Director food called concentration
Percent solids 14. Percent solids, % (PS) PS =TMTS x 100 (Digester feed solid concentration 15. Desired digester feed solids concentration, % (DDFSC) =
	Digester effective volume 17. Digester effective volume, ft ³
14. Percent solids, % (PS) $PS = \frac{TMTS \times 100}{TMP \times 62.4} = \frac{() \times 100}{() \times 62.4} = {}$ Daily manure inflow 16. Daily manure inflow, ft ³	Digester effective volume 17. Digester effective volume, ft ³

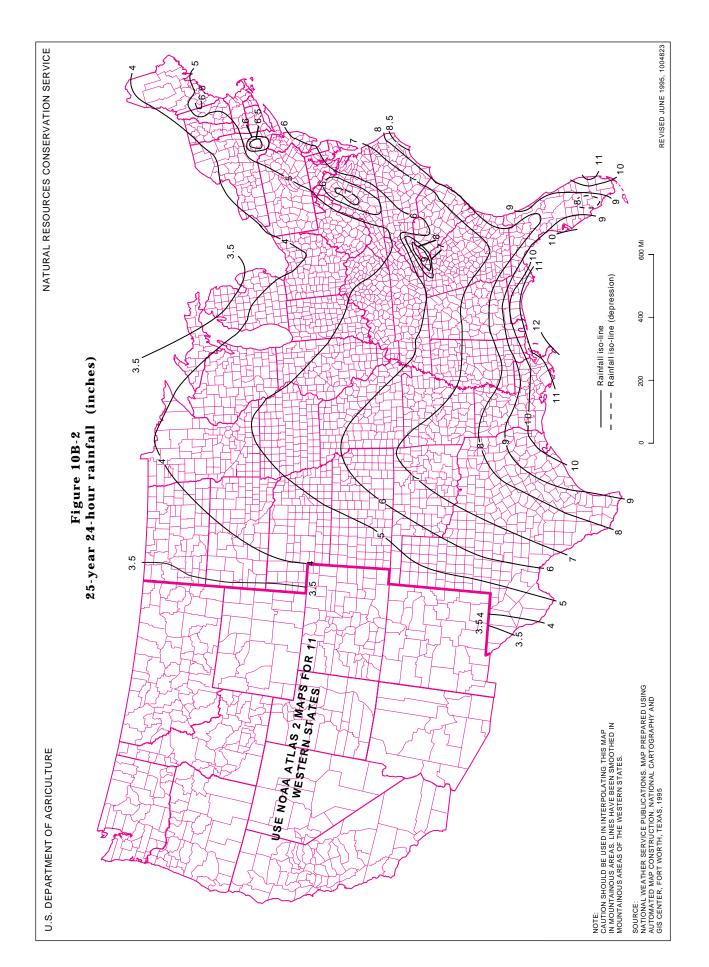
Worksheet 10A-6—Monthly precipitation minus evaporation

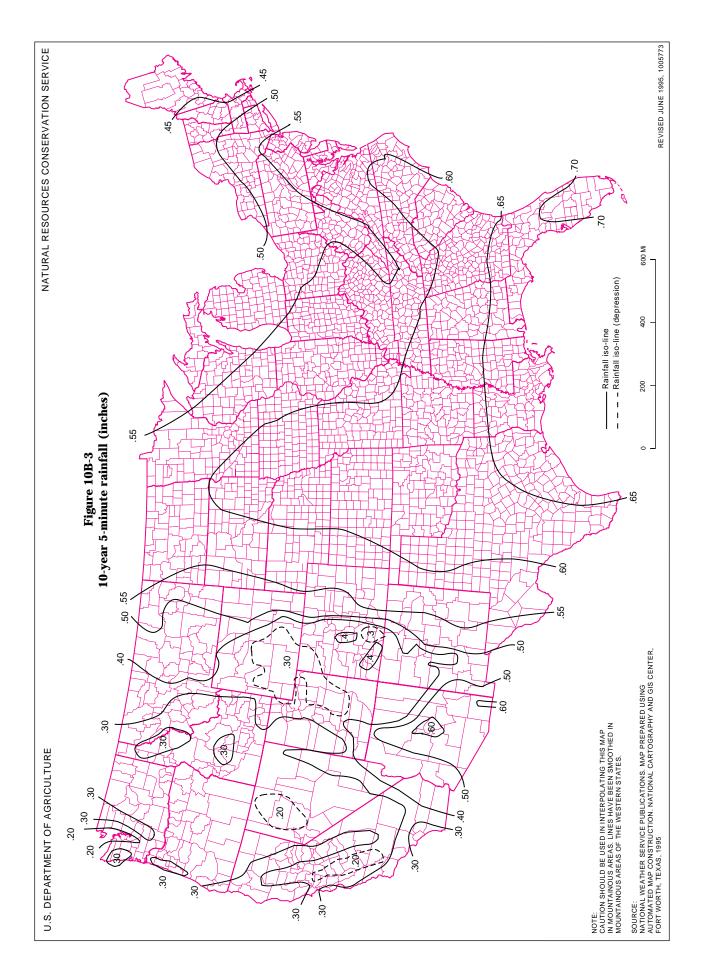
Site: Month Monthly	Decisionmaker:			Date:	
Month precipitation annual evaporation of annual evaporation MP (inches)* Me (inche	Site:			I	
January February March April May June July August September October November December **ME = FWS x MPAE Storage or treatment period, days (D) = months = Critical successive months Monthly precipitation less evaporation Month Mont	Annual FWS Evaporat	ion (<i>FWS</i>) =inc	ches		
February March April May June July August September October November December **ME = FWS x MPAE Storage or treatment period, days (D) = months = Critical successive months Monthly precipitation less evaporation MPLE (inches) Month less evaporation MPLE (inches)	Month	Monthly precipitation MP (inches)	Monthly portion of annual evaporation MPAE (percent)	Monthly evaporation ME (inches)*	Monthly precipitation less evaporation MPLE (inches)
March April May June July August September October November December **ME = FWS x MPAE Storage or treatment period, days (D) = months = **Critical successive months Monthly precipitation less evaporation MPLE (inches) Month Month MPLE (inches)	January				
April May June July August September October November December **ME = FWS x MPAE Storage or treatment period, days (D) = months = Critical successive months Monthly precipitation less evaporation Month MPLE (inches) Month MPLE (inches)	February				
May June July August September October November December *ME = FWS x MPAE Storage or treatment period, days (D) = months = Critical successive months Monthly precipitation less evaporation MPLE (inches) Month MPLE (inches) Month MPLE (inches)	March				
July August September October November December **ME = FWS x MPAE Storage or treatment period, days (D) = months = **Critical successive months Monthly precipitation less evaporation MPLE (inches) Month Month MPLE (inches) Month MPLE (inches)	April				
July August September October November December **ME = FWS x MPAE Storage or treatment period, days (D) = months = Critical successive months Monthly precipitation less evaporation Month MPLE (inches) Month MPLE (inches) Month MPLE (inches)	May				
August September October November December **ME = FWS x MPAE Storage or treatment period, days (D) = months = **Critical successive months Monthly precipitation less evaporation MPLE (inches) Month MPLE (inches) Month MPLE (inches)	June				
September October November December *ME = FWS x MPAE Storage or treatment period, days (D) = months = Critical successive months Monthly precipitation Monthly precipitation less evaporation MPLE (inches) MPLE (inches)	July				-
October November December *ME = FWS x MPAE Storage or treatment period, days (D) = months = Critical successive months Monthly precipitation Month less evaporation MPLE (inches) Month MPLE (inches) Month MPLE (inches)	August				
November	September .				
December *ME = FWS x MPAE Storage or treatment period, days (D) = months = Critical successive months Monthly precipitation Monthly precipitation less evaporation Monthly precipitation Mess evaporation Mess evapo	October				
*ME = FWS x MPAE Storage or treatment period, days (D) = months = **ME = FWS x MPAE Storage or treatment period, days (D) = months = **ME = FWS x MPAE Storage or treatment period, days (D) = months = Monthy = Monthly precipitation Monthly precipitation less evaporation MPLE (inches) MPLE (inches) **ME = FWS x MPAE **MOnthly precipitation Monthly precipita	November				
Storage or treatment period, days (D) = months = Critical successive months Monthly precipitation	December				
months = Critical successive months Monthly precipitation	*ME = FWS	x MPAE			
Critical successive months Monthly precipitation Month Mont	Storage or treatmer	nt period days (D) -			
		Monthly precipitation less evaporation		Month	less evaporation
Total		MPLE (inches)			MPLE (inches)
Total					
	Total			–	

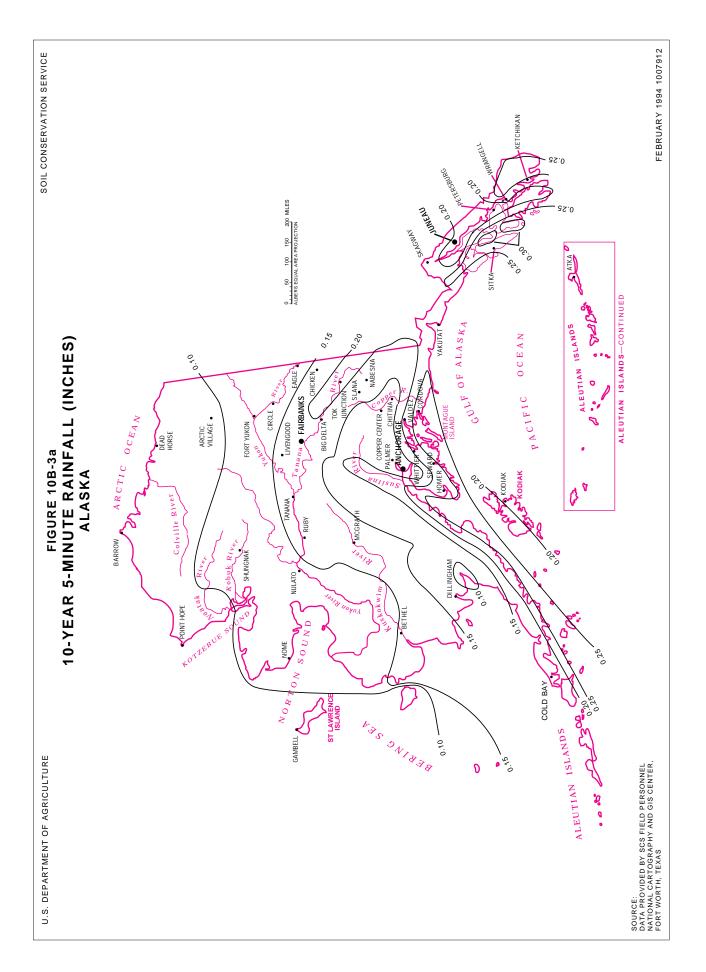
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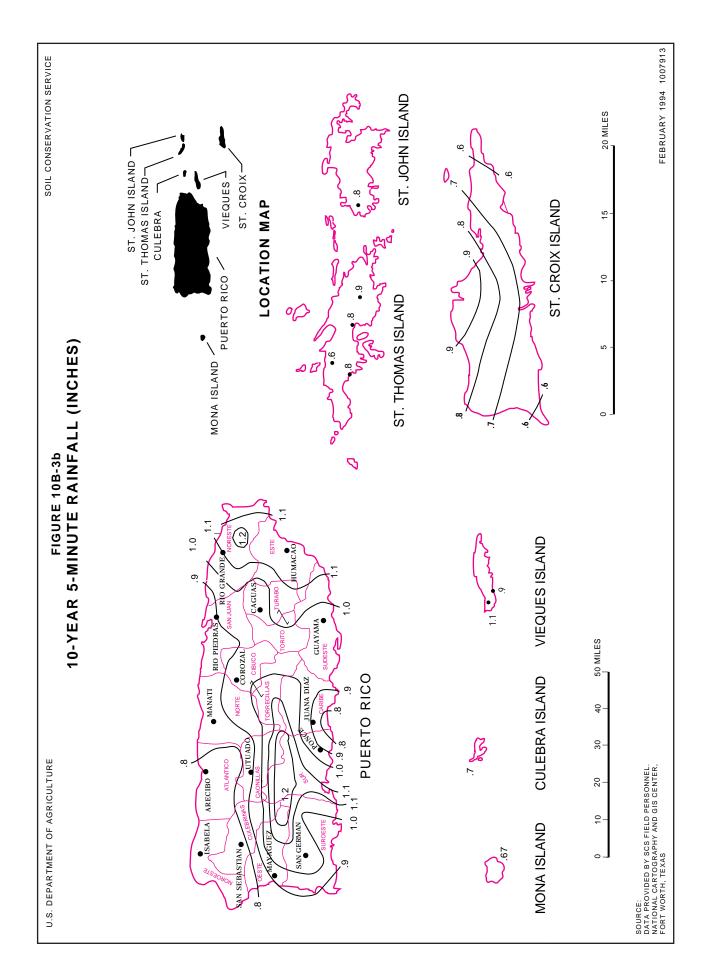
651.1060 Appendix 10B—Rainfall Intensity Maps

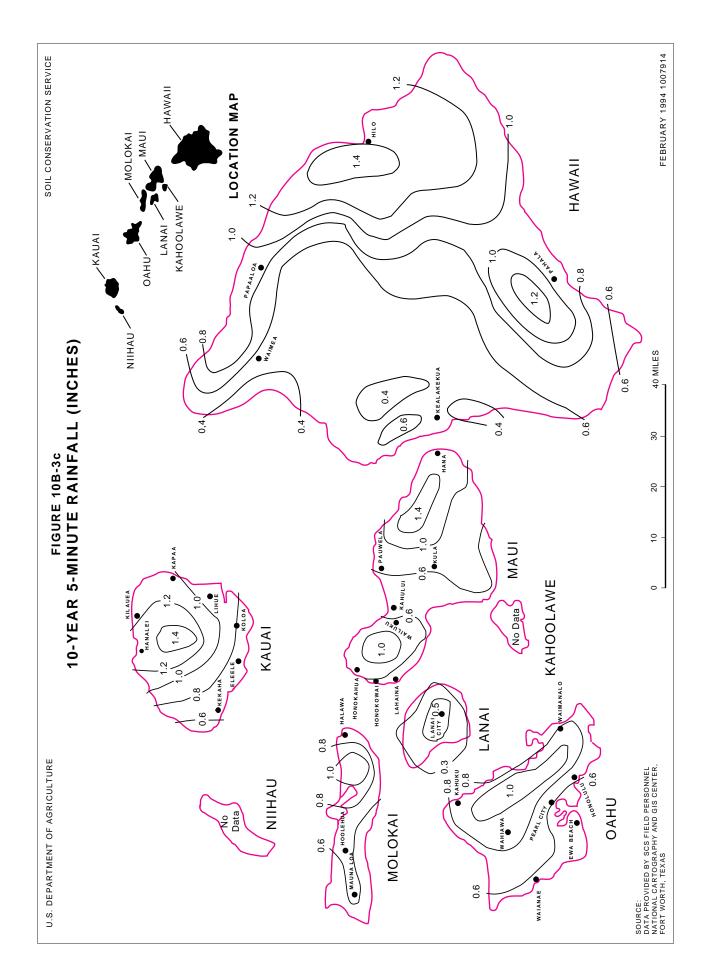


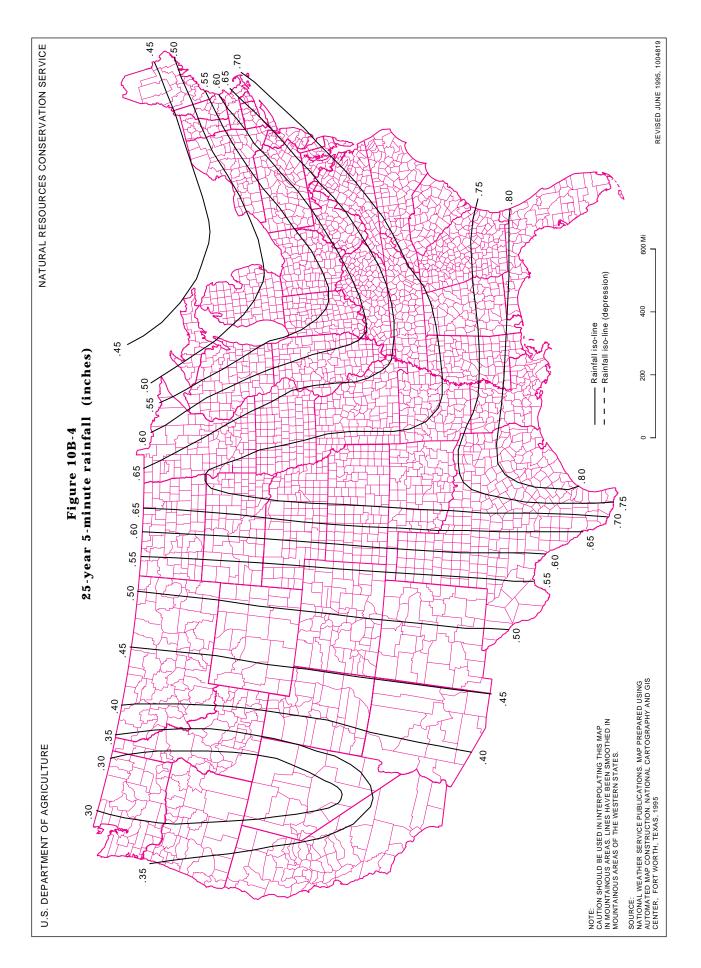












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651.1070 Appendix 10C—Runoff From Feedlots and Evaporation

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(a) Runoff

Runoff must be handled if feedlots or other components of the livestock production unit are exposed to the weather. Contaminated runoff should be collected in settling basins and storage ponds.

A paved or surfaced feedlot typically has a runoff curve number (RCN) of about 97; an RCN of 90 is representative of an unpaved or unsurfaced feedlot. Based on these RCN's, the amount of runoff from feedlots can be estimated as a percentage of the precipitation that is expected over a period of time.

Figures 10C–1 and 10C–2 describe for the continental United States the percentage of annual precipitation that will occur as runoff from unsurfaced and surfaced feedlots, respectively. Figures 10C–3 through 10C–14 describe the percentage of monthly precipitation that will occur as runoff from unsurfaced feedlots. Figures 10C–15 through 10C–26 describe the percentage of monthly precipitation that will occur as runoff from surfaced feedlots.

Other available sources give the annual or monthly precipitation data to which the runoff percentages are applied. One such source is "Climatography of the United States No. 81 (by state) Monthly Normals of Temperature, Precipitation, and Heating and Cooling Degree Days, 1941–70," prepared by the U.S. Department of Commerce, National Oceanic and Atmospheric Administration, Environmental Data Service. Another source available in many counties is the local soil survey, which contains a section on climatic data.

The runoff percentage from figures 10C–1 through 10C–26 is multiplied by the precipitation from the corresponding time period to determine the amount of runoff. This is the runoff volume (ROV) value used in several of the worksheets in chapter 10.

Design example 10C-1—Runoff from a concrete feedlot

Determine the annual runoff from a concrete feedlot near Portland, Oregon. From the reference cited, the mean annual precipitation is 37.6 inches. From figure 10C–2, the annual runoff is 49 percent of the precipitation. Therefore, the annual ROV = $(37.6 \text{ in. } \times 0.49)$ = 18.4 inches.

Design example 10C-2—Runoff from an earth feedlot

Determine the runoff to be expected from an earth feedlot near Dallas, Texas, for the period October to March.

Month	Precip.	Runoff	
	(inches)	%	(inches)
Oct.	3.18	36	1.14
Nov.	2.60	27	0.70
Dec.	2.34	24	0.56
Jan.	1.96	20	0.39
Feb.	2.57	20	0.51
Mar.	3.04	22	0.67
Total			3.97

(b) Evaporation

Storage and treatment facilities require an allowance for precipitation less evaporation for the most critical design period. For example, for a 90-day storage period, an allowance for storage is planned using the three successive months that result in the greatest sum of precipitation less evaporation that is critical.

Some ponds or structures, especially those containing dairy manure and straw bedding, develop a crust on the surface, and evaporation may be limited. This will vary among areas and individual farms. For a conservative design when crusting is anticipated, the allowance evaporation in the pond sizing can be omitted.

Local records are almost always available for the average monthly precipitation for each month of the year. Local records may also be available for average monthly evaporation. If evaporation data are not readily available, however, the annual free water surface evaporation (shallow lake evaporation) may be determined using figure 10C–27. Monthly free water surface evaporation may be determined using table 10C–1, which gives the approximate mean monthly percent of the annual evaporation for selected stations in the continental United States.

Table 10C–1 was developed for use in obtaining monthly evaporation for selected stations from annual Class A pan evaporation maps. This table is to be used

on free water surface maps. Although the information in this table is not completely correct, the monthly percentages are adequate for estimating free water surface evaporation. Several other factors prevent an exact correlation between evaporation from waste storage ponds and lagoon surfaces and Class A pan evaporation. Factors causing differences include effects of salinity, coloration, and floating surface material, such as bedding, on evaporation rates.

Worksheet 10A–6 can be used to determine the monthly precipitation less evaporation value for each month.

Design example 10C-3

Mr. Austin Peabody of Rocky Mount, North Carolina, has selected an alternative for an agricultural waste management system that includes a waste storage pond. Designing the depth of the pond requires that an allowance for containing the precipitation evaporation minus evaporation for the storage period be determined. Using worksheet 10A–6, determine the precipitation less evaporation value to use for a 180-day storage period.

- The annual FWS evaporation (FWS) is selected from figure 10C–27.
- The monthly precipitation (MP) values are selected from local data.
- The monthly portion of annual evaporation (MPAE) is determined using the appropriate station in table 10C–1.
- The monthly evaporation (ME) is computed by the equation:

$$ME = FWS \times MPAE$$

• The monthly precipitation less evaporation (MPLE) is determined by the equation:

$$MPLE = MP - ME$$

• The 180-day storage period is about 6 months; therefore, the successive 6 months that are critical are determined by inspection. For this example, the storage period is September through February.

 The total precipitation less evaporation depth that must be accommodated in the waste storage pond is the sum of monthly values for September through February.

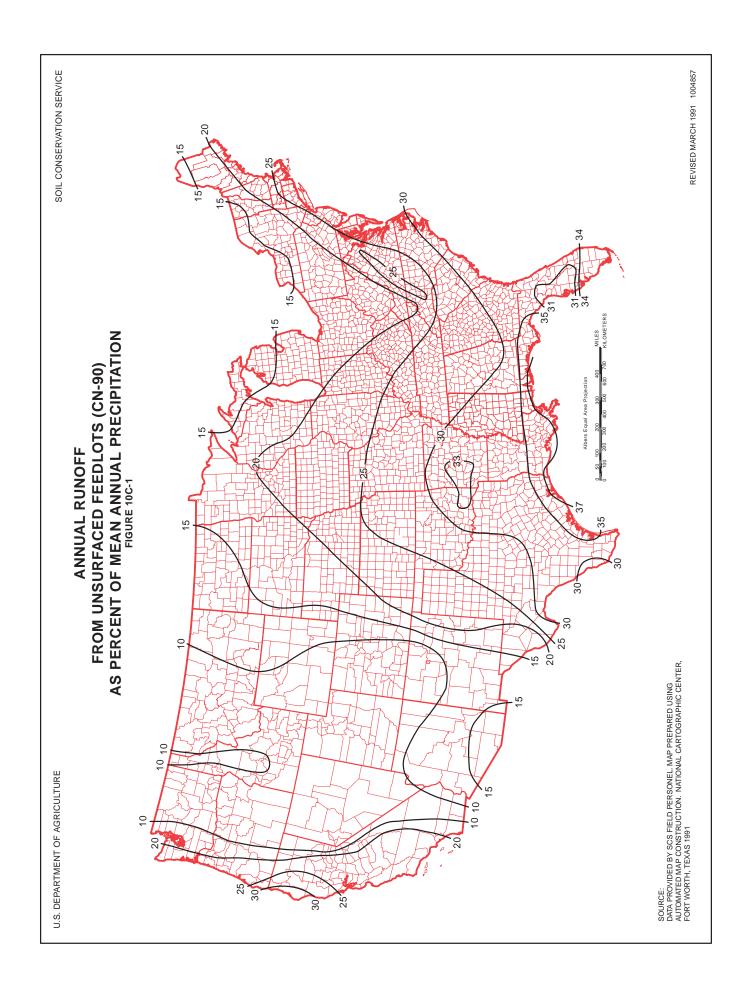
Completed worksheet for design example 10C-3

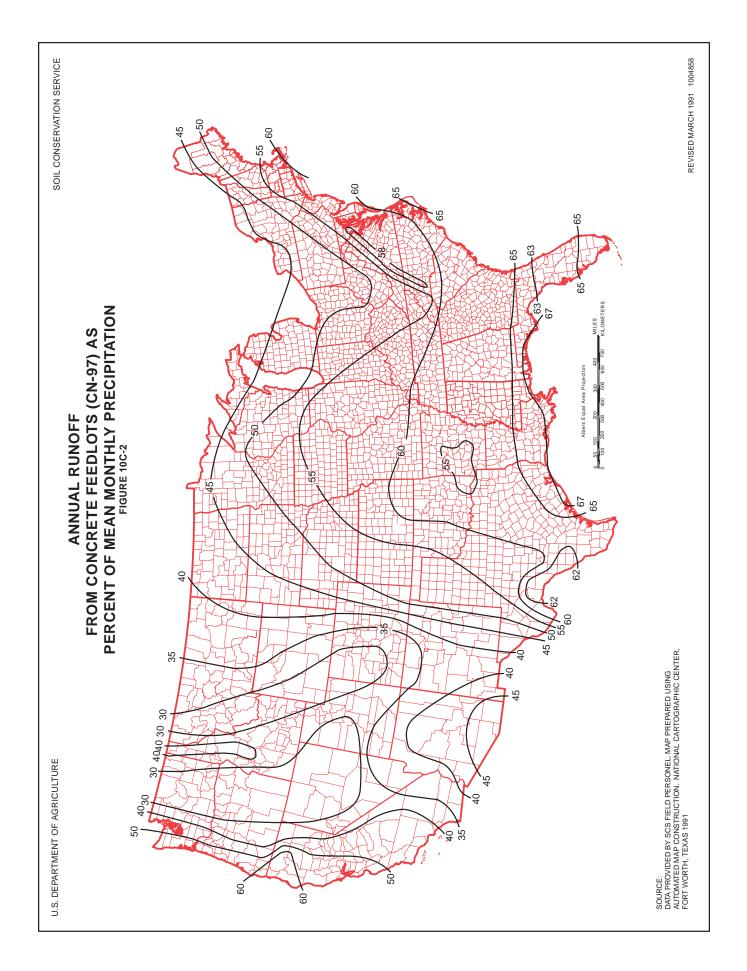
Worksheet 10A-6 – Monthly precipitation in Decisionmaker: Austin Peabody			Date:	Date:	
ite:					
nnual FWS Evaporati	ion (FWS)= 39	inches			
Month	Monthly precipitation MP (inches)	Monthly portion of annual evaporation MPAE (percent)	Monthly evaporation ME (inches)*	Monthly precipitation less evaporation MPLE (inches)	
January	3.53	3	1.17	2.36	
February	3.71	5	1.95	1.76	
March	3.49	8	3.12	0.37	
April	3.50	10	3.90	-0.40	
May	3.61	12	4.68	-1.07	
June	4.47	13	5.07	-0.60	
July	5.58	13	5.07	0.51	
August	4.45	12	4.68	-0.23	
September	3.95	9	3.15	0.44	
October	2.79	7	2.73	0.06	
November	2.24	5	1.95	0.29	
December	3.49	3	1.17	2.32	
*ME = FWS Storage or treatme	ent period, days (D) = 18C				
Storage or treatme	ent period, days (D) = 180 months = 6				
Storage or treatme	ent period, days (D) = $\frac{180}{}$ months = $\frac{6}{}$		Month	Monthly precipitation less evaporation MPLE (inches)	
Storage or treatme	months = 6 months		Month	less evaporation	
Storage or treatments Critical success Jonth SEPT Oct	months = 6 months = 6 months = 6 Monthly precipitation less evaporation MPLE (inches) 0.44 0.06		Month	less evaporation	
Storage or treatments Critical success fronth SEPT Oct NOV	months = 6 months = 6 months = 6 months = 6 Monthly precipitation less evaporation MPLE (inches) 0.44 0.06 0.29		Month	less evaporation	
Storage or treatments Critical success Jonth SEPT Oct	months = 6 months = 6 months = 6 Monthly precipitation less evaporation MPLE (inches) 0.44 0.06		Month	less evaporation	
Storage or treatments Critical success fronth SEPT Oct NOV	months = 6 months = 6 months = 6 months = 6 Monthly precipitation less evaporation MPLE (inches) 0.44 0.06 0.29		Month	less evaporation	
Storage or treatments Critical success SEPT Oct NOV DEC	months = 6 Monthly precipitation less evaporation MPLE (inches) 0.44 0.06 0.29 2.32		Month	less evaporation	

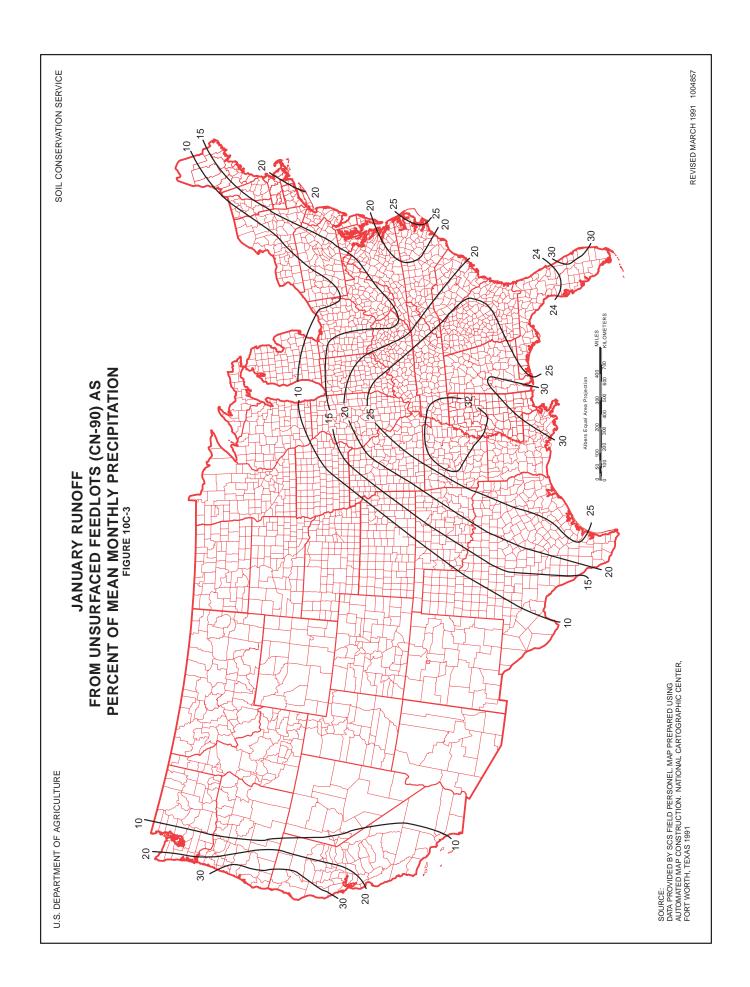
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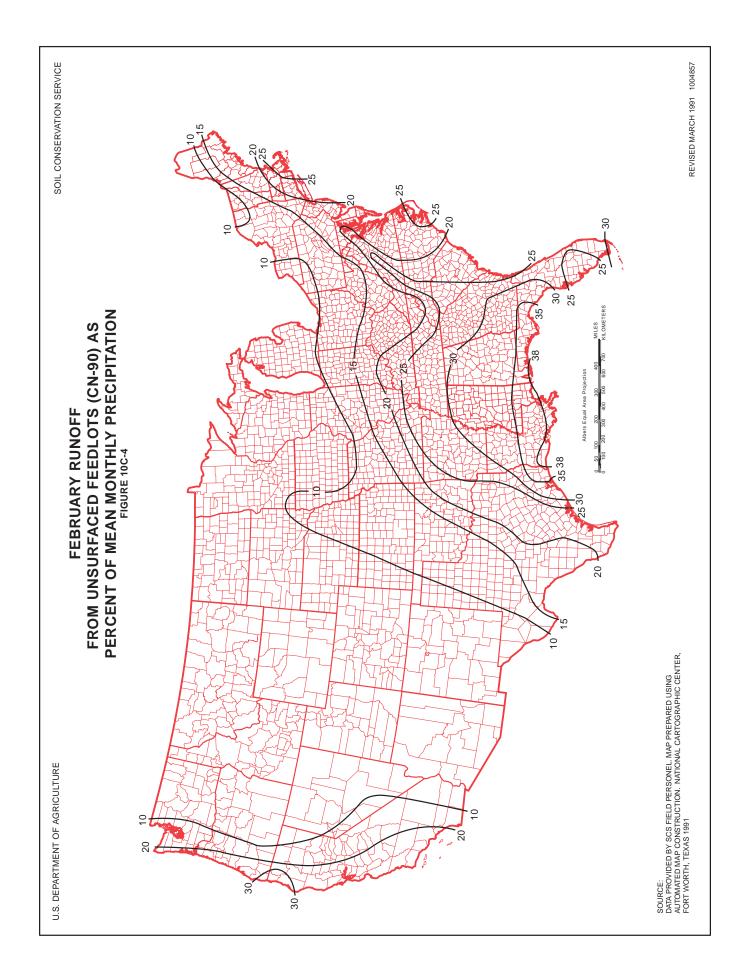
Table 10C-1 Adjusted approximate mean monthly free water surface evaporation for selected stations

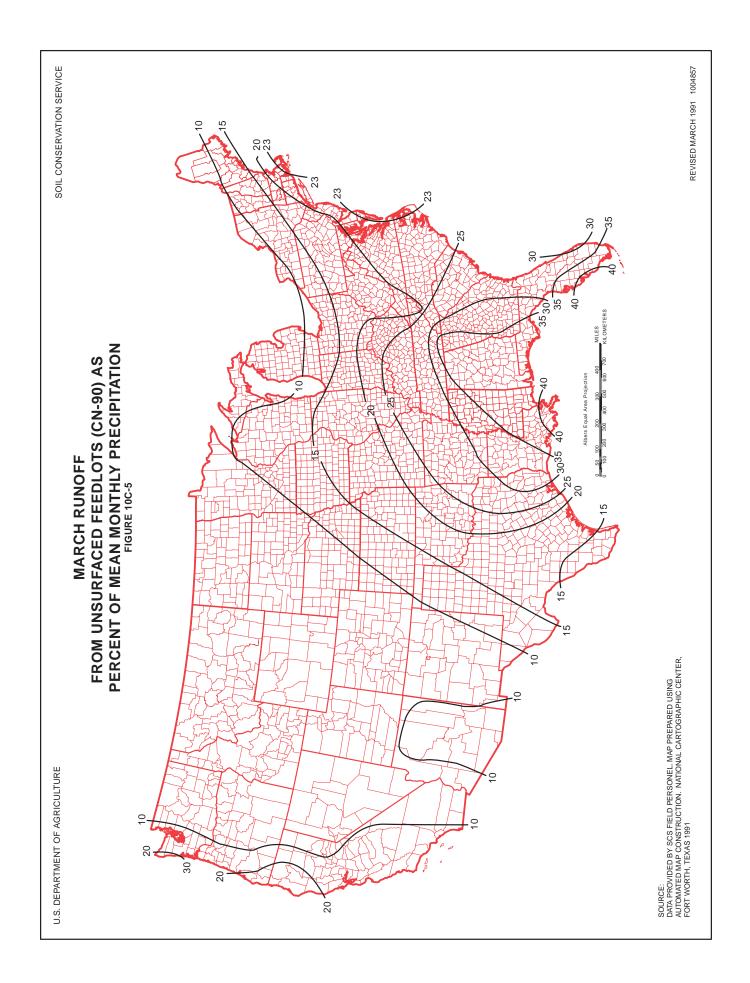
Source: Adapted from Evaporation Atlas for the Contiguous 48 United States, NOAA Technical Report NWS 33, Table 3-Adjusted mean monthly Class A pan evaporation for selected stations, 1956-70.

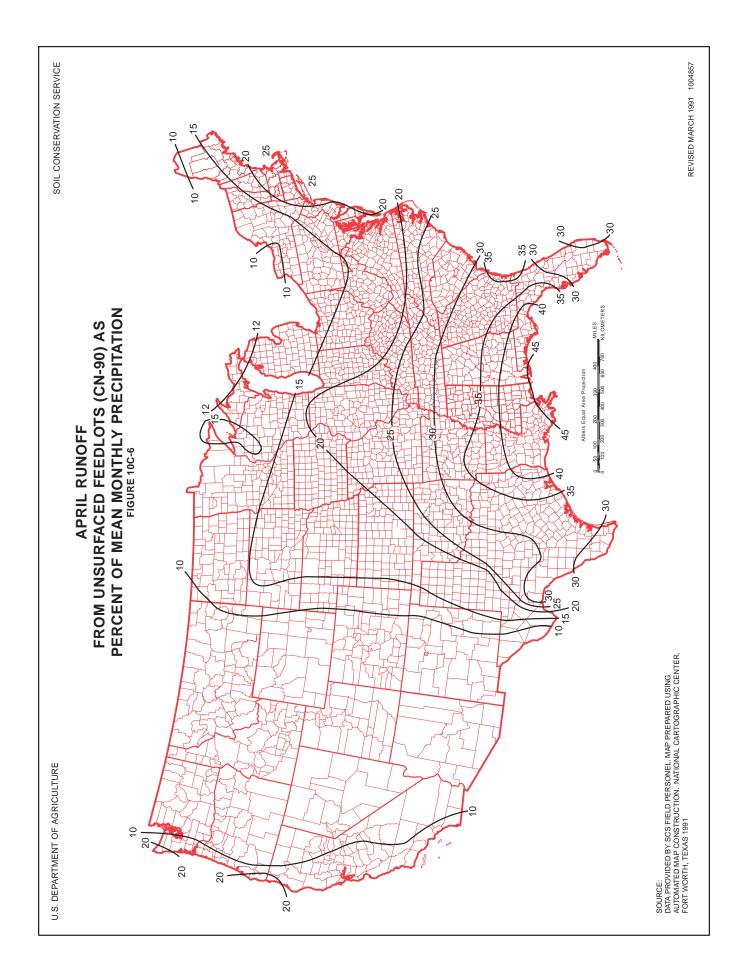


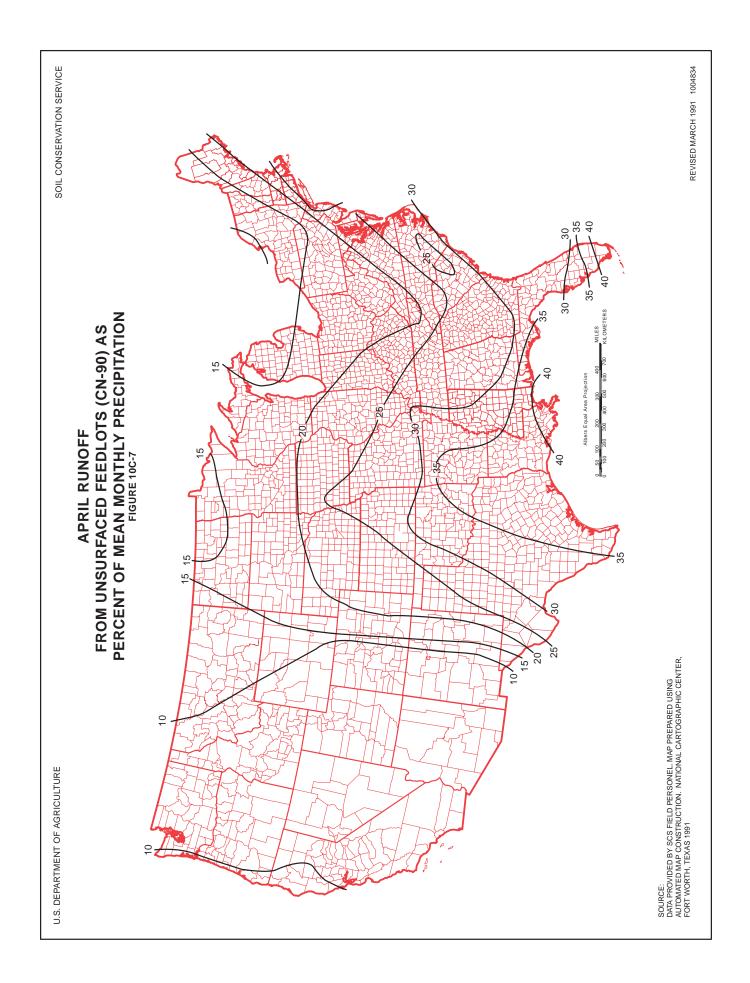


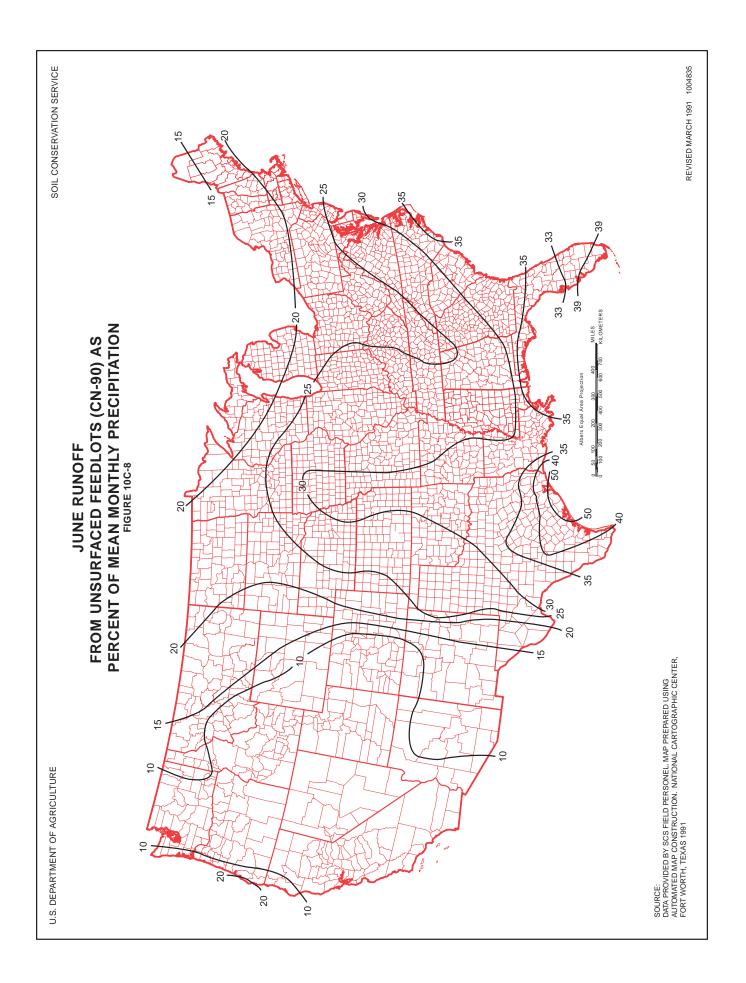


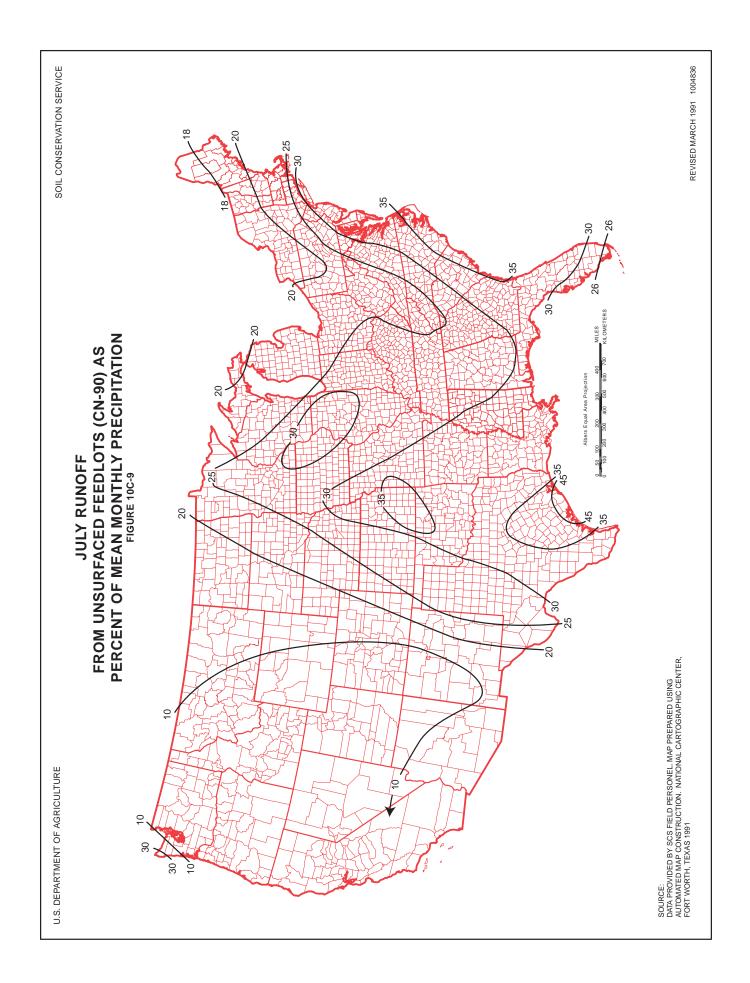


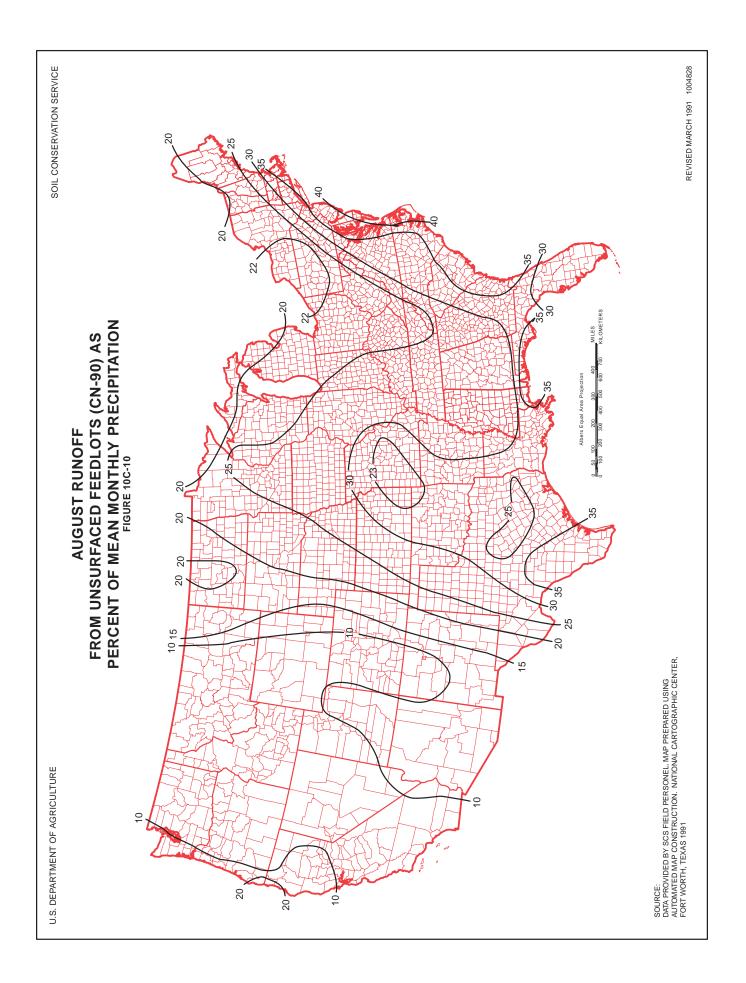


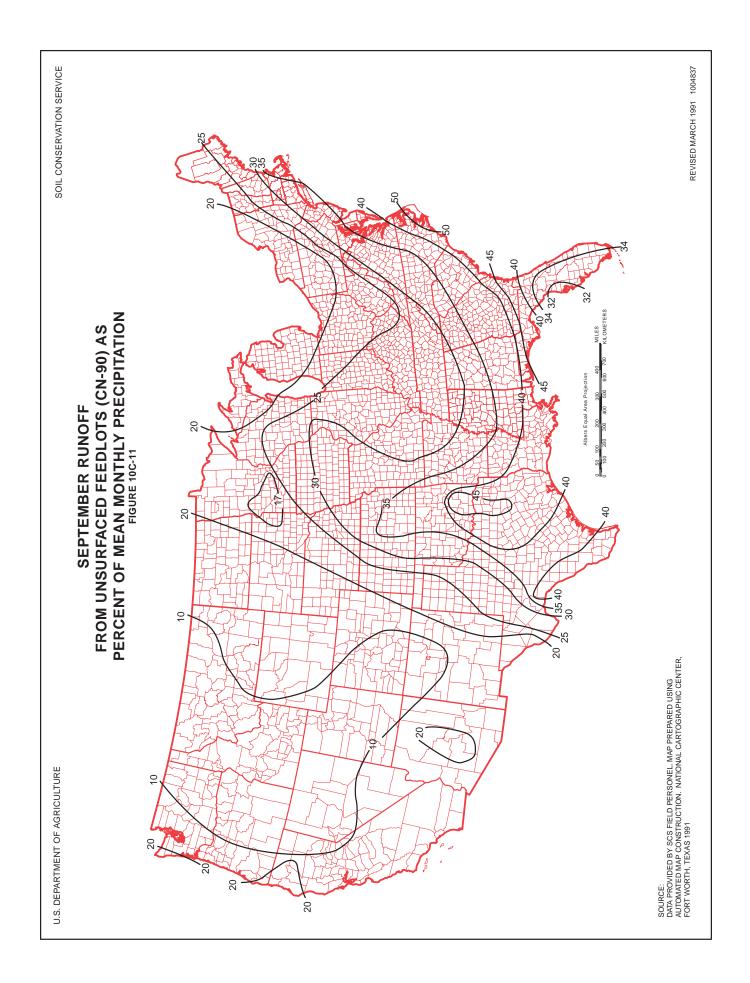


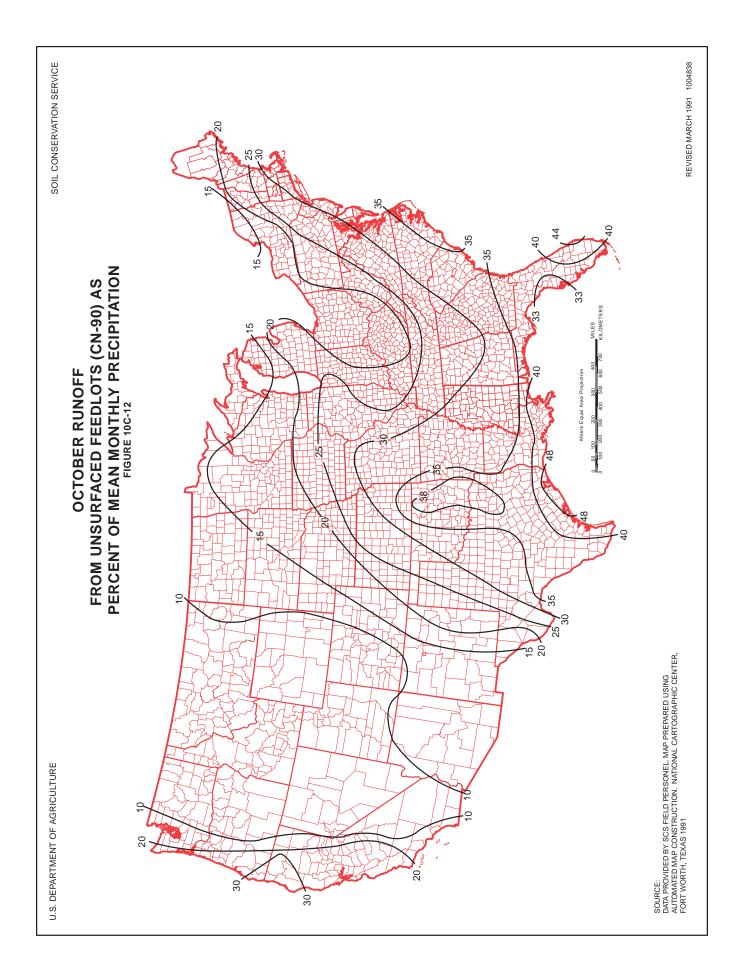


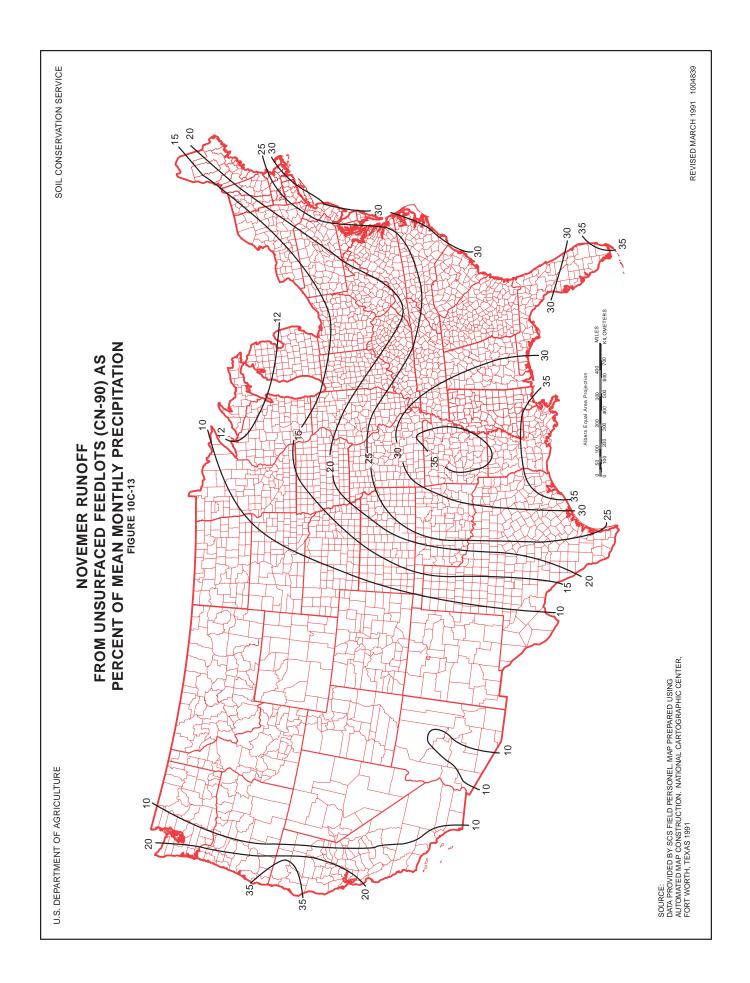


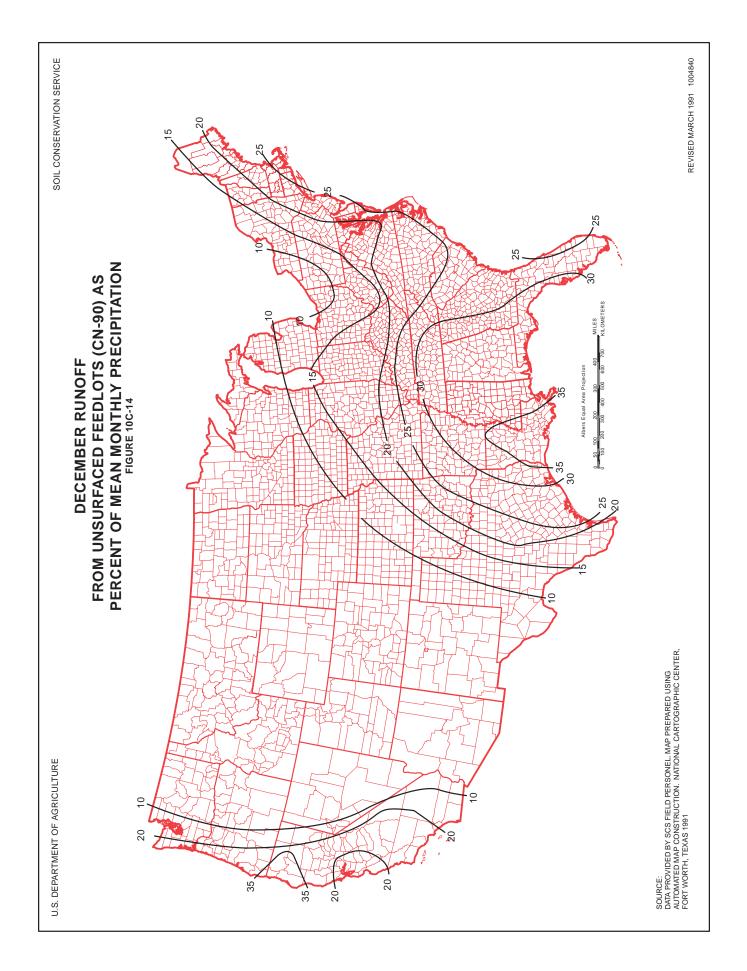


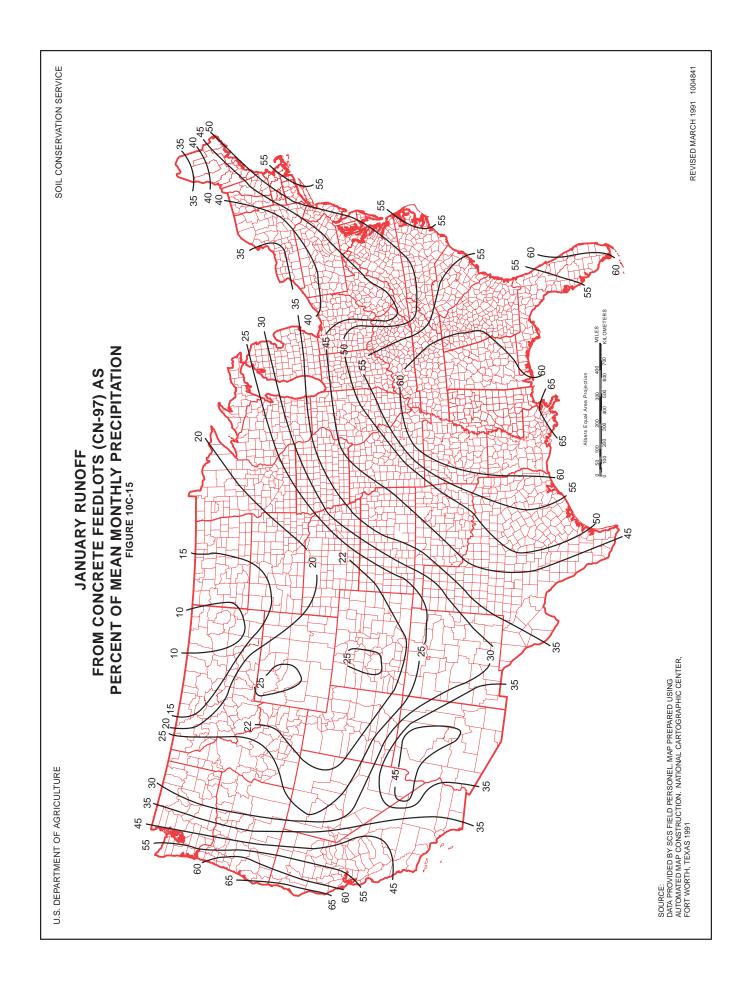


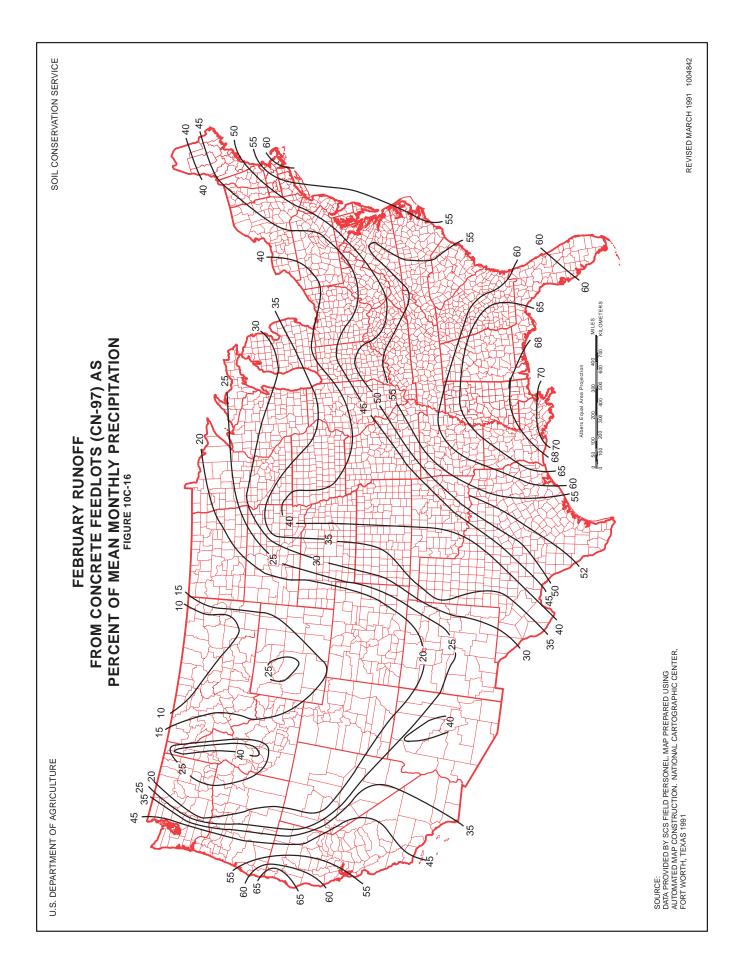


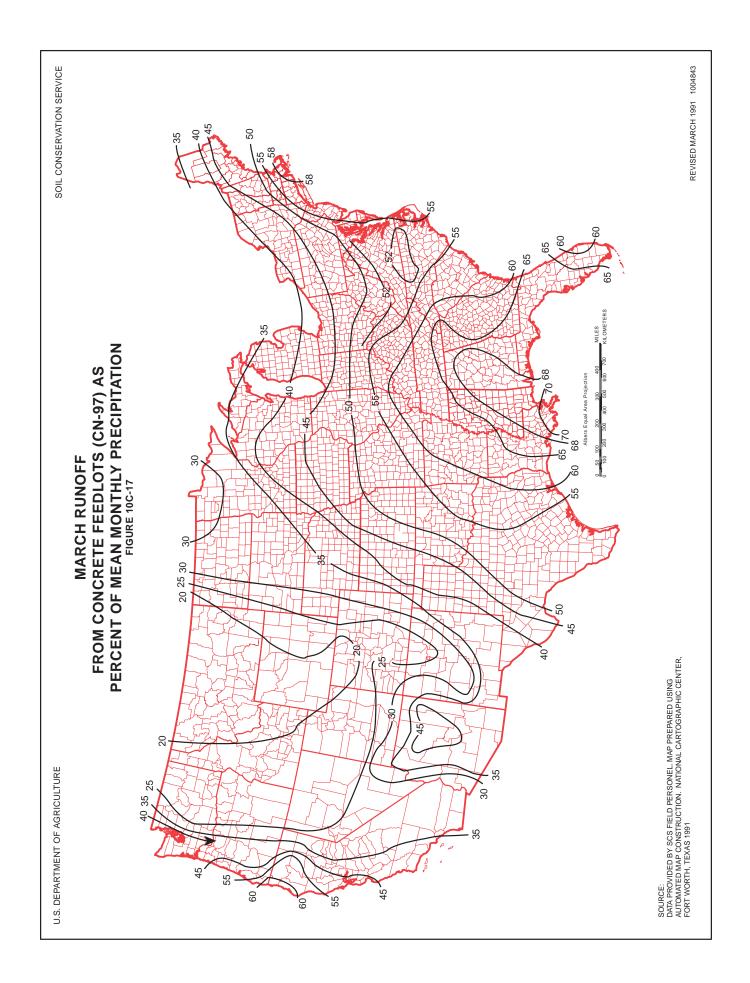


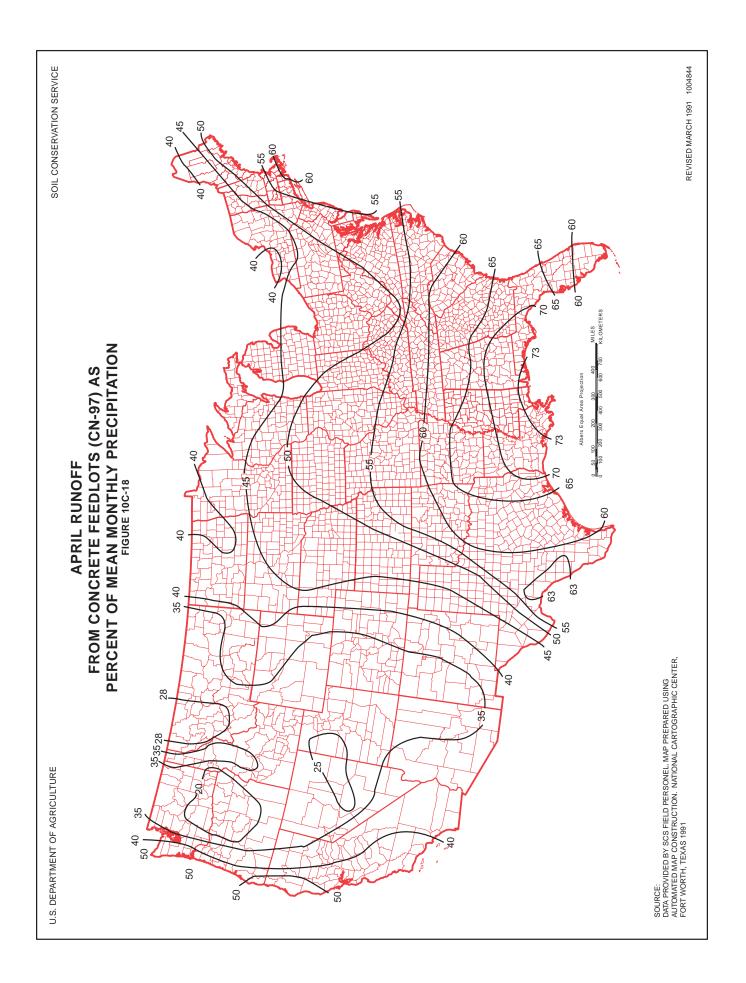


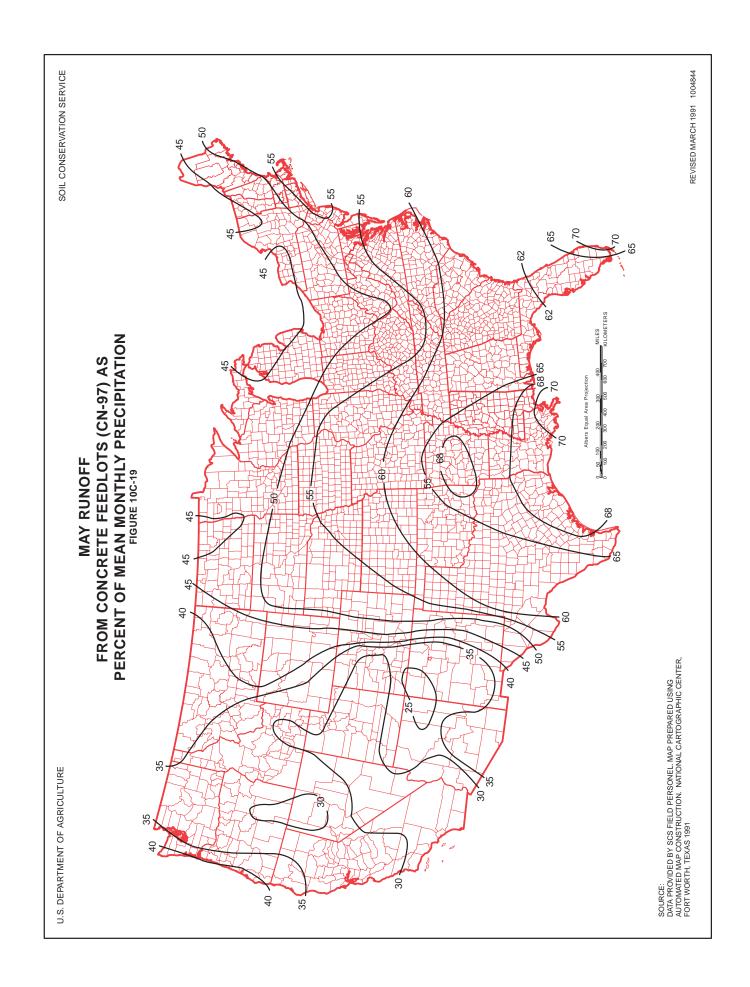


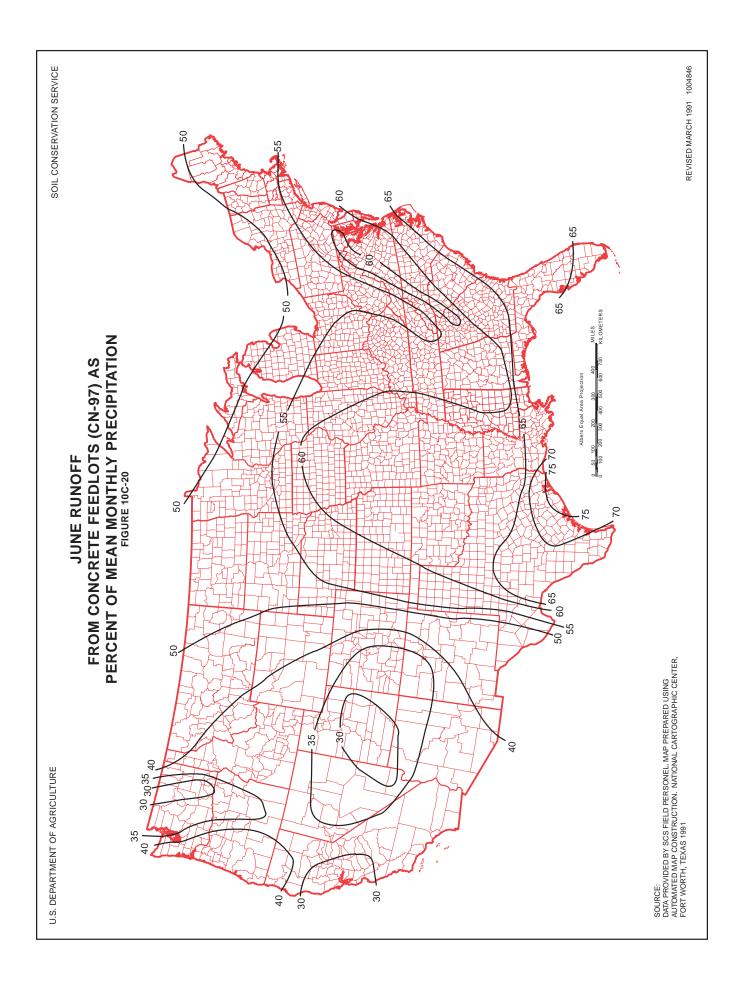


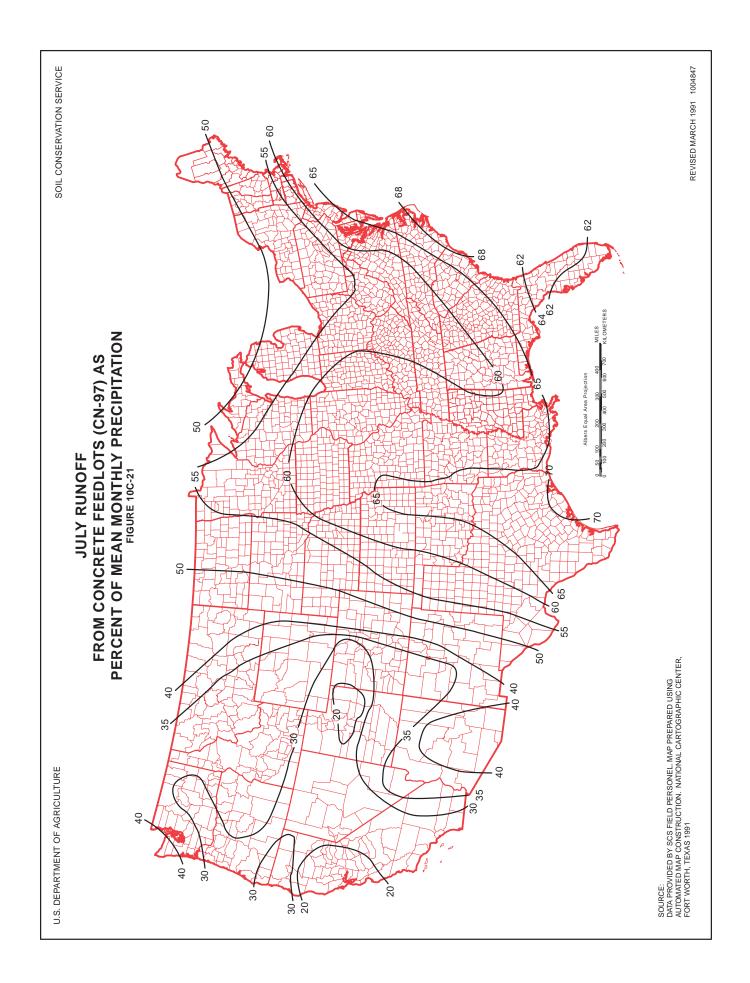


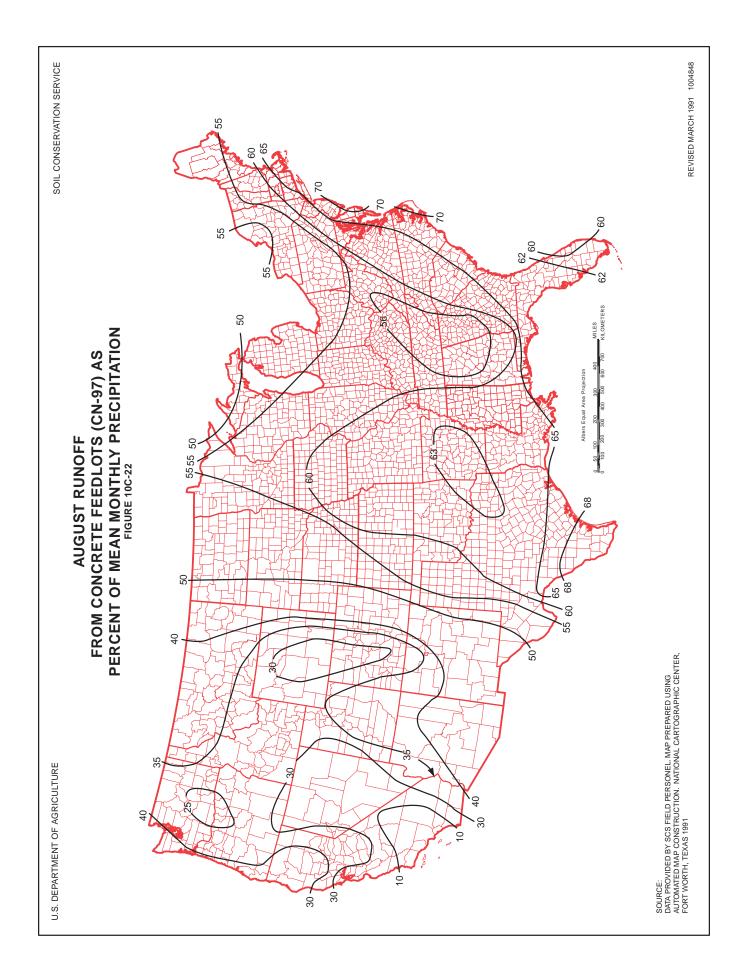


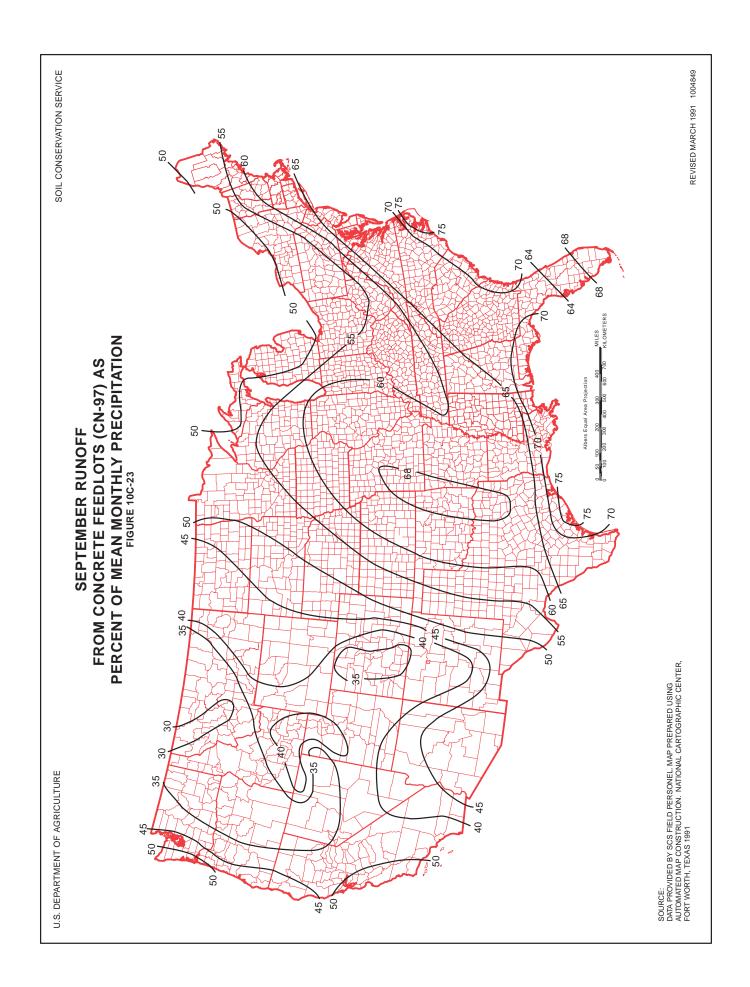


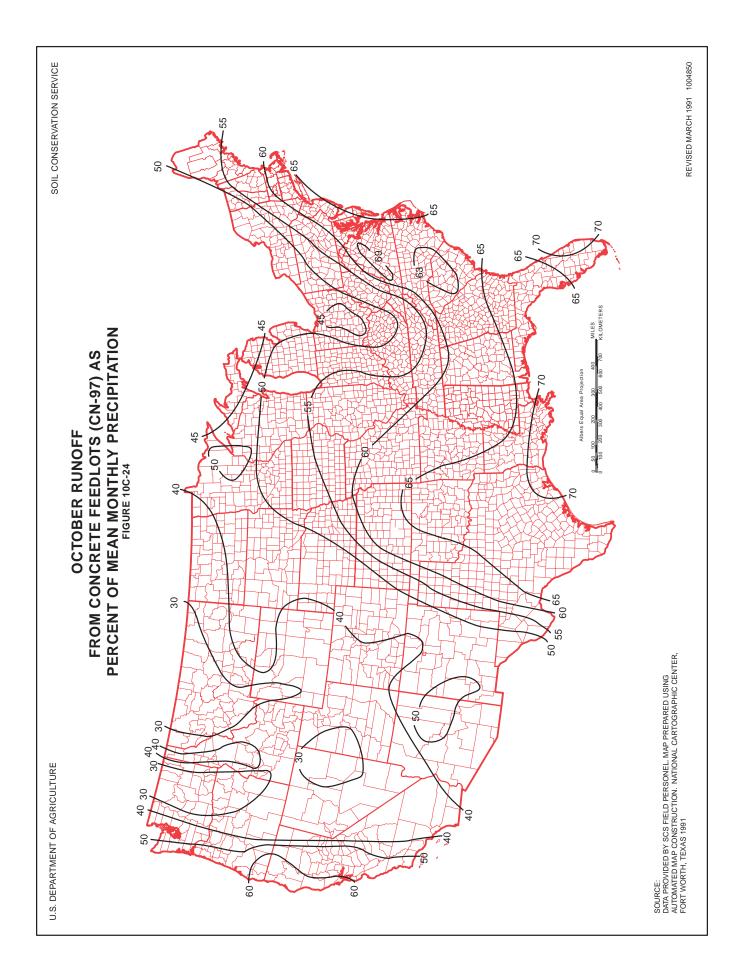


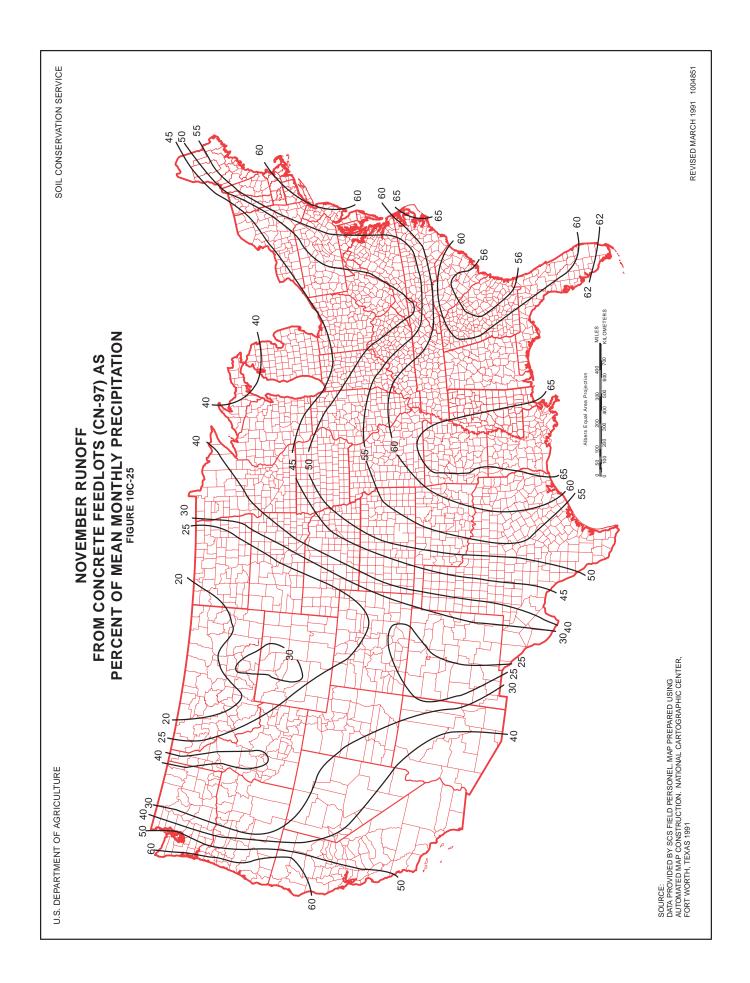


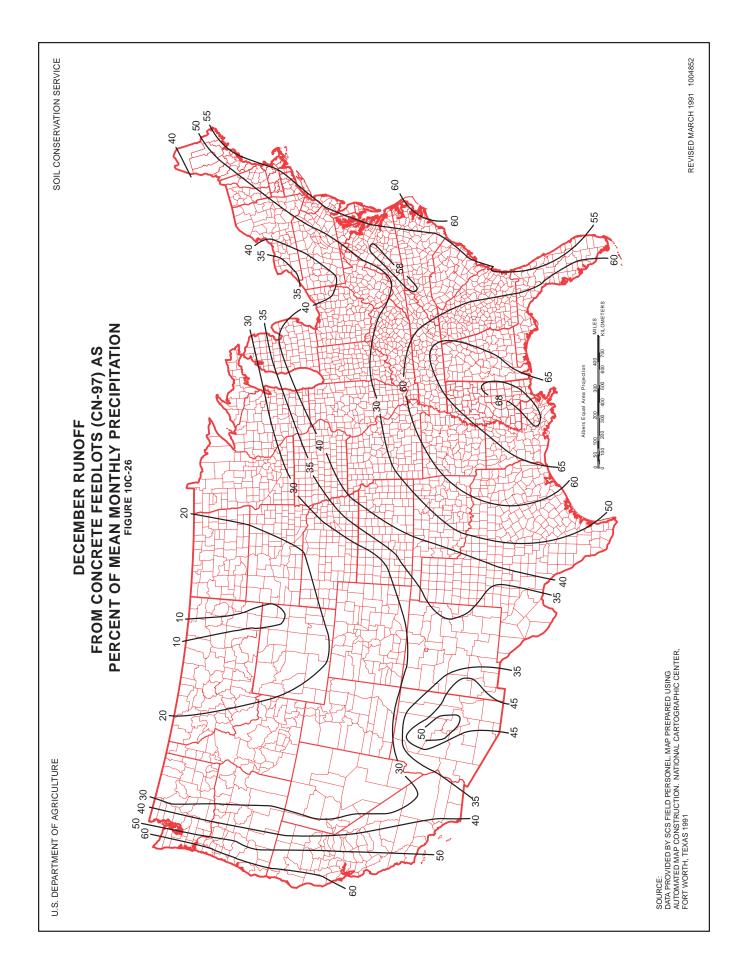


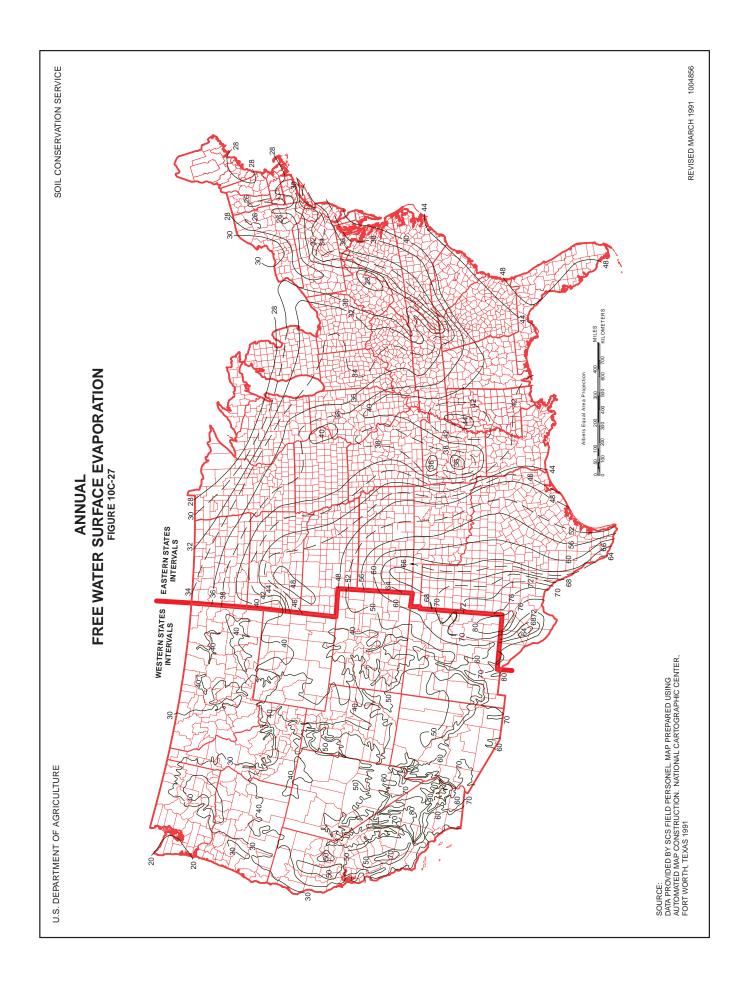












Appendix 10D	Agricultural Waste Management System	Part 654
	Component Design	National Engineering Handbook

Appendix 10D
Design and Construction
Guidelines for Waste
Impoundments Lined with Clay or
Amendment-Treated Soil

Appendix 10D	Agricultural Waste Management System Component Design	Part 651 Agricultural Waste Management Field Handbook

Appendix 10D

Design and Construction Guidelines for Impoundments Lined with Clay or Amendment-Treated Soil

Introduction

Waste storage ponds and treatment lagoons are used in agricultural waste management systems to protect surface and ground water and as a component in a system for properly utilizing wastes. Seepage from these structures has the potential to pollute surface water and underground aquifers. The principal factors determining the potential for downward and/or lateral seepage of the stored wastes are the:

- permeability of the soil and bedrock horizons near the excavated limits of a constructed waste treatment lagoon or waste storage pond,
- depth of liquid in the pond that furnishes a driving hydraulic force to cause seepage, and
- thickness of low permeability horizons between the boundary of the lagoon bottom and sides and the distance to the aquifer or water table

In some circumstances, where permitted by local and/ or State regulations, designers may consider whether seepage may be reduced from the introduction of manure solids into the reservoir. Physical, chemical, and biological processes can occur that reduce the permeability of the soil-liquid interface. Suspended solids settle out and physically clog the pores of the soil mass. Anaerobic bacteria produce by-products that accumulate at the soil-liquid interface and reinforce the seal. The soil structure can also be altered in the process of metabolizing organic material.

Chemicals in waste, such as salts, can disperse soil, which may also be beneficial in reducing seepage. Researchers have reported that, under some conditions, the seepage rates from ponds can be decreased by up to an order of magnitude (reduced 1/10th) within a year following filling of the waste storage pond or treatment lagoon with manure. Manure with higher solids content is more effective in reducing seepage than manure with fewer solids content. Research has shown that manure sealing only occurs when soils have a minimal clay content. A rule of thumb supported by research is that manure sealing is not effective unless soils have at least 15 percent clay content for monogastric animal generated waste and 5 percent clay content for ruminant animal generated waste (Barrington, Jutras, and Broughton 1987a, 1987b). Manure sealing is not considered effective on relatively clean sands and gravels, and these soils always require a liner as described in the following sections.

Animal waste storage ponds designed prior to about 1990 assumed that seepage from the pond would be minimized by the accumulation of manure solids and a biological seal at the foundation surface. Figure 10D–1 shows one of these early sites, where the soils at grade were somewhat permeable sands. Monitoring wells installed at some sites with very sandy soils showed that seepage containing constituents from the pond was still occurring even after enough time had passed that manure sealing should have occurred.

This evidence caused U.S. Department of Agriculture (USDA) Natural Resources Conservation Service (NRCS) engineers to reconsider guidance on suitable soils for siting an animal waste storage pond. In the late 1980s guidance was developed that designs should not rely solely on the seepage reduction that might occur from the accumulation of manure solids in the bottom and on the sides of the finished structure. That initial design document was entitled "South National Technical Center (SNTC) Technical Guide 716." It suggested that if any of four site conditions were present at a proposed structure location, a clay liner or other method of reducing seepage would be used in NRCS designs. A few revisions were made, and the document was re-issued in September 1993.

Figure 10D-1

Animal waste storage pond constructed before the implementation of modern design guidelines



NRCS was reorganized in 1994, and guidance in old SNTC documents was not part of the revised document system of the Agency. Consequently, the 716 document was revised considerably, and the revised material was incorporated into appendix 10D of the Agricultural Waste Field Management Handbook (AWMFH) in October 1998. This 2008 version of appendix 10D continues to update and clarify the process of designing an animal waste storage pond that will meet NRCS-specified engineering design criteria and stated specified permeability requirements.

General design considerations

Limiting seepage from an agricultural waste storage pond has two primary goals. The first is to prevent any virus or bacteria from migrating out of the storage facility to an aquifer or water source. The second is to prevent the conversion of ammonia to nitrate in the vadose zone. Nitrates are very mobile once they are formed by the nitrification process. They can then accumulate significantly in ground water. The National drinking water standard for nitrate is 10 parts per million, and excessive seepage from animal waste storage ponds could increase the level of nitrates in ground water above this threshold. Other constituents in the liquid manure stored in ponds may also be potential contaminants if the seepage from the pond is unacceptably high.

Defining an acceptable seepage rate is not a simple task. Appendix 10D recommends an allowable seepage quantity that is based on a historically accepted tenet of clay liner design, which is that a coefficient of permeability of 1×10^{-7} centimeters per second is reasonable and prudent for clay liners. This value, rightly or wrongly, has a long history of acceptability in design of impoundments of various types, including sanitary landfills. The seepage rate considered acceptable by NRCS is based on this permeability rate, also considering the following:

 When credit for a reduction of seepage from manure sealing (described later in the document) is allowed, NRCS guidance considers an acceptable initial permeability value to be 1×10⁻⁶ centimeters per second. This higher value used for design assumes that manure sealing will result in a tenth reduction in the initial seepage. Other assumptions are that typical NRCS waste impoundments have a depth of liquid of about 9 feet and typical clay liners are 1 foot thick. The computed seepage rate before manure sealing took effect is then about 9,240 gallons per acre per day, and this rate would reduce to 924 gallons per acre per day when manure sealing reduced the seepage by one tenth. To introduce some conservatism into the design, the NRCS guidance allows a seepage rate of 5,000 gallons per acre per day for initial designs unless State or local regulations are more restrictive, in which case those requirements should be followed.

• If State or local regulations prohibit designs from taking credit for future reductions in seepage from manure sealing, then NRCS recommends the initial design for the site be based on a seepage rate of 1,000 gallons per acre per day, the approximate seepage predicted for a site with 9 feet of head, a 1-foot-thick clay liner and a coefficient of permeability in the liner of 1×10^{-7} centimeters per second. Applying an additional safety factor to this value is not recommended because it conservatively ignores the potential benefits of manure sealing.

One problem with basing designs on a unit seepage value is that the approach considers only unit area seepage. The same criterion applies for small and large facilities. More involved three-dimensional type analyses would be required to evaluate the potential impact of seepage on ground water regimes on a whole-site basis. In addition to unit seepage, studies for large storage facilities should consider regional ground water flow, depth to the aquifer likely to be affected, and other factors.

The procedures in appendix 10D to the AWMFH provide a rational approach to selecting an optimal combination of liner thickness and permeability to achieve a relatively economical, but effective, liner design. It recognizes that manipulating the permeability of the soil liner is usually the most cost-effective approach to reduce seepage quantity. While clay liners obviously allow some seepage, the limited seepage from a properly designed site should have minimal impact on ground water quality. Numerous studies, such as those done by Kansas State University (2000), have shown that waste storage ponds located in low permeability soils

of sufficient thickness have a limited impact on the quality of ground water.

If regulations or other considerations cause a design to be devised with a goal of reducing unit seepage to less than 500 gallons per acre per day (1/56 inch per day), NRCS engineers' opinions are that synthetic liners such as high-density polyethylene (HDPE), linear low-density polyethylene (LLDPE), ethylene propylene diene monomer (EPDM), or geosynthetic clay liners (GCL), concrete liners, or aboveground storage tanks

will be required to achieve lower rates. Figure 10D–2 shows a pond lined with a synthetic liner, figure 10D–3 shows a concrete-lined excavated pond, and figure 10D–4 shows an aboveground concrete tank. Aboveground tanks may be also constructed of fiberglass-lined steel. NRCS has significant expertise in the selection, specification, and construction of sites using these products in addition to clay liners. Guidance on these other technologies is contained in other chapters of the AWMFH.

Figure 10D-2 Pond with synthetic liner (Photo credit NRCS)



Figure 10D-3 Excavated animal waste storage pond with concrete liner (*Photo credit NRCS*)

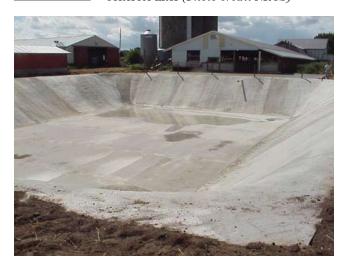


Figure 10D-4 Aboveground storage tank for animal waste (Photo credit Mitch Cummings, Oregon NRCS)



Progressive design

Waste storage ponds and waste treatment lagoons are usually designed with specific objectives that include cost, allowable seepage, aesthetics, and other considerations. Designs are usually evaluated in a progressive manner, with less costly and simple methods considered first, and more costly and complex methods considered next. These design concepts should generally be considered in the order listed to provide the most economical, yet effective, design of these structures. The following descriptions cover details on design and installation of these individual design measures.

- The least expensive and least complex design is to locate a waste impoundment in soils that have a naturally low permeability and where horizons are thick enough to reduce seepage to acceptable levels. The site should also be located where the distance to the water table conforms to requirements of any applicable regulations.
- Soils underlying the excavated boundaries of the pond may not be thick enough or slowly permeable enough to limit seepage to acceptably low values. In this case, the next type of design often considered is a liner constructed of compacted clay or other soils with appropriate amendments. This type of liner may be constructed with soils from the excavation itself or soil may be imported from nearby borrow sources. If the soils require amendments such as bentonite or soil dispersants, the unit cost of the compacted liner will be significantly higher than for a liner that only requires compaction to achieve a satisfactorily low permeability.
- A synthetic liner may be used to line the impoundment to reduce seepage to acceptable levels. Various types of synthetic materials are available.
- A liner may be constructed of concrete, or a concrete or fiberglass-lined steel tank can be constructed above ground to store the wastes.

A useful tool in comparing design alternatives is to evaluate unit costs. Benefits of alternatives may then be compared against unit costs to aid in selecting a design alternative. Benefits may include reduced seepage, aesthetics, or other considerations. Many geomembrane suppliers may be able to provide rough cost estimates based on the size and locale of the site. In estimating the cost of a compacted clay liner, one should evaluate the volume of compacted fill involved in a liner of given thickness. Table 10D–1 illustrates a cost comparison for different thicknesses of compacted clay liners. If methods other than compacted clay liners are used, higher unit costs may apply (table 10D–2).

Table 10D-1 Cost comparisons of design options for compacted clay liner

Thickness of compact- ed liner (ft)	Number of cubic yards of fill per square foot (yd³)	Assumed cost of compacted fill, per cubic yard (\$)	Unit cost of stated thickness liner (\$/ft ²)
1.0	0.037037	3.00-5.00	0.11-0.19
1.5	0.055555	3.00-5.00	0.17-0.28
2.0	0.074074	3.00-5.00	0.22-0.37
3.0	0.111111	3.00-5.00	0.33-0.56

Table 10D-2 Cost comparison for other design options

Liner type	Unit costs (\$/ft ²)
Geosynthethic	0.50-1.25
Concrete, reinforced 5 inches thick	7.50-8.00

Soil properties

The permeability of soils at the boundary of a waste storage pond depends on several factors. The most important factors are those used in soil classification systems such as the Unified Soil Classification System (USCS). The USCS groups soils into similar engineering behavioral groups. The two most important factors that determine a soil's permeability are:

- The percentage of the sample which is finer than the No. 200 sieve size, 0.075 millimeters.
 The USCS has the following important categories of percentage fines:
 - Soils with less than 5 percent fines are the most permeable soils.
 - Soils with between 5 and 12 percent fines are next in permeability.
 - Soils with more than 12 percent fines but less than 50 percent fines are next in order of permeability.
 - Soils with 50 percent or more fines are the least permeable.
- The plasticity index (PI) of soils is another parameter that strongly correlates with permeability.

When considered together with percent fines, a grouping of soils into four categories of permeability is possible. The following grouping of soils is based on the experience of NRCS engineers. It may be used to classify soils at grade as an initial screening tool. Estimating permeability is difficult because so many factors determine the value for a soil. For *in situ* soils, the following factors, in addition to percent fines and PI, affect the permeability of the natural soils:

- The dry density of the natural soil affects the permeability. Soils with lower dry densities have higher percentage of voids (porosity) than more dense soils.
- Structure strongly affects permeability. Many clay soils, particularly those with PI values above 20, develop a blocky structure from desiccation. The blocky structure creates preferential flow paths that can cause soils to have an unexpectedly high permeability. Albrecht

- and Benson (2001) and Daniel and Wu (1993) describe the effect of desiccation on the permeability of compacted clay liners.
- While not considered in the USCS, the chemical composition of soils with clay content strongly affects permeability. Soils with a preponderance of calcium or magnesium ions on the clay particles often have a flocculated structure that causes the soils to be more permeable than expected based simply on percent fines and PI. Soils with a preponderance of sodium or potassium ions on the clay particles often have a dispersive structure that causes the soils to be less permeable than soils with similar values of percent fines and PI. The NRCS publication TR–28, Clay Minerals, describes this as follows:

In clay materials, permeability is also influenced to a large extent by the exchangeable ions present. If, for example, the Ca (calcium) ions in a montmorillonite are replaced by Na (sodium) ions, the permeability becomes many times less than its original value. The replacement with sodium ions reduces the permeability in several ways. For one thing, the sodium causes dispersion (disaggregation) reducing the effective particle size of the clay minerals. Another condition reducing permeability is the greater thickness of water adsorbed on the sodium-saturated montmorillonite surfaces which diminishes the effective pore diameter and retards the movement of fluid water.

- Alluvial soils may have thin laminations of silt or sand that cause them to have a much higher horizontal permeability than vertical permeability. This property is termed anisotropy and should be considered in flow net analyses of seepage.
- Other types of deposits may have structure resulting from their mode of deposition. Loess soils often have a high vertical permeability resulting from their structure. Glacial tills may contain fissures and cracks that cause them to have a permeability higher than might be expected based only on their density, percent fines and PI of the fines.

The grouping of soils in table 10D–3 is based on the percent passing the No. 200 sieve and PI of the soils. Table 10D–4 is useful to correlate the USCS groups to one of the four permeability groups.

Table 10D-3

Grouping of soils according to their estimated permeability. Group I soils are the most permeable, and soils in groups III and IV are the least permeable soils

Group	Description
I	Soils that have less than 20 percent passing a No. 200 sieve and have a PI less than 5
П	Soils that have 20 percent or more passing a No. 200 sieve and have PI less than or equal to 15. Also included in this group are soils with less than 20 percent passing the No. 200 sieve with fines having a PI of 5 or greater
III	Soils that have 20 percent or more passing a No. 200 sieve and have a PI of 16 to 30
IV	Soils that have 20 percent or more passing a No. 200 sieve and have a PI of more than 30

Table 10D–4 Unified classification versus soil permeability groups ^{1/}

Unified Soil Classification	1 -	Soil permeability group number and occurrence of USCS group in that soil				
System Group Name	I	II	III	IV		
СН	N	N	S	U		
MH	N	S	U	s		
CL	N	S	U	S		
ML	N	U	S	N		
CL-ML	N	A	N	N		
GC	N	S	U	S		
GM	S	U	S	S		
GW	A	N	N	N		
SM	S	U	S	S		
SC	N	S	U	S		
SW	A	N	N	N		
SP	A	N	N	N		
GP	A	N	N	N		

- ASTM Method D-2488 has criteria for use of index test data to classify soils by the USCS.
- A = Always in this permeability group
- N = Never in this permeability group
- S = Sometimes in this permeability group (less than 10 percent of samples fall in this group)
- U = Usually in this permeability group (more than 90 percent of samples fall in this group)

Permeability of soils

Table 10D–5 shows an approximate range of estimated permeability values for each group of soils in table 10D–3. The ranges are wide because the classification system does not consider other factors that affect the permeability of soils, such as the electrochemical nature of the clay in the soils. Two soils may have similar percent finer than the No. 200 sieves and PI values but have very different permeability because of their different electrochemical makeup. The difference can easily be two orders of magnitude (a factor of 100). The most dramatic differences are between clays that have a predominance of sodium compared to those with a preponderance of calcium or magnesium. High calcium soils are more permeable than high sodium soils.

Table 10D–5 summarizes the experienced judgment of NRCS engineers and generally used empirical correlations of other engineers. The correlations are for in situ soils at medium density and without significant structure or chemical content. Information shown in figure 10D–5 is also valuable in gaining insight into the probable permeability characteristics of various soil and rock types.

Some soils in groups III and IV may have a higher permeability than indicated in table 10D–5 because they contain a high amount of calcium. High amounts of calcium result in a flocculated or aggregated structure in soils. These soils often result from the weathering

Table 10D-5

Grouping of soils according to their estimated permeability. Group I soils are the most permeable and soils in groups III and IV are the least permeable soils.

Group	Percent fines	PI	Estimated permeabili	-
			Low	High
I	< 20	< 5	3×10 ⁻³	2
п	≥20	≤ 15	5×10 ⁻⁶	5×10 ⁻⁴
II	< 20	≥ 5	9×10	5×10
III	≥20	$16 \le PI \le 30$	5×10 ⁻⁸	1×10 ⁻⁶
IV	≥20	> 30	1×10 ⁻⁹	1×10^{-7}

of high calcium parent rock, such as limestone. Soil scientists and published soil surveys are helpful in identifying these soil types.

High calcium clays should usually be modified with soil dispersants to achieve the target permeability goals. Dispersants, such as tetrasodium polyphosphate, can alter the flocculated structure of these soils by replacement of the calcium with sodium. Because manure contains salts, it can aid in dispersing the structure of these soils, but design should not rely on manure as the only additive for these soil types.

Soils in group IV usually have a very low permeability. However, because of their sometimes blocky structure, caused by desiccation, high seepage losses can occur through cracks that can develop when the soil is allowed to dry. These soils possess good attenuation properties if the seepage does not move through cracks in the soil mass. Soils with extensive desiccation cracks should be disked, watered, and recompacted to destroy the structure in the soils to provide an acceptable permeability. The depth of the treatment required should be based on design guidance given in the section **Construction considerations for compacted clay liners**.

High plasticity soils like those in group IV should be protected from desiccation in the interim period between construction and filling the pond. Ponds with intermittent storage should also consider protection for high PI liners in their design.

Figure 10D-5 Permeability of various geologic material (from Freeze and Cherry 1979)

					cm ³ /cm ² /s (c	em/s)					
	10^{1}	1	10-1	10-2	10-3	10^{-4}	10-5	10-6	10-7		10-8
	•				ft³/ft²/d (ft	:/d)		•	•		<u> </u>
10^{5}	10^{4}	10	10^3 10^2	10	1 1	10)-1	10-2	10-3	10^{-4}	10-5
	<u> </u>			I	ft³/ft²/min (ft	/min)					
	10^{1}	1	10-1	10-2	10-3	10^{-4}	10) ⁻⁵	10-6	10-7	10-8
			•	<u> </u>	gal/ft²/d (gal/	/ft²/d)					
	10^{5}	10^{4}	10^{3}	10^{2}	10^{1}	1	1	0-1	10-2	10-3	10-4
		•			m³/m²/day (m/d)					
	10^{4}	10^{3}	10^{2}	10^{1}	1	10-1	10-2	10-3	10-	1	10-5
		•		1	relative perme	eability					

	Very high	High	Moderate	Low	Very low
		Re	presentative materials		
Soil types	Clean gravel (GP)	Clean sand, clean sand and gravel mixes (GW, GP, SW, SP, SM)	Fine sand, silty sand and gravel mixes (SP, SM, GM, GW–GM, GP–GM, SW–SM, SP–SM)	Silt, clay, and sand-silt- clay mixes, organic silts, organic clays (GM, GC, SM, SC, MH, ML, ML–CL, OL, OH, GW–GC, GC–GM, SW-SC, SP–SC, SC–SM)	Massive clay, no soil joints or other macropores (CL, CH)
			Any soil mass with joints, cr	acks or other macroporosity	

Rock Cavernous and karst limestones types and dolomites, permeable basalts

Limestones, dolomites, clean sandstones

Interbedded sandstones, siltstones, and shales

Most massive rocks, unfractured and unweathered

Fractured igneous and metamorphic rocks

In situ soils with acceptable permeability

For screening purposes, NRCS engineers have determined that if the boundaries of a planned pond are underlain on the sides and bottom both by a minimum thickness of natural soil in permeability groups III or IV, the seepage from those ponds is generally low enough to cause no degradation of ground water. This assumes that soils do not have a flocculated structure. Unless State regulations or other requirements dictate a more conservative method of limiting seepage, it is the position of NRCS that special design measures generally are not necessary where agricultural waste storage ponds or treatment lagoons are constructed in these soils, provided that:

- at least 2 feet of natural soil in groups III or IV occur below the bottom and sides of the lagoon
- the soils are not flocculated (high calcium)
- no highly unfavorable geologic conditions, such as karst formations, occur at the site
- the planned depth of storage is less than 15 feet

Ponds with more than 15 feet of liquid should be evaluated by more precise methods. If the permeability and thickness of horizons beneath a structure are known, the predicted seepage quantities may be estimated more precisely. In some cases, even though a site is underlain by 2 feet of naturally low permeability soil, an acceptably low seepage rate satisfactory for some State requirements cannot be documented. In those cases, more precise testing and analyses are suggested. The accumulation of manure can provide a further decrease in the seepage rate of ponds by up to 1 order of magnitude as noted previously. If regulations permit considering this reduction, a lower predicted seepage can be assumed by designers.

Definition of pond liner

Compacted clay liner—Compacted clay liners are relatively impervious layers of compacted soil used to reduce seepage losses to an acceptable level. A liner for a waste impoundment can be constructed in several ways. When soil alone is used as a liner, it is often called a clay blanket or impervious blanket. A

simple method of providing a liner for a waste storage structure is to improve a layer of the soils at the excavated grade by disking, watering, and compacting the soil to a thickness indicated by guidelines in following sections. Compaction is often the most economical method for constructing liners if suitable soils are available nearby or if soils excavated during construction of the pond can be reused to make a compacted liner. Soils with suitable properties can make excellent liners, but the liners must be designed and installed correctly. Soil has an added benefit in that it provides an attenuation medium for many types of pollutants. NRCS Conservation Practice Standard (CPS) 521D, Pond Sealing or Lining Compacted Clay Treatment, addresses general design guidance for compacted clay liners for ponds.

If the available soils cannot be compacted to a density and water content that will produce an acceptably low permeability, several options are available, and described in the following section. The options involve soil additives to improve the permeability of the soils and adding liners constructed of materials other than natural soils.

Treat the soil at grade with bentonite or a soil dispersant—Designers must be aware of which amendment is appropriate for adding to specific soils at a site. In the past, bentonite has been inappropriately used to treat clay soils and soil dispersants have inappropriately been used to treat sands with a small clay content.

The following guidelines are helpful and should be closely followed.

• When to use bentonite—Soils in groups I and II have unacceptably high permeability because they contain an insufficient quantity of clay or the clay in the soils is less active than required. A useful rule of thumb is that soils amenable for treatment with bentonite will have PI values less than 7, or they will have less than 30 percent finer than the No. 200 sieve, or both.

Bentonite is essentially a highly concentrated clay product that can be added in small quantities to a sand or slightly plastic silt to make it relatively low in permeability. CPS 521C, Pond Sealing or Lining Bentonite Treatment, covers this practice. NRCS soil mechanics laboratories have found it important to use the same type

and quality of bentonite planned for construction in the laboratory permeability tests used to design the soil-bentonite mixture. Both the quality of the bentonite and how finely ground the product is before mixing with the soil will strongly affect the final permeability rate of the mixture. It is important to work closely with both the bentonite supplier and the soil testing facility when designing treated soil liners.

• When to use soil dispersants—Soils in groups III and IV may have unacceptably high permeability because they contain a preponderance of calcium or magnesium on the clay particles. Unfortunately, field or lab tests to determine when soils are likely to have this problem are not available. High calcium soils often occur when parent materials have excessive calcium. Many soils developed from weathering of limestone and gypsum may have this problem. See the section Design and construction of clay liners treated with soil dispersants, for more detail. Some States require the routine use of soil dispersants in areas that are known to have high calcium clay soils.

Use of concrete or synthetic materials such as geomembranes and geosynthetic clay liners

(GCLs)—Concrete has advantages and disadvantages for use as a liner. A disadvantage is that it will not flex to conform to settlement or shifting of the earth. In addition, some concrete aggregates may be susceptible to attack by continued exposure to chemicals contained in or generated by the waste. An advantage is

Figure 10D-6

Agricultural waste storage impoundment lined with a geomembrane ($Photo\ credit\ NRCS$)



that concrete serves as an excellent floor from which to scrape solids. It also provides a solid support for equipment such as tractors or loaders.

Geomembranes and GCLs are the most impervious types of liners if designed and installed correctly. Care must be exercised both during construction and operation of the waste impoundment to prevent punctures and tears. The most common defects in these liners arise from problems during construction. Forming seams in the field for geomembranes can require special expertise. GCLs have the advantage of not requiring field seaming, but overlap is required to provide a seal at the seams. Geomembranes must contain ultraviolet inhibitors if exposed to sunlight. Designs should include provision for protection from damage during cleaning operations. Concrete pads, double liners, and soil covering are examples of protective measures. Figure 10D-6 shows an agricultural waste storage facility with a geomembrane liner with ultraviolet inhibitors.

When a liner should be considered

A constructed liner may be required if any of the conditions listed are present at a planned impoundment.

Proposed impoundment is located where any underlying aquifer is at a shallow depth and not confined and/or the underlying aquifer is a domestic or ecologically vital water supply—State or local regulations may prevent locating a waste storage impoundment within a specified distance from such features. Even if the pond bottom and sides are underlain by 2 feet of naturally low permeability soil, if the depth of liquid in the pond is high enough, computed seepage losses may be greater than acceptable. The highest level of investigation and design is required on sites like those described. This will ensure that seepage will not degrade aquifers at shallow depth or aquifers that are of vital importance as domestic water sources.

Excavation boundary of an impoundment is underlain by less than 2 feet of suitably low permeability soil, or an equivalent thickness of soil with commensurate permeability, over bedrock—Bedrock that is near the soil surface is often fractured or jointed because of weathering and stress relief.

Many rural domestic and stock water wells are developed in fractured rock at a depth of less than 300 feet. Some rock types, such as limestone and gypsum, may have wide, open solution channels caused by chemical action of the ground water. Soil liners may not be adequate to protect against excessive leakage in these bedrock types. Concrete or geomembrane liners may be appropriate for these sites. However, even hairline openings in rock can provide avenues for seepage to move downward and contaminate subsurface water supplies. Thus, a site that is shallow to bedrock can pose a potential problem and merits the consideration of a liner. Bedrock at a shallow depth may not pose a hazard if it has a very low permeability and has no unfavorable structural features. An example is massive siltstone.

Excavation boundary of an impoundment is underlain by soils in group I—Coarse grained soils with less than 20 percent low plasticity fines generally have higher permeability and have the potential to allow rapid movement of polluted water. The soils are also deficient in adsorptive properties because of their lack of clay. Relying solely on the sealing resulting from manure solids when group I soils are encountered is not advisable. While the reduction in permeability from manure sealing may be one order of magnitude, the final resultant seepage losses are still likely to be excessive, and a liner should be used if the boundaries of the excavated pond are in this soil group.

Excavation boundary of an impoundment is underlain by some soils in group II or problem soils in group III (flocculated clays) and group IV (highly plastic clays that have a blocky structure)—Soils in group II may or may not require a liner. Documentation through laboratory or field permeability testing and computations of specific discharge (unit seepage quantities) is advised. Higher than normal permeability can occur when soils in group III or IV are flocculated or have a blocky structure. These are special cases, and most soils in groups III and IV will not need a liner provided the natural formation is thick enough to result in acceptable predicted seepage quantities.

These conditions do not always dictate a need for a liner. Specific site conditions can reduce the potential risks otherwise indicated by the presence of one of these conditions. For example, a thin layer of soil over high quality rock, such as an intact shale, is less risky than if the thin layer occurs over fractured or fissured rock. If the site is underlain by many feet of intermediate permeability soil, that site could have equivalent seepage losses as one underlain by only 2 feet of low permeability soil.

Some bedrock may contain large openings caused by solutioning and dissolving of the bedrock by ground water. Common types of solutionized bedrock are limestone and gypsum. When sinks or openings are known or identified during the site investigation, these areas should be avoided and the proposed facility located elsewhere. However, when these conditions are discovered during construction or alternate sites are not available, concrete or geosynthetic liners may be required, but only after the openings have been properly cleaned out and backfilled with concrete.

Specific discharge

Introduction

One way to require a minimal design at a site is to require a minimum thickness of a given permeability soil for a natural or constructed liner. An example of this would be to require that a clay liner constructed at a waste storage pond should be at least 1 foot thick, and the soil should have a coefficient of permeability of 1×10^{-7} centimeters per second or less.

However, using only permeability and thickness of a boundary horizon as a criterion ignores the effect of the depth of liquid on the predicted quantity of seepage from an impoundment. Using this approach would mean that the same design would be used for a site with 30 feet of water as one with 8 feet of water, for instance. A more rational method for stating a limiting design requirement is to compute seepage using Darcy's law for a unit area of the pond bottom.

A rational method of comparing design alternatives at a given site is needed. Such a method allows designers to evaluate the effect of changing one or more of the design elements in a site on the predicted seepage quantities. This document presents methods for computing the term "specific discharge" to use in comparing alternatives and to document a given design goal for a site. Specific discharge is defined as unit seepage.

It does not reflect the total seepage from a site, but rather provides a value of seepage per square unit area of pond bottom.

This document uses calculations of specific discharge to compare design alternatives and to determine if a given design meets regulatory requirements and guidelines. In some cases, the total seepage from a pond may be of interest, particularly for larger ponds in highly environmentally sensitive environments.

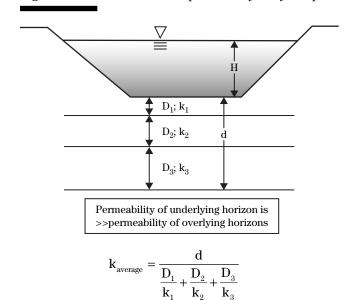
In those cases, more elaborate three-dimensional seepage computations using sophisticated finite-element computer programs may be warranted. It is outside the scope of this document to describe these types of analyses. Specialists who are experienced in using the complex software used for these computations should be consulted.

The parameters that affect the seepage from a pond with a natural or constructed clay liner are:

 The size of the pond—The total bottom area and area of the exposed sides of the pond holding the stored waste solids and liquids.

- The thickness of low permeability soil at the excavation limits of the pond—For design, the thickness of the soil at the bottom of the pond is often used because that is where seepage is likely to be highest. In some cases, however, seepage from the sides of the pond may also be an important factor. Seepage from the sides of ponds is best analyzed using finite element flow net programs. In some cases, rather than a single horizon, multiple horizons may be present.
- The depth of liquid in the pond—The depth of liquid at the top of the reservoir when pumping should commence is normally used.
- The coefficient of permeability of the soil forming the bottom and sides of the pond—In layered systems, an average or weighted permeability may be determined as shown in figure 10D-7.

Figure 10D-7 Conversion of permeability in layered profile to single value

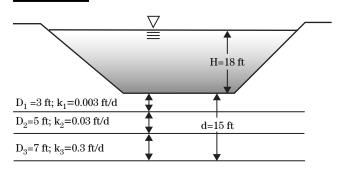


Example 10D–1 shows how to convert a multiple layer system into a single equivalent permeability. Using this method allows a designer to compute specific discharge when several horizons of constructed or natural soils occur below a site.

Example 10D-1

The excavated pond is underlain by 15 feet of soil consisting of three different horizons (fig. 10D–8). The thickness and permeability of each horizon is shown in the sketch. Compute the average vertical permeability of the 15 feet of soil.

Figure 10D–8 Idealized soil profile for example 10D–1



Solution

$$\mathbf{k}_{
m average} = rac{\mathbf{d}}{rac{\mathbf{D}_1}{\mathbf{k}_1} + rac{\mathbf{D}_2}{\mathbf{k}_2} + rac{\mathbf{D}_3}{\mathbf{k}_3}}$$
 $\mathbf{k}_{
m a} = rac{15}{\mathbf{D}_3} = -\frac{1}{2}$

$$\frac{3}{0.003} + \frac{5}{0.03} + \frac{7}{0.3}$$

Definition of specific discharge

The term "specific discharge" has been coined to denote the unit seepage that will occur through the bottom of a pond with a finite layer of impervious soil. Specific discharge is the seepage rate for a unit cross-sectional area of a pond. It is derived from Darcy's law as follows. First, consider Darcy's law.

$$Q = k \times i \times A$$

For a pond with either a natural or constructed liner, the hydraulic gradient is the term i in the equation, and it is defined in figure 10D–9 as equal to (H+d)/d.

Given

The Darcy's law for this situation becomes:

$$Q = k \times \frac{H + d}{d} \times A$$

where:

Q = total seepage through area A (L³/T)

k = coefficient of permeability (hydraulic

conductivity) $(L^3/L^2/T)$

i = hydraulic gradient (L/L)

H = vertical distance measured between the top of the liner and top of the liquid storage of the waste impound-

ment (fig. 10D-9) (L)

(L)

d = thickness of the soil liner (fig. 10D-9)

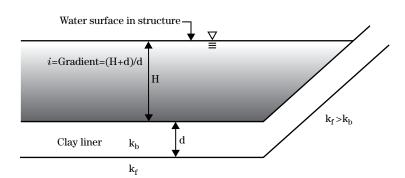
A = cross-sectional area perpendicular to

flow (L^2)

L = length

T = time

Figure 10D-9 Definition of terms for clay liner and seepage calculations



Rearrange terms:

$$\frac{Q}{A} = \frac{k(H+d)}{d}$$
 (L/T)

By definition, unit seepage or specific discharge, is $Q \div A$. The symbol v is used for specific discharge:

$$v = \frac{k(H+d)}{d} \tag{$L^3/L^2/T$}$$

Specific discharge may be confused with permeability because the units are the same. In the metric system, specific discharge and permeability are often expressed in units of centimeters per second. The actual units are cubic centimeters of flow per square centimeter of cross section per second, but this reduces to centimeters per second. Specific discharge is different than permeability because specific discharge is an actual flow rate of liquid through a cross section of a soil mass, whereas permeability is a property of the soil mass itself. Permeability is independent of the hydraulic gradient in a particular site, whereas specific discharge accounts for both permeability of the soil and the gradient causing the flow, as illustrated in figure 10D-9. Because hydraulic gradient is dimensionless, the units of specific discharge and permeability are then the same.

Because specific discharge expressed as L/T has the same units as velocity, specific discharge is often misunderstood as representing the average rate or velocity of water moving through a soil body rather than a quantity rate flowing through the soil. Because the water flows only through the soil pores, the actual cross-sectional area of flow is computed by multiplying the soil cross section (A) by the porosity (n). The seepage velocity is then equal to the unit seepage or specific discharge, ν , divided by the porosity of the soil, n. Seepage velocity = (ν/n) . In compacted liners, the porosity usually ranges from 0.3 to 0.5. The result is that the average linear velocity of seepage flow is two to three times the specific discharge value. The units of seepage velocity are L/T.

To avoid confusion between specific discharge and permeability, a strong recommendation is to use different units for specific discharge than for the coefficient of permeability. Common units for permeability are recommended to be in feet per day or centimeters per second. Units for specific discharge should be in

gallons per acre per day, acre-feet per acre per day, or acre-inches per acre per day.

To illustrate a typical computation for specific discharge, assume the following:

- A site has a liquid depth of 12 feet.
- The site is underlain by 2 feet of soil that has a coefficient of permeability of 1×10^{-6} centimeters per second (assume that a sample was obtained at the grade of the pond and sent to a laboratory where a flexible wall permeability test was performed on it).
- Compute the specific discharge, v. First, the coefficient of permeability may be converted to units of feet per day by multiplying the given units of centimeters per second by 2,835.

$$k = (1 \times 10^{-6} \text{ cm/s}) \times 2,835 = 0.002835 \text{ ft/d}$$

Then, the specific discharge ν is computed as follows:

$$v = k \times \frac{H + d}{d}$$

$$= 0.002835 \times \frac{12 + 2}{2}$$

$$\approx 0.02 \text{ ft}^3/\text{ft}^2/\text{d}$$

$$\approx 0.02 \text{ ft}/\text{d}$$

Conversion factors for specific discharge are given in table 10D-6.

 Table 10D-6
 Conversion factors for specific discharge

To convert from	To units of	Multiply by
ft ³ /ft ² /d	in ³ /in ² /d	12
ft ³ /ft ² /d	gal/acre/d	325,829
in ³ /in ² /d	gal/acre/d	27,152.4
in ³ /in ² /d	$\mathrm{cm}^3/\mathrm{cm}^2/\mathrm{s}$	2.94×10^{-5}
cm ³ /cm ² /s	gal/acre/d	9.24×10^{8}
cm ³ /cm ² /s	$in^3/in^2/d$	34,015
cm ³ /cm ² /s	$\mathrm{ft}^3/\mathrm{ft}^2/\mathrm{d}$	2,835

To convert the computed specific discharge in the example into units of gallons per acre per day and cubic inches per square inch per day (in/d), use conversion factors given in table 10D–6.

- 0.02 foot per day×325,829 ≅ 6,500 gallons per acre per day
- 0.02 foot per day×12 = 0.24 cubic inch per square inch per day

A variety of guidelines have been used and regulatory requirements stated for specific discharge. Usually, guidelines require the specific discharge for a given waste storage structure to be no higher than a stated value. The following example demonstrates the unit seepage that will result from a typical size animal waste storage lagoon or storage pond with 2 feet of either very good natural soil or a very well constructed, 2-foot-thick clay liner in the bottom of the lagoon. A practical lower limit for the assumed permeability of a compacted clay or a very good natural liner is a coefficient of permeability equal to 5×10^{-8} centimeters per second. This is based on considerable literature on field and laboratory tests for compacted clay liners used in sanitary landfills.

The specific discharge for this ideal condition follows, assuming:

- The pond has a liquid depth of 15 feet.
- The site is underlain by 2 feet of soil (either a natural layer or a constructed clay liner) that has a coefficient of permeability of 5×10⁻⁸ centimeters per second
- Compute the specific discharge, v. First, the coefficient of permeability is converted to units of feet per day by multiplying the given units of centimeters per second by 2,835. Then,

Table 10D-7 Typical requirement for specific discharge used by State regulatory agencies

Example specific discharge value	Equivalent value in gallons per acre per day
1/56 in ³ /in ² /d	485
$1/8 \text{ in}^3/\text{in}^2/\text{d}$	3,394
$1/4 \text{ in}^3/\text{in}^2/\text{d}$	6,788
$1 \times 10^{-6} \text{ cm}^3/\text{cm}^2/\text{s}$	924

$$k = (1 \times 10^{-6} \text{ cm/s}) \times 2,835 = 0.002835 \text{ ft/d}$$

Then, the specific discharge v is computed as follows:

$$\begin{aligned} v &= k \times \frac{H + d}{d} \\ &= 1.42 \times 10^{-4} \text{ ft/d} \times \frac{15 \text{ ft} + 2 \text{ ft}}{2 \text{ ft}} \\ &\cong 0.0012 \text{ ft}^3/\text{ft}^2/\text{d} \\ &\cong 0.0012 \text{ ft/d} \end{aligned}$$

Converting this into units of gallons per acre per day:

$$0.0012 \text{ ft/d} \times 325,829 \cong 393 \text{ gal/acre/d}$$

Table 10D–7 lists typical specific discharge values used by State regulatory agencies. Requirements vary from State to State. Individual designers may regard minimum requirements as too permissive. Some States permit a designer to assume that the initial computed seepage rate will be reduced in the future by an order of magnitude by taking credit for a reduction in permeability resulting from manure sealing. The State or local regulations should be used in design for a specific site.

If one assumes at least one order of magnitude of reduction in permeability will occur, the initial specific discharge can be 10 times greater, and the final value for specific discharge will approach a tenth of the initial rate after sealing.

Design of compacted clay liners

If a site does not have a sufficient thickness of $in\ situ$ low permeability soil horizons to limit seepage to an acceptably low value, a clay liner may be required. Some State regulations may also require a constructed clay liner regardless of the nature of the $in\ situ$ soils at a site. Regulations sometimes require a specific thickness of a compacted soil with a documented permeability of a given value. An example of this is a State requirement that a waste storage pond must have in the bottom and sides of the pond at least 2 feet of compacted clay with a documented coefficient of permeability of 1×10^{-7} centimeters per second.

Clay liners may also be designed based on a stated allowable specific discharge value. Computations

may be performed as detailed in following sections to determine a design that will meet a design specific discharge goal.

Detailed design steps for clay liners

The suggested steps for design of a compacted clay or amendment-treated liner are:

Step 1—Size the impoundment to achieve the desired storage requirements within the available construction limits and determine this depth or the height, H, of storage needed.

Step 2—Determine (from a geologic investigation) the thickness and permeability of horizons of natural clay underlying the bottom of the planned excavated pond. Investigate to a minimum of 2 feet below the planned grade of the pond or to depths required by State regulations, if greater. If natural low permeability horizons at least 2 feet thick or an equivalent thickness of soil with different permeability do not underlie the site, assume that a compacted clay liner (with or without amendments) will be constructed. The liner may be constructed of soils from the excavation if they are suitable for use, or soil may be imported from a nearby borrow source.

Step 3—Measure or estimate the permeability of the natural horizons or the compacted liner planned at the site. Use procedures shown in example 10D–1 to obtain a weighted permeability for the natural horizons.

Step 4—Compute the specific discharge using the values of head in the pond and thickness of natural horizons and their equivalent permeability in the specific discharge equation. If State or local regulations provide a required value for allowable specific discharge, design on the basis of those regulations. Currently, State regulations for specific discharge range from a low of about 500 gallons per acre per day (1/56 inch per day) to a high of about 6,800 gallons per acre per day (1/4 inch per day). If no regulations exist, a value of 5,000 gallons per acre per day may be used. If a designer feels that more conservative limiting seepage is advisable, that rate should be used in computations. It is seldom technically or economi-

cally feasible to meet a design specific discharge value of less than 500 gallons per acre per day using compacted clay liners or amendment-treated soil liners. To achieve lower values of unit seepage usually requires synthetic liners, concrete liners, or aboveground storage tanks.

Step 5—If the computed specific discharge meets design objectives, the site is satisfactory without additional design and may be designed and constructed.

Step 6—If the computed specific discharge at the site does not meet design objectives, use either method A or method B shown in following sections to design a compacted clay liner or a liner with soil amendment.

Notes to design steps:

- The calculated thickness of the soil liner required is sensitive to the relative values of soil
 permeability and the assumed allowable specific discharge value.
- The best and most economical way to reduce the required liner thickness is by reducing the soil's permeability. Liner permeability may be reduced by compacting soils to a higher degree, compacting them at a higher water content, and by using an appropriate additive such as bentonite or soil dispersants.
- By using higher compaction water contents and compacting soils to a high degree of saturation, permeability often can be reduced by a factor of 1/100.
- The liner soil must be filter compatible with the natural foundation upon which it is compacted.
 Filter compatibility is determined by criteria in NEH 633, chapter 26. As long as the liner soil will not pipe into the foundation, the magnitude of hydraulic gradient across the liner need not be limited.
- Filter compatibility is most likely to be a significant problem when a liner is constructed directly on top of very coarse soil, such as poorly graded gravels and gravelly sands.
- The minimum recommended thickness of a compacted clay liner is given in CPS 521D. The minimum thickness varies with the depth of liquid in the pond.

- Clay liners constructed by mixing bentonite with the natural soils at a site should have a minimum thickness shown in CPS 521C. These minimum thicknesses are based on construction considerations rather than calculated values for liner thickness requirement from the specific discharge equations. In other words, if the specific discharge equations indicate a 7-inch thickness of compacted bentonite-treated liner is needed to meet suggested seepage criteria, the CPS 521C could dictate a thicker liner. That guidance should be considered in addition to the specific discharge computations.
- Natural and constructed liners must be protected against damage by mechanical agitators or other equipment used for cleaning accumulated solids from the bottom of the structure. Liners should also be protected from the erosive forces of waste liquid flowing from pipes during filling operations. CPSs provide guidance for protection.
- Soil liners may not provide adequate confidence against ground water contamination
 if foundation bedrock beneath the pond contains large, connected openings. Collapse of
 overlying soils into the openings could occur.
 Structural liners of reinforced concrete or
 geomembranes should be considered because
 the potential hazard of direct contamination of
 ground water is significant.
- Liners should be protected against puncture from animal traffic and roots from trees and large shrubs. The subgrade must be cleared of stumps and large angular rocks before construction of the liner.
- If a clay liner (or a bentonite-treated liner) is allowed to dry, it may develop drying cracks or a blocky structure. Desiccation can occur during the initial filling of the waste impoundment and later when the impoundment is emptied for cleaning or routine pumping. Disking, adding water, and compaction are required to destroy this structure created by desiccation. A protective insulating blanket of less plastic soil may be effective in protecting underlying more plastic soil from desiccation during these times the liner is exposed. CPSs address this important consideration.

 Federal and State regulations may be more stringent than the design guidelines given, and they must be considered in the design. Examples later in this section address consideration of alternative guidelines.

Two methods for designing constructed clay liner

Two methods for designing a clay liner are available. In method A, designers begin with an assumed or required value for allowable specific discharge. Using the depth of liquid storage in the pond and known or estimated values of the liner's coefficient of permeability, a required thickness of liner is computed. If the value obtained is unrealistic, different values for the liner permeability are evaluated to determine what values produce a desirable thickness of liner. CPSs also determine minimum liner thicknesses.

In method B, designers begin with a desired thickness of liner and an assumed or required value for specific discharge. Using the depth of liquid storage in the pond and the desired thickness of liner, a required coefficient of permeability for the liner is computed. If the value obtained is unrealistic, different values for the liner thickness are evaluated to determine what values produce an achievable permeability. Coordinating with soil testing laboratories is helpful in evaluating alternatives that can provide the required permeability for the liner.

Each of these methods is illustrated with detailed design examples as follows:

Method A—Using assumed values for the coefficient of permeability of a compacted clay based on laboratory tests of the proposed liner soil, compute the required thickness of a liner to meet the given specific discharge design goal. In the absence of more restrictive State regulations, assume an acceptable specific discharge of 5,000 gallons per acre per day.

The required thickness of a compacted liner can be determined by algebraically rearranging the specific discharge equation, as follows. Terms have been previously defined.

$$d = \frac{k \times H}{v - k}$$

Note: If the k value assumed for the liner is equal to or greater than the assumed allowable specific discharge, meaningless results are attained for d, the calculated thickness of the liner in the last equation. The reason is that the denominator would be zero, or a negative number. Another way of stating this is that the allowable specific discharge goal cannot be met if the liner soils have k values equal to or larger than the assumed allowable specific discharge, in consistent units. Note also that CPS 521D has requirements for minimum thickness of compacted clay liners. If the computed value for the required thickness is less than that given in CPS 521D, then the values in the CPS must be used.

Example 10D–2—Design a clay liner using method A

Given:

Site design has a required depth of waste liquid, H, in the constructed waste impoundment of 12 feet. A soil sample was obtained and submitted to a soil mechanics laboratory for testing. A permeability test on a sample of proposed clay liner soil resulted in a permeability value of 6.5×10^{-7} centimeters per second (0.00184) ft/d) for soils compacted to 95 percent of maximum Standard Proctor dry density at a water content 2 percent wet of optimum. The State requirement for the site requires a specific discharge no greater than an eighth of an inch per day. Compute the required thickness of liner to be constructed of soil having the stated permeability that will achieve this specific discharge. What would be the effect of manure sealing on this computed requirement, if assumed reduction of seepage from manure sealing were permitted and was elected for use in the design?

Solution:

Step 1—First, convert the required specific discharge into the same units as will be used for the coefficient of permeability. Using values for permeability of feet per day, convert the stated eighth of an inch per day specific discharge requirement into feet per day. To convert, divide an eighth by 12 to obtain a specific discharge requirement of 0.010417 foot per day. It is given that the k value at the design density and water content is 0.00184 foot per day. Calculate the required minimum thickness of compacted liner as follows:

The equation for required d is:

$$d = \frac{k \times H}{v - k}$$

Using English system units, substituting the given values for H and k, assuming an allowable specific discharge, ν , of 0.010417 foot per day, then

$$d = \frac{0.00184 \text{ ft/d} \times 12 \text{ ft}}{0.010417 \text{ ft/d} - .00184 \text{ ft/d}} = 2.6 \text{ ft}$$

CPS 521D requires a pond with a depth of water of 12 feet to have a minimum thickness liner of 1 foot, so the 2.6 foot requirement governs.

Step 2—Assume that regulations permit considering the benefit of seepage reduction for manure sealing of one order of magnitude. Then, the design specific discharge may be 10 times the stated permissible value because manure sealing will reduce the initial seepage to the stated acceptable limits in a year or so of operation. The allowable specific discharge then becomes:

$$10 \times (0.010417 \text{ ft/d}) = 0.10417 \text{ ft/d}$$

Substituting into the equation solving for thickness of liner required:

$$d = \frac{k \times H}{v - k}$$

$$d = \frac{0.00184 \text{ ft/d} \times 12 \text{ ft}}{0.10417 \text{ ft/d} - 0.00184 \text{ ft/d}} = 0.2 \text{ ft}$$

Conclusion:

A compacted clay liner this thin is impractical. In this case, the minimum thickness liner required in CPS 521D of 1 foot would be used for design.

Method B—Using a given value for depth of liquid in the pond, assumed values for the thickness of a compacted clay based on construction considerations, CPS 521D requirements, State regulations, or the preference of the designer, compute the required permeability of a liner to meet the given specific discharge design goal. In the absence of more restrictive State regulations, assume an acceptable specific discharge of 5,000 gallons per acre per day. The required permeability of a compacted liner can be determined by algebraically rearranging the specific discharge equation as follows. Terms have been previously defined.

$$k = \frac{v \times d}{H + d}$$

If the computed value for the required permeability is less than 5×10^{-8} centimeters per second $(1.4\times10^{-4} \, \text{ft/d})$, NRCS engineers' experience is that lower values are not practically obtainable and a thicker liner or synthetic liners should be used to achieve design goals.

Example 10D–3—Designer a clay liner using method B

Given:

Site design has a required depth of waste liquid, H, in the constructed waste impoundment of 19 feet. CPS 521D requires a liner that is at least 18 inches (1.5 feet) thick. The site is in a State that allows NRCS design guidance of 5,000 gallons per acre per day to be used in the design. The NRCS guidance assumes that manure sealing will reduce this seepage value further and no additional credit should be taken.

Solution:

Step 1 First, convert the required specific discharge into the same units as will be used for the coefficient of permeability. Using values for permeability of feet per day, convert the stated 5,000 gallons per acre per day specific discharge requirement into feet per day. To convert using conversions shown in table 10D–6, divide 5,000 by 325,829 to obtain a specific discharge requirement of 0.0154 foot per day. The thickness of liner is given to be 1.5 feet. Calculate the required coefficient of permeability of the compacted liner as follows:

$$k = \frac{v \times d}{H + d}$$

Using English system units, substituting the given values for H of 19 feet and for d of 1.5 feet, assuming an allowable specific discharge, ν , of 0.0154 foot per day, then:

$$k = \frac{.0154 \text{ ft/d} \times 1.5 \text{ ft}}{19 \text{ ft} + 1.5 \text{ ft}}$$
$$= 1.1 \times 10^{-3} \text{ ft/d}$$

Convert to centimeters per second by dividing by 2,835.

$$k = \frac{1.1 \times 10^{-3} \text{ ft/d}}{2,835}$$
$$k = 4.0 \times 10^{-7} \text{ cm/s}$$

Step 2—The designer should coordinate testing with a laboratory to determine what combinations of degree of compaction and placement water content will result in this value of permeability or less. Design of the 1.5-foot-thick liner may proceed with those recommendations.

Construction considerations for compacted clay liners

Thickness of loose lifts

The permissible loose lift thickness of clay liners depends on the type of compaction roller used. If a tamping or sheepsfoot roller is used, the roller teeth should fully penetrate through the loose lift being compacted into the previously compacted lift to achieve bonding of the lifts. A loose lift thickness of 9 inches is commonly used by NRCS specifications. If the feet on rollers cannot penetrate the entire lift during compaction, longer feet or a thinner lift should be specified. A loose layer thickness of 6 inches may be needed for some tamping rollers that have larger pad type feet that do not penetrate as well.

Method of construction

Several methods are available for constructing a clay liner in an animal waste impoundment. Each has its advantages and disadvantages as described in following sections. A designer should consider the experience of local contractors and the relative costs of the methods in selecting the most appropriate design for a given site. The thickness of the planned soil liner, haul distance, planned side slopes for the pond, and other factors also guide a designer's decision on the best method to use.

Bathtub construction

This method of construction consists of a continuous thickness of soil compacted up and down or across the slopes. Figure 10D–10 shows the orientation of the lifts of a compacted liner constructed using this method, as contrasted to the stair step method, which is covered next. Figure 10D–11 shows two sites where the bathtub method of construction is being used.

This construction method has the following advantages over the stair-step method:

- The layers of compacted clay are oriented perpendicular to flow through the liner in this method. If the lifts making up the liner are not bonded well, the effect on seepage is minor, compared to the stair-step method.
- This method lends itself to constructing thinner lifts, which is more economical.

The bathtub construction method has the following disadvantages compared to the stair-step method:

- Side slopes must be considerably flatter than
 for the stair-step method, creating a pond with
 a larger surface area. A pond with a larger surface area has to store more precipitation falling
 on it, which could be considered an extra cost
 of the method.
- To permit equipment traversing up and down the slopes, slopes must be an absolute minimum of 3H:1V. Shearing of the soil by the equipment on steeper slopes is a concern. To prevent shearing of the compacted soil, the slopes of many compacted liners in ponds constructed using this method use 4H:1V slopes so that equipment will exert more normal pressure on the slope than downslope pressure.

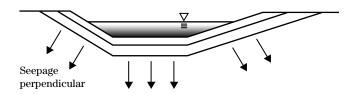
Stair-step construction

The stair-step method of construction is illustrated in figure 10D–10. Construction of the liner consists of compacting lifts of soil around the perimeter of the liner in a stair-step fashion, finishing the job by shaving off some of the side liner and placing it in the bottom of the pond. This method of construction is required if the side slopes of the pond are any steeper than about 3H:1V. Advantages of this method of construction are:

- A thicker blanket, measured normal to the slope, will result compared to the bathtub method of construction (fig. 10D-10). This is a positive factor in seepage reduction.
- It allows steeper side slopes, and thus the surface area of the pond exposed to rainwater accumulation is smaller than a bathtub construction would permit.
- The thicker blanket reduces the impact of shrinkage cracks, erosive forces, and potential mechanical damage to the liner.

Figure 10D–10 Methods of liner construction (after Boutwell 1990)

Bathtub construction



Stair-step construction

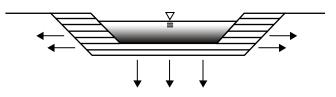


Figure 10D–11 Bathtub construction of clay liner (photo courtesy of NRCS Virginia (top) and NRCS Nebraska (bottom))





 Ponds constructed with this method are deeper for a given volume of waste than ponds constructed with the bathtub method, which favors anaerobic processes in the pond.

Disadvantages of the method are:

- This method may be more expensive than the bathtub method because the liner on the sides of the pond are thicker.
- Flow is parallel to the orientation of the layers forming the compacted liner on the pond sides.
 If care is not taken to obtain good bonding between lifts, seepage through the interface between lifts could be higher than expected.
- Contractors may be less familiar with this method of operation of equipment.

In the stair-step method of construction, the pond is first excavated. Borrow soil is then imported with a truck or scraper and spread in thin lifts (8 to 9 in thick) prior to compaction. Figure 10D–12a shows the first layer being constructed on the sides of the pond. This pond used a bentonite application. Each lift of soil is compacted with a sheepsfoot roller to obtain the desired dry density at the specified water content (fig. 10D–12b). The interior liner is constructed by bringing up lifts the full depth of the pond. Photo 10D–12c provides an overview of the stair-step process of constructing a clay liner in an animal waste storage pond. After the sides are constructed, some of the liner is shaved off and used to construct a liner in the bottom of the pond (fig. 10D–12c).

Soil type

Soils in groups III and IV are the most desirable for constructing a clay liner (table 10D–3). Some soils in group II may also be good materials for a clay liner, but definitely require laboratory testing to document their permeability characteristics. Soils in group I always require bentonite to form a liner with acceptably low permeability. Some soils in group II may also require bentonite to be an acceptable material for a liner. Some soils in groups III and IV require a soil dispersant to create an acceptably low permeability.

Classification

The most ideal soils for compacted liners are those in group III. The soils have adequate plasticity to provide a low permeability, but the permeability is not exces-

Figure 10D-12

Stair-step method (Photo credit John Zaginaylo, PA, NRCS)

(a)



(b)



(c)



sively high to cause poor workability. Group IV soils can be useful for a clay liner, but their higher plasticity index (PI greater than 30) means they are more susceptible to desiccation. If clay liners are exposed to hot dry periods before the pond can be filled, desiccation and cracking of the liner can result in an increase in permeability of the liner. A protective layer of lower PI soils is often specified for protection of higher PI clay liners to prevent this problem from developing.

Highly plastic clays like those in group IV are also difficult to compact properly. Special effort should be directed to processing the fill and degrading any clods in high plasticity clays to prevent this problem.

Size of clods

The size and dry strength of clay clods in soil prior to compaction have a significant effect on the final quality of a clay liner. Soil containing hard clayey clods is difficult to break down and moisten thoroughly. Adding water to the soil is difficult because water penetrates the clods slowly. High speed rotary pulverizers are sometimes needed if conditions are especially unfavorable. If soils containing large clay clods are not treated properly, the resultant permeability will be much higher than might otherwise be true. Figure 10D-13 shows the structure that results from compacting soils containing clods that are not adequately broken down.

Natural water content of borrow

The water content of soils used to construct a clay liner is the most important factor in obtaining a low permeability liner for a given soil. If soils are too dry, they cannot effectively be compacted to a condition where their structure is acceptable and their permeability may be higher than desirable. Compacting a soil at the proper water content creates a structure that is most favorable to a low permeability. Adding water to compacted clay liners is an additional expense that must be considered. A good rule of thumb is that it requires about 3.2 gallons of water to increase the water content of a cubic yard of compacted soil by 1 percent.

Dry conditions in the borrow

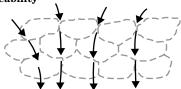
If soils in the borrow area are dry, several problems may need to be addressed. If the soils are clays with relatively high plasticity (PI values greater than about 20), they are likely to be very cloddy when excavated. Water is slow to penetrate the clods and compaction is less likely to degrade clods if enough time has not elapsed between adding the water and compaction. More descriptions follow in subsequent sections, and figure 10D–13 illustrates how clods left in the compacted fill will likely cause the soil to have a higher than expected permeability.

If the water content of borrow soils is more than 3 or 4 percent drier than required for specified compaction conditions, consideration should be given to wetting the soils in the borrow prior to construction. Adding large amounts of water during processing on the fill is

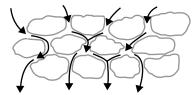
Figure 10D-13

Macrostructure in highly plastic clays with poor construction techniques (from Hermann and Elsbury 1987)

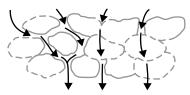
Micropermeability

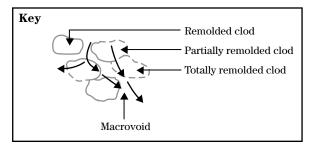


Macropermeability



Intermediate situation





difficult and inefficient. Sprinklers can be set up in the borrow some time before construction is planned and then time will allow water to soak into the soils more thoroughly.

Wet conditions in the borrow

If the natural water content of the borrow soil is significantly higher than optimum water content, achieving the required degree of compaction may be difficult. A good rule of thumb is that a soil will be difficult to compact if its natural water content exceeds about 90 percent of the theoretical saturated water content at the dry density to be attained. The following procedure can help to determine if the soils in the borrow are too wet for effectively compacting them.

Step 1 Measure the natural water content of the soil to be used as a borrow source for the clay liner being compacted.

Step 2 Compute the highest dry density to which the soil can be compacted at this water content using the following equation, which assumes that the highest degree of saturation achievable is 90 percent:

$$A chievable \ \gamma_{dry} \ lb/ft^3 = \frac{62.4}{\frac{W_n\%}{90} + \frac{1}{G_s}}$$

where:

w_n(%) = natural water content of borrow soils, % G_s = specific gravity of the soil solids (dimensionless)

Specific gravity values are obtained by ASTM Standard Test Method D854. An average value for specific gravity is often assumed to be 2.68. However, soils with unusual mineralogy may have values significantly different. Soils with volcanic ash may have specific gravity values as low as 2.3, and soils with hematite in them may have values as high as 3.3, based on NRCS laboratory results.

Step 3 Perform a Standard Proctor (ASTM D698) compaction test on the same soil and determine the maximum dry density value. Compute the achievable degree of compaction by dividing the computed value of achievable dry density by the maximum Standard Proctor dry density.

Step 4 If the computed achievable degree of compaction is less than 95 percent, then drying of the sample will probably be required. In rare cases, compaction to a lower degree, such as 90 percent of Standard Proctor, at higher water contents will achieve an acceptably low permeability. Laboratory tests should be performed to evaluate whether a lower degree of compaction will result in an acceptable permeability value.

Note: The experience of NRCS engineers is that when the natural water content of a soil is more than 4 percent above optimum water content, it is not possible to achieve 95 percent compaction. Computations should always be performed, as this rule of thumb sometimes has exceptions. In most cases, drying clay soils by only disking is somewhat ineffective, and it is difficult to reduce their water content by more than 2 or 3 percent with normal effort. It may be more practical to delay construction to a drier part of the year when the borrow source is at a lower water content. In some cases, the borrow area can be drained several months before construction. This would allow gravity drainage to decrease the water content to an acceptable level.

Step 5 Another way of examining this problem is to assume that soils must be compacted to 95 percent of their Standard Proctor (ASTM D698) dry density and then compute the highest water content at which this density is achievable. Commonly, soils are difficult to compact to a point where they are more than 90 percent saturated. The following equation is used to determine the highest feasible placement water content at which the dry density goal is achievable:

$$\label{eq:energy} \text{Highest placement w(\%)} = \frac{90(\%)}{100} \times \left[\frac{62.4}{\gamma_{\text{dry}} \text{ lb/ft}^3} - \frac{1}{G_{\text{s}}} \right]$$

Example 10D–4—Compute the achievable dry density of a potential borrow source

Given

A borrow source is located and found to be in a desirable group III type soil. The soil has 65 percent finer than the No. 200 sieve and a PI of 18. The soil was sampled and placed in a water tight container and shipped to a soils laboratory. The natural water content of the soil was measured to be 21.8 percent. The lab also performed a specific gravity (G_s) test on the soil, and

measured a value of 2.72. A Standard Proctor Test was performed on the sample and values for maximum dry density of 108.5 pounds per cubic foot and an optimum water content of 17.0 percent were measured.

Solution:

The maximum degree of compaction of this soil at the measured water content. If the soil is too wet to be compacted to 95 percent of maximum standard Proctor dry density, how much will it have to be dried to achieve compaction to 95 percent of maximum density?

$$\begin{split} & \text{Achievable } \gamma_{\rm dry} \ lb/ft^3 = \frac{62.4}{\frac{w_{\rm n}\%}{90} + \frac{1}{G_{\rm s}}} \\ & \text{Achievable } \gamma_{\rm dry} \ lb/ft^3 = \frac{62.4}{\frac{21.8\%}{90} + \frac{1}{2.72}} = 102.3 \ lb/ft^3 \end{split}$$

Next, compute the achievable degree of compaction by dividing the achievable dry density by the maximum Standard Proctor dry density, expressed as a percentage. The achievable degree of compaction is then equal to 102.3 divided by 108.5×100=94.3 percent.

Now, determine how wet the sample could be and still achieve 95 percent compaction. Ninety-five percent of the maximum Standard Proctor dry density is $0.95 \times 108.5 = 103.1$ pounds per cubic foot. Substitute this value into the equation given:

$$\begin{split} & \text{Highest placement } w\% = \frac{90}{100} \times \left[\frac{62.4}{\gamma_{\text{dry}} \ \text{lb/ft}^3} - \frac{1}{G_{\text{S}}} \right] \\ & \text{Highest placement } w\% = \frac{90}{100} \times \left[\frac{62.4}{103.1 \ \text{lb/ft}^3} - \frac{1}{2.72} \right] = 21.4\% \end{split}$$

This computation confirms the rule of thumb given that it is difficult to achieve 95 percent degree of compaction if the natural water content is greater than 4 percent above optimum. The stated value for optimum water content is 17.0 percent, so the rule of thumb says that if the natural water content exceeds 21.0 percent, achieving 95 percent degree of compaction will be difficult.

Methods of excavating and processing clay for liners

Clods in borrow soil

If borrow soils are plastic clays at a low water content, the soil will probably have large, durable clods. Disking may be effective for some soils at the proper water content, but pulverizer machines may also be required. To attain the highest quality liner, the transported fill should be processed by adding water and then turned with either a disk or a high-speed rotary mixer before using a tamping roller. Equipment requirements depend on the strength and size of clods and the water content of the soil.

Placement of lifts

Individual lifts of soil usually consist of an equipment width (often about 8 to 10 feet wide) layer of soil about 6 inches thick, after compaction. These lifts should be staggered to prevent preferential flow along the inter-lift boundaries. Figure 10D–14(a) shows the preferred way of offsetting the lifts. Figure 10D–14(b) shows a method that should be avoided. Bonding between the 6-inch lifts is also important so that if water does find its way down the boundary between two lanes of compacted soil that it cannot flow laterally and find the offset boundary.

Macrostructure in plastic clay soils

Clods can create a macrostructure in a soil that results in higher than expected permeability because of preferential flow along the interfaces between clods.

Figure 10D-14 Construction methods to limit interlift preferential flow paths

(a) Lanes for lift placement should be staggered to prevent preferential flow at sides of lifts. Bonding of lifts is also important to prevent flow along poorly bonded lifts.

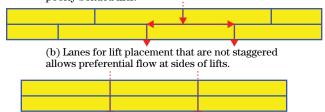


Figure 10D–13 illustrates the structure that can result from inadequate wetting and processing of plastic clay. The permeability of intact clay particles may be quite low, but the overall permeability of the mass is high because of flow between the intact particles.

Dry density and optimum water content

Compaction specifications for most earthfill projects normally require a minimum dry density (usually referenced to a specified compaction test procedure) and an accompanying range of acceptable water contents (referenced to the same compaction test procedure). This method of fill specification is usually based on engineering property tests such as shear strength, bearing capacity, and permeability. When permeability is the primary engineering property of interest, as would be the case for a compacted clay liner, an alternative type of compaction specification should be considered. The reason for this is a given permeability value can be attained for many combinations of compacted density and water contents (Daniels and Benson 1990). Figure 10D–15 illustrates a window of compacted dry density and water content in which a given permeability could be obtained for an example soil. The principles involved can be illustrated as follows.

Assume that a given soil is being used to construct a clay liner for an animal waste impoundment. A moderately plastic silty clay classifying as CL in the USCS is used. In case 1, the soil being obtained from a nearby borrow area has a relatively high natural water content. The contractor elects to use lighter construction equipment that applies a relatively low energy in compacting the soil. The result is the soil is compacted to a condition where the compacted density is relatively low and the placement water content is relatively high. This is labeled as point 1 in the figure 10D-15. In case 2, the same soil is being used, but the site is being constructed in a drier time of year. The contractor elects to use a larger sheepsfoot roller and apply more passes of the equipment to achieve the desired product. This time the same soil is compacted to a significantly higher density at a significantly lower water content. This is labeled point 2 in the figure 10D-15.

Laboratory tests can be used to establish the boundary conditions and arrive at a window of acceptable densities and water contents for a clay liner. Figure 10D–16 shows how a different structure results between soils compacted wet of optimum and those compacted dry of optimum water content. It also illustrates that soils compacted with a higher compactive effort or energy have a different structure than those compacted with low energy.

Figure 10D–15 Range at acceptable moisture/density for a typical clay liner

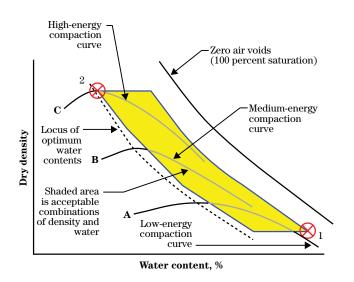
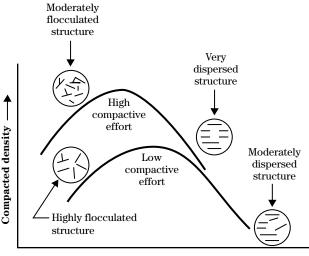


Figure 10D–16 Effect of water content and compactive effort on remolding of soil structure in clays (from Lambe 1958)



Molding water content →

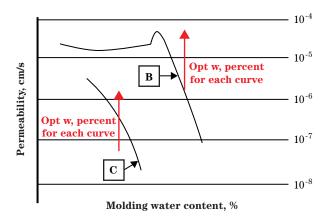
Mitchell (1965) was instrumental in explaining how the permeability of clay soils is affected by the conditions under which they were compacted. Figure 10D–17 illustrates results of one series of experiments summarized in the study. Two samples of a soil were compacted using different energy at different water contents and their permeability was measured. Soil C was compacted using higher energy, like that used when a heavy sheepsfoot roller passed over each compacted lift multiple times. Soil B was compacted using a lower energy, equating to a smaller roller with a smaller number of passes used in the compaction process.

The curves show the relationship between the permeability of the compacted soil and the compaction water content, for the two energies used. The following general principles are seen:

- The permeability of the low energy soil (curve B) is high unless the compaction water content is significantly wet of optimum. Very high permeability results for compaction dry of optimum.
- The permeability of the higher energy soil (curve C) is relatively high for water contents less than optimum.

Lambe (1958) explains how the energy used and the water content of the soil at the time of compaction affect the permeability of the soil by creating structure in the soil. Figure 10D–16 summarizes his explanation

Figure 10D–17 Plot showing effect of molding water content on permeability (Mitchell 1965)



of how different soil structures results from these two factors. Soils compacted with higher energy (heavier equipment and numerous passes of the equipment) at a higher water content have a dispersed structure. This structure creates very small plate-shaped voids that are resistant to water flow. Soils that are compacted with lower energy and/or lower water contents have a flocculated structure. This structure involves larger voids that are more conducive to water flow.

Percent saturation importance

Benson and Boutwell (2000) studied the correlation between field measured permeability values on compacted liners with laboratory measured values. The study found that when soils were compacted at drier water contents, even if a high density were obtained, that correlation between field and lab permeability test values was poor. The study found good correlation when soils were compacted at relatively higher water contents. Clods in clay soils are probably not broken down as well at lower compaction water contents which explains the higher permeability in the field. In lab tests, breaking down clods and obtaining test specimens without a structure is easier than done with field compaction procedures.

The conclusions of Benson and Boutwell's research were that if a designer is going to rely on laboratory permeability tests to predict the permeability of a compacted clay liner, the following rules of thumb apply.

• Soils should generally be compacted wet of the line of optimums. The line of optimums is illustrated in figure 10D-15. It is the locus of optimum water content values for a given soil for a range of compactive energy. A soil compacted with a low energy (like that resulting from a small sheepsfoot roller), curve A in figure 10D–15, will have a relatively low maximum density and high optimum water content. A soil compacted with a high energy (like that resulting from using a large heavy tamping roller), curve C in figure 10D-15, will have a high value for maximum density and a low value of optimum water content. The line of optimums is the locus of points connecting the values of optimum water content. Remember that optimum water content depends on the energy used and that Standard Proctor (ASTM D698) is only one standard type of compaction test. ASTM D1557,

the modified energy test is also used for design of some clay liners.

- Eighty percent of field tests of dry density and water content should plot to the right of the line of optimums if the field permeability is expected to reflect the same values obtained in laboratory testing.
- The average water content of all quality control tests should be from 2 to 4 percent wetter than the line of optimums as defined.

Energy level of compaction

The relationship of maximum dry density and optimum water content varies with the compactive energy used to compact a soil. Higher compactive energy results in higher values of maximum dry unit weight and lower values of optimum water content. Lower compactive energy results in lower values of maximum dry unit weight and higher values of optimum water content. Because optimum water content varies with the energy used in compaction, its nomenclature can be misleading. The optimum water content of a soil varies with the particular energy used in the test to measure it.

Compactive energy is a function of the weight of the roller used, thickness of the lift, and number of passes of the roller over each lift. Rollers should be heavy enough to cause the projections (teeth or pads) on the roller to penetrate or almost penetrate the compacted lift. Enough passes must be used to attain coverage and break up any clods. Additional passes do not compensate for rollers that are too light.

Roller size is often specified in terms of contact pressure exerted by the feet on sheepsfoot or tamping rollers. Light rollers have contact pressures less than 200 pounds per square inch, while heavy rollers have contact pressures greater than 400 pounds per square inch.

Limited data are available for various sizes of equipment to correlate the number of passes required to attain different degrees of compaction. Typically, from 4 to 8 passes of a tamping roller with feet contact pressures of 200 to 400 pounds per square inch are required to attain degrees of compaction of from 90 to 100 percent of maximum Standard Proctor dry density.

However, this may vary widely with the soil type and weight of roller used. Specific site testing should be used when possible.

Equipment considerations

Size and shape of teeth on roller

Older style sheepsfoot-type projections on rollers are best suited for compacting clay soils to achieve the lowest possible permeability. They are better suited than the modern style rollers called tamping rollers that have more square, larger area projections. The longer teeth on the older style sheepsfoot rollers are better at remolding plastic clay soils that are wet of optimum water content, and they are better at degrading clods in the soils (fig. 10D–18). The modern tamping-type rollers are effective in compacting soils at a drier water content when high bearing capacity is needed, like soils being compacted for highway subgrades (fig. 10D–19). The older style of sheepsfoot roller compactors are better suited for compaction to achieve low permeability.

Total weight of roller

To attain penetration of the specified loose lift, the roller weight must be appropriate to the specified thickness and the shape of the roller projections. Many modern rollers are too heavy to compact soils that are more than 1 or 2 percent wet of optimum water content. When the specified compaction water content is 2 percent or more wet of optimum water content, lighter rollers are essential. Permeability of clays is minimized by compaction at water contents wet of optimum.

Speed of operation

Heavy rollers operated at excessive speed can shear the soil lifts being compacted, which may result in higher permeability. Close inspection of construction operations should indicate if this problem is occurring, and adjustments to equipment or the mode of operation should then be made.

Vibratory versus nonvibratory sheepsfoot and tamping rollers

Some sheepsfoot and tamping rollers have an added feature, a vibratory action. This feature can usually be activated or deactivated while soils are being compacted. Vibratory energy adds little to the effectiveness

of these rollers when the soils being compacted are clays. At the same time, the vibration of the equipment is not usually detrimental. One condition in which the vibratory energy of this type of equipment might be detrimental is when a clay liner is being constructed on a subgrade of low plasticity silts or sands that are saturated. The vibration of the equipment often causes these types of foundation soils to become dilatant as they densify, and the water expelled in this process can create a trafficability problem. For this reason, when subgrade soils are saturated low plasticity silts and sands, the vibratory action of the compaction equipment should be disabled.

Vibratory smooth-wheeled rollers

Vibratory smooth-wheeled rollers are well suited to compacting bentonite-treated liners. They should not be used for compacting clay liners, however. The smooth surface of the roller results in poor bonding between lifts and can cause problems like those shown in figure 10D–14. The load distribution of the rollers also causes the top of a lift to be compacted well but the bottom of the lift not as well, when finegrained soils are being compacted. A vibratory smooth wheeled roller is shown in figure 10D–20.

Figure 10D-18 Longer style of teeth preferable for compacting soils for clay liner



Figure 10D-20 Smooth-wheeled steel roller compactor



Figure 10D-19 Modern type of tamping roller less well suited for compacting soils for clay liner



Freeze-thaw and desiccation

Freeze-thaw

Compacted clay liners may become damaged when the liner is exposed during freezing weather. Articles by Kim and Daniel (1992) and Benson and Othman (1993) describe the effects of freezing on clay liners and how the damage resulting from freezing may be permanent. Laboratory tests show that permeability rates may increase by 2 to 3 orders of magnitude (100-1,000 times). Freeze-thaw damage is more likely to affect the side slopes of a clay-lined pond than it will the bottom of the pond after it is filled. If freeze-thaw damage is regarded as likely to increase the permeability of the soils on the side slopes of the pond, a thicker liner or protective cap of cover soil should be considered. The extra cost of freeze-thaw protection may cause a designer to consider a synthetic liner alternative for reasons of economy and confidence in the low permeability of the synthetic liner. For instance, Minnesota designs often include the use of GCL liners for this reason.

Desiccation

Compacted clay liners may also be damaged when the liner is exposed during hot, dry weather after construction and before the pond is filled. Desiccation may also occur during periods the pond is emptied. Articles by Daniel and Wu (1993) and Kleppe and Olson (1985) describe factors that affect desiccation. Using the sandiest soil available that will be adequately impermeable is helpful. Compacting the soil as dense and dry as practical while still achieving the design permeability goal is also helpful. Protective layers must be at least 12 inches thick to be effective, and even thicker layers may be needed for more plastic clay liners, those with PI values of 30 or higher.

Design and construction of bentonite amended liners

When soils at grade of an excavated pond are low plasticity sands and silts in groups I or II of table 10D–3, an unlined pond will result in unacceptably high seepage losses. Several design options are normally considered for this situation. The options are listed as follows in order of increasing cost:

- Clay soils suitable for a clay liner are located in a nearby borrow area and imported to the site to construct a compacted clay liner. CPS 521D applies to this practice.
- Soils from the excavation and at the excavated subgrade are treated with bentonite to create a compacted liner with the required permeability and thickness. CPS 521C applies to this practice.
- The pond may be lined with geosynthetic, a GCL, or lined with concrete. An aboveground storage tank is also an option.

Bentonite type and quality

Several types of bentonite are mined and marketed for use in treating soils to produce a low permeability liner. The most effective type of bentonite (less volume required per cubic foot of treated soil) is finely ground sodium bentonite that is mined in the area of northeast Wyoming, southeast Montana, and western South Dakota. This sodium bentonite is derived from weathered volcanic ash. Sodium bentonite is a smectite clay composed primarily of the mineral montmorillonite (Bentofix 2007). It has the ability to swell up to 10 to 15 times its dry natural volume when exposed to water. Other types of bentonite, usually calcium bentonite are also mined and marketed for treating soils. These types of bentonites are less active (less free swell potential) and more volume of bentonite per treated cubic yard of soil will be required to produce a target permeability than would be required if sodium bentonite were used.

Two methods of evaluating a bentonite source being considered for use as an additive for a liner has high swell properties exist. They are:

- Determine the level of activity based on its Atterberg limit values as determined in a soil testing laboratory. High-quality sodium bentonite has LL values greater than 600 and PI values greater than 550.
- High-quality sodium bentonite has a free swell value of 22 milliliter or higher, based on experience of NRCS engineers and generally accepted guidance. An ASTM Standard test method to evaluate the free swell potential of bentonite is used to verify the quality of bentonite used

in GCL liners and is also suitable for evaluating bentonite proposed for a liner being constructed using CPS 521C. The ASTM method is D5890. A summary of the method follows.

- Prepare a sample for testing that consists of material from the total sample that is smaller than a No. 100 sieve.
- Partially fill a 100-milliliter graduated cylinder with 90 milliliters of distilled water.
- Add 2 grams of bentonite in small increments to the cylinder. The bentonite will sink to the bottom of the cylinder and swell as it hydrates. Wash the sides of the cylinder and fill to the 100-milliliter level.
- After 2 hours, inspect the hydrating bentonite column for trapped air or water separation in the column. If present, gently tip the cylinder at a 45-degree angle and roll slowly to homogenize the settled bentonite mass.
- After 16 hours from the time the last of sample was added to the cylinder, record the volume level in milliliters at the top of the settled bentonite. Record the volume of free swell, for example, 22 milliliters free swell in 24 hours.

Figure 10D–21 shows an excellent quality bentonite reaction to the test. It has a free swell of about 27 milliliters.

Figure 10D-21 Free swell test for bentonite ASTM D5890



Bentonite is furnished in a range of particle sizes for different uses. Fineness provided by the bentonite industry ranges from very finely ground, with most particles finer than a No. 200 sieve, to a granular form, with particles about the size of a No. 40 sieve. Laboratory permeability tests have shown that even though the same bentonite is applied at the same volumetric rate to a sample, a dramatic difference in the resulting permeability can occur between a fine and a coarse bentonite. It is important to use in construction the same quality and fineness as was used by the soils laboratory for the permeability tests to arrive at recommendations. Fineness for use in treating liners for waste impoundment can also be specified by an acceptable bentonite by supplier and designation, or equivalent. An example specification is Wyo Ben type Envirogel 200, CETCO type BS-1, or equivalent.

Design details for bentonite liner

The criteria given in CPS 521C, Pond Sealing or Lining, Bentonite Treatment, provide minimum required liner thicknesses for various depth of liquids.

CPS 521C provides guidance on rates of application of bentonite for preliminary planning purposes or where the size and scope of the project does not warrant obtaining samples and having laboratory tests performed. These preliminary recommended rates of application are based on using high-quality sodium bentonite that is finely ground. The CPS 521C includes a table that shows a range of recommended application rates which vary with the type of soil being treated. Higher rates of application are needed for coarse, clean sands and lower rates for silts. The table shows a recommended application rate expressed in pounds of bentonite per square foot per inch of liner to be built. For example, a typical rate of application for a relatively clean sand would be about 0.625 pounds per square foot per inch of compacted bentonite-treated liner. The most up-to-date CPS 521C should always be consulted for recommended rates, in case they have changed since this document was written.

For planning purposes, using these recommended rates, the amount of bentonite needed for a job can be estimated. For example, assume that a pond is to be constructed with an area of the sides and bottom totaling one acre. Assume that considering the planned depth of water in the pond, a design has been formu-

lated that calls for a 1-foot-thick bentonite-treated liner and that an application rate of 0.625 pounds per square foot per inch is needed. The total amount of bentonite required per square foot will be

$$0.625 \text{ lb/ft}^2 \times 12 \text{ in/ft} = 7.5 \text{ lb}$$

of bentonite per square foot. For an acre of pond area, the total amount needed will be

$$7.5 \text{ lb/ft}^2 \times 43,560 \text{ ft}^2/\text{acre} = 326,700 \text{ lb}$$

= 163 tons

The cost of bentonite is affected strongly by freight, and the further a site is from the area of the United States where bentonite is produced, the more costly it will be. Better unit prices are available for larger quantities.

Remember that the preliminary rates of application provided in CPS 521C assume that finely ground highswell sodium bentonite is used. If plans anticipate that a lower quality bentonite with a free swell less than about 22 milliliters or a coarsely ground bentonite may be used, laboratory testing is required to establish a rate of application that will create a suitably low permeability. Design using the specific discharge approach will establish what the target permeability value should be.

The recommended procedure to arrive at a design for a bentonite-treated liner then is as follows:

- Step 1 Obtain a sample of the soil to which the bentonite is to be added. Have the sample tested in a soils laboratory to determine its basic index properties, including percent fines and plasticity.
- Step 2 Have a standard Proctor (ASTM D698) test performed to determine the maximum dry density and optimum water content.
- Step 3 From the preliminary design of the site, determine the depth of water in the structure. Use CPS 521C to determine the minimum thickness of liner required.
- Step 4 Using given or assumed values for allowable specific discharge, compute the required permeability of the bentonite-treated liner.
- Step 5 Coordinate with a soils laboratory on testing to determine what degree of compac-

tion, water content, and rate of application of the proposed additive is required to obtain this permeability. Consider whether high quality (free swell > 22 mL) is being used and whether finely ground or coarsely ground bentonite is proposed.

Step 6 Design the final liner based on the results of step 5.

Example 10D–5—Design of a bentonite-treated liner

Given:

A waste storage pond is planned with a depth of liquid of 21 feet. The State requirement for the location is a specific discharge no greater than one-fifty-sixth of an inch per day of seepage. Assume the soils at grade have been tested and found to be suitable for bentonite treatment. Find the minimum thickness liner required according to CPS 521C, and determine the required permeability to meet this specific discharge requirement.

First, consult CPS 521C to determine the minimum required thickness. Assume the current CPS requires a liner that is 18 inches thick (1.5 ft).

Convert the specified unit seepage rate (specific discharge) of one-fifty-sixth of an inch per day into the same units as will be used for permeability (centimeters per second). To convert, use conversion values shown in table 10D–6, multiply:

$$v = \frac{1}{56}$$
 in/d×2.94×10⁻⁵ = 5.25×10⁻⁷ cm/s

The thickness of the liner and depth of liquid in the pond must also be converted to metric units. To convert the liner thickness of 18 inches to centimeters, multiply by 2.54, which equals a liner thickness, d, of 45.72 centimeters. The liquid depth, H, of 21 feet is equal to

$$H = 21 \text{ ft} \times 12 \text{ in/ft} \times 2.54 \text{ cm/in} = 640.1 \text{ cm}$$

Using the equation described previously, solve for the required permeability:

$$k = \frac{v \times d}{H + d}$$

$$k = \frac{5.25 \times 10^{-7} \text{ cm/s} \times 45.72 \text{ cm}}{640.1 \text{ cm} + 45.72 \text{ cm}} = 3.5 \times 10^{-8} \text{ cm/s}$$

The designer should coordinate with a soils laboratory to determine how much bentonite of given quality is required to obtain this low a permeability. In the experience of NRCS engineers, relying on this low a permeability means that construction quality control must be excellent and all the procedures and materials used are of highest quality. Seldom should designs for clay liners rely on a design permeability much lower than 5×10^{-8} centimeters per second. A designer might want to proceed with this design but require a slightly thicker liner (24 in) to provide additional assurance of obtaining the design specific discharge.

Considerations for protective cover

CPS 521C recommends considering the addition of a protective soil cover over the bentonite-treated compacted liner in waste impoundments. There are several reasons why a soil cover should be provided:

- Desiccation cracking of the liner after construction and prior to filling is a significant problem because the bentonite used in treatment is highly plastic.
- Desiccation cracking of the liner on the side slopes may occur during periods when the impoundment is drawn down for waste utilization or sludge removal. Desiccation cracking would significantly change the permeability of the liner. Rewetting generally does not completely heal the cracks.
- Bentonite-treated liners are generally thinner than compacted clay liners. Because the liner is thin, it can be more easily damaged by erosion from rainfall and runoff while the pond is empty. Rills in a thin liner provide a direct pathway for seepage.
- Over excavation by mechanical equipment during sludge removal can damage the liner. A minimum thickness of 12 inches measured normal to the slope and bottom is recommended for a protective cover. The protective cover should be compacted to reduce its erodibility.

Construction specifications for bentonite liner

The best equipment for compacting bentonite-treated liners is smooth-wheeled steel rollers, as shown in figure 10D-20. Crawler tractor treads are also effective. Sheepsfoot rollers that are often used in constructing clay liners are not as effective. CPS 521C specifies that for mixed layers, the material shall be thoroughly mixed to the specified depth with disk, rototiller, or similar equipment. In addition, intimate mixing of the bentonite is essential to constructing an effective liner. If a standard disk is used, several passes should be specified. A high-speed rotary mixer is the best method of obtaining the desired mix (fig. 10D-22). A minimum of two passes of the equipment is recommended to assure good mixing. When multiple passes of equipment are used for applying and mixing the bentonite, the passes should be in directions perpendicular to each other. This encourages a more homogeneous mixture.

Another construction consideration is the moisture condition of the soil into which the bentonite is to be mixed. Unless the soil is somewhat dry, the bentonite will most likely ball up and be difficult to thoroughly mix. Ideally, bentonite should be spread on a relatively dry soil, mixed thoroughly, then watered and compacted.

Depending on the type of equipment used, tearing of the liner during compaction can occur on slopes of 3H:1V or steeper. Compacting along, rather than up and down slopes, could be unsafe on 3H:1V or steeper side slopes. For most sites, slopes of 3.5H:1V or 4H:1V should be considered.

Figure 10D–22 Pulvermixer (high-speed rotary mixer) (Photo credit Stacy Modelski, NRCS)



Bentonite-treated liners are often constructed in lifts that are 4-inch compacted thickness. Liners should be designed in multiples of 4 inches for this reason. Often, the first layer of bentonite-treated soil is the soil exposed in the bottom of the excavation. By applying bentonite to the exposed grade, disking it in to a depth of about 6 inches, and compacting it, the first layer is formed. Subsequent lifts are formed by importing loose fill adequate to form additional 4-inch-thick lifts.

Design and construction of clay liners treated with soil dispersants

Previous sections of this appendix caution that soils in groups III and IV containing high amounts of calcium may be more permeable than indicated by the percent fines and PI values. Groups III and IV soils predominated by calcium usually require some type of treatment to serve as an acceptable liner. The most common method of treatment to reduce the permeability of these soils is use of a soil dispersant additive containing sodium.

Types of dispersants

The dispersants most commonly used to treat high calcium clays are soda ash ($\mathrm{Na_2CO_3}$) and polyphosphates. The two most common polyphosphates are tetrasodium pyrophosphate (TSPP), and sodium tripolyphosphate (STPP). Common salt (NaCl) has been used in the past, but it is considered less permanent than other chemicals and is not permitted in the current CPS 521B. NRCS experience has shown that usually about twice as much soda ash is required to effectively treat a given clay when compared to the other two dispersants. However, because soda ash is often less expensive, it may be the most economical choice in many applications.

Design details for dispersant-treated clay liner

CPS 521B, Pond Sealing or Lining, Soil Dispersant, provides minimum thicknesses of liners using the dispersant-treated layer method, based on the depth

of liquid in the pond. CPS 521B provides guidance on approximate rates of application of soil dispersants based on testing performed by the NRCS laboratories. Rates provided in the CPS are in terms of pounds of dispersant required per 100 square feet for each 6-inch layer of liner. The total amount of dispersant per 100 square feet is then equal to the number of 6 inch lifts in the completed liner multiplied by the rate per lift.

Example 10D–6—Steps in design of a dispersant-treated liner

Assume for the purposes of this example that a soil has been tested at a site and found to be a flocculated clay with an unacceptably high permeability. The designer chooses to evaluate a soda ash-treated liner. Consult the current CPS 521B for guidance on application rates for soda ash. Assume that the current CPS suggests an application rate of 15 pounds of soda ash per 100 square feet of liner for each 6-inch-thick lift of finished liner. Next, assume that based on the depth of water in the pond that the CPS 521B requires a total liner thickness of 12 inches. Then, because each 6-inch-thick lift requires 15 pounds of soda ash per 100 square feet, the total amount of soda ash required for this example would be 30 pounds of soda ash per 100 square feet. The most up-to-date CPS 521B should always be consulted for recommended rates, in case they have changed since this document was written.

The recommended rates of application of dispersants in CPS 521B are based on the most up-to-date information from the NRCS soils testing laboratories. The rates are in general conservative, and if a designer wanted to evaluate lower rates of application, samples should be obtained and sent to a laboratory for documenting the efficacy of lower rates. If this procedure is followed, the following steps are usually implemented.

- Step 1 Obtain a sample of the soil to which the dispersant is to be added. Have the sample tested in a soils laboratory to determine its basic index properties, including percent fines and plasticity.
- Step 2 A standard Proctor (ASTM D698) test is performed to determine the maximum dry density and optimum water content.
- Step 3 From the preliminary design of the site, determine the depth of water in the structure and use CPS 521B to determine the minimum thickness of liner required.

Step 4 Using given or assumed values for allowable specific discharge, compute the required permeability of the dispersant-treated liner.

Step 5 Coordinate with a soils laboratory on testing to determine what degree of compaction, water content, and rate of application of the proposed additive is required to obtain this permeability. Consider local practice and consult suppliers to determine the relative costs of soda ash versus polyphosphates.

Step 6 Design the final liner based on the results from previous steps.

Example 10D–7—Comprehensive example for a dispersant-treated liner

Given:

A waste storage pond is planned with a depth of liquid of 18 feet. The State requirement for the location is a specific discharge no greater than 2,000 gallons per acre per day of seepage. Assume the soils at grade have been tested and found to require dispersant treatment. Assume that the current CPS 521B requires a minimum liner thickness of 1.5 feet. The example problem is to determine what permeability is required to meet the stated specific discharge requirement.

Solution:

First, the required specific discharge value, which is given in units of gallons per acre per day has to be converted the same units that will be used for required permeability. Assume that permeability will be expressed in centimeters per second, so use table 10D–6 to convert the value of 2,000 gallons per acre per day to centimeters per second as follows:

$$v = \frac{2,000 \text{ gal/acre/d}}{9.24 \times 10^8} = 2.2 \times 10^{-6} \text{ cm/s}$$

Next, convert the liner thickness and depth of liquid from units of feet to centimeters:

$$d = 18 \text{ in} \times 2.54 \text{ cm/in} = 45.72 \text{ cm}$$

$$H = 18 \text{ ft} \times 12 \times 2.54 \text{ cm/ft} = 548.64 \text{ cm}$$

Using the equation described previously, solve for the required permeability:

$$k = \frac{v \times d}{H + d}$$

$$= \frac{2.2 \times 10^{-6} \text{ cm/s} \times 45.72 \text{ cm}}{548.64 \text{ cm} + 45.72 \text{ cm}}$$

$$= 1.7 \times 10^{-7} \text{ cm/s}$$

The designer should coordinate with a soils laboratory to determine how much soil dispersant of the desired type is required to obtain this low a permeability. In the experience of NRCS engineers, obtaining this value of permeability using a soil dispersant should not require special effort or unusual amounts of additive. At the same time, seldom should designs for dispersant-treated clay liners rely on a design permeability much lower than 5×10^{-8} centimeters per second. A designer should proceed with this design specifying the application rate recommended by the soils lab and a 1.5-foot-thick liner to obtain the design specific discharge.

Construction specifications for a dispersant-treated clay liner

The best equipment for compacting clays treated with dispersants is a sheepsfoot or tamping type of roller. CPS 521B specifies that the material shall be thoroughly mixed to the specified depth with a disk, high speed rotary mixer, or similar equipment. Because small quantities of soil dispersants are commonly used, uniform mixing of the dispersants is essential to constructing an effective liner. If a standard disk plow is used, several passes should be specified. A high-speed rotary mixer is also essential to obtain a thorough mixture of the dispersant with the clay being amended. Figure 10D–23 shows this type of equipment. At least two passes of the equipment is recommended to assure good mixing.

Other construction considerations are also important. Using the bathtub method of construction on slopes of 3H:1V or steeper can cause tearing of the liner during compaction and reduce the effectiveness of compaction equipment. Slopes as flat as 3.5H:1V or 4H:1V should be considered for this factor alone, for bathtub type construction.

Current CPSs usually require a liner thicker than 6 inches. A liner generally can be satisfactorily constructed in a series of lifts by mixing in the required amount of soil dispersant to a 9-inch-thick loose depth and then compacting it to the 6 inches. Thicker liners should be constructed in multiple lifts, with the final compacted thickness of each lift being no greater than 6 inches.

Figure 10D-23 High-speed rotary mixer used to mix dispersants into clays (*Photo credit Jody Kraenzel*, *NRCS*)



Uplift pressures beneath clay blankets

A clay blanket may be subject to uplift pressure from a seasonal high water table in the foundation soil underneath the clay liner. The uplift pressure in these cases can exceed the weight of the clay liner, and failure in the clay blanket can occur (fig. 10D–24). This problem is most likely to occur during the period before the waste impoundment is filled and during periods when the impoundment may be emptied for maintenance and cleaning. Figure 10D–25 illustrates the parameters involved in calculating uplift pressures for a clay blanket. The most critical condition for analysis typically occurs when the pond is emptied. Thicker blankets to attain a satisfactory safety factor should be used if they are required.

The factor of safety against uplift is the ratio of the pressure exerted by a column of soil to the pressure of the ground water under the liner. It is given by the equation:

$$FS = \frac{\gamma_{sat} \times d \times \cos(\alpha)}{z \times \gamma_{water}}$$

where:

d = thickness of liner, measured normal to the

slope

 α = slope angle

 γ_{water} = unit weight or density of water γ = saturated unit weight of clay liner

z = vertical distance from middle of clay liner

to the seasonal high water table

Figure 10D-24 Failure of compacted liner from uplift forces below clay blanket (Photo credits NRCS, TX)





A factor of safety of at least 1.1 should be attained. The safety factor can be increased by using a thicker blanket or providing some means of intercepting the ground water gradient and lowering the potential head behind the blanket. Often, sites where seasonal high water tables are anticipated designs include a perimeter drain to collect the water and prevent this type of damage. Another option is a concrete structure above ground.

Another situation where a clay liner may be damaged from hydrostatic pressure is one where a site is located in a flood plain of a stream or river. The site may have to be built above ground level in this location to avoid a seasonal high water table. Figure 10D-26 illustrates the problem that may occur that must be considered by designers. A temporary flood condition in the flood plain can subject the agricultural waste impoundment to a differential head when the pond is empty. The pond could be empty shortly following construction or it could be empty to apply waste to crops. Uplift pressure may cause piping of sandy horizons underlying the site and boils, and sloughing of side slopes can occur as shown in figure 10D-26. The photo shows a claylined animal waste impoundment where the clay liner was damaged from excessive hydrostatic uplift forces caused by temporary storage of flood waters outside the embankment. The liner must be thick enough to resist predicted buoyant forces if it is possible for the pond to be empty or near empty during a flood. Drains will be ineffective because in a flood, outlets will be submerged.

Figure 10D–26 Uplift conditions caused by temporary flood stage outside lagoon (*Photo credit NRCS*, *WA*)

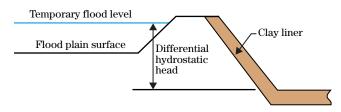
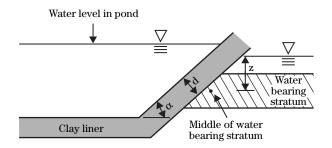






Figure 10D-25 Uplift calculations for high water table and clay blanket (from Oakley 1987)



Perimeter drains for animal waste storage ponds

When a high water table is anticipated and uplift pressures are anticipated, one approach to solving the problem is to install a drain around the pond. The drain may completely encircle the pond if a designer anticipates a general elevated water table in the site vicinity. At other sites with a more sloping ground surface, the perimeter drain may only be installed on the side(s) of the impoundment where the elevated water table is anticipated. Drains may be used both for clay liners and geosynthetic liners.

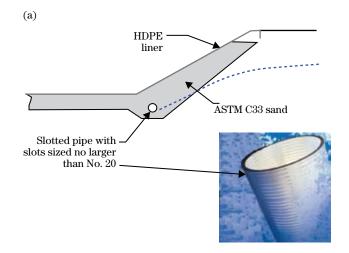
Drains usually are constructed by

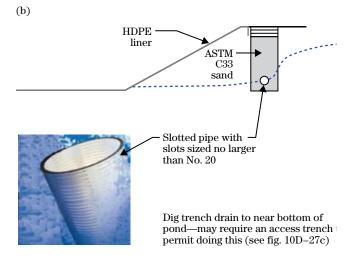
- digging a trench to the depth needed to draw down the water table
- placing a perforated or slotted drainage pipe
- surrounding the drain with granular material that is compatible with both the slot size in the pipe and the gradation of the surrounding foundation soils

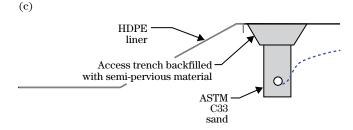
Pipes with small slots that are compatible with a filter sand like ASTM C-33 are preferred to avoid having to use two filter gradations. If pipes with larger perforations are used, they should be surrounded with gravel to prevent particles from moving into the pipe. Figure 10D-27 (a, b, and c) show typical installations where a single filter and perforated pipe is used. Another approach to installing a drain is to dig a trench, line it with geotextile, and after putting a slotted collector pipe in the trench, filling it with gravel. Figure 10D-28 shows this type of installation.

Several types of drain pipe may be used. One type is a low strength corrugated pipe with slots or perforations surrounded by a filter envelope of granular material. Figure 10D–29 is an example of this type of collector pipe. If a higher strength pipe is required, figure 10D–30 shows another type of pipe that is sometimes used for these types of installations.

Figure 10D–27 Typical drain installations using single filter with well-screened collector pipe







Illustrated access trench construction to permit installing deeper trench drain. Access trench filled with semi-pervious soil to limit infiltration of surface runoff.

Figure 10D-28

Perforated collector pipe installed the gravel envelope with trench lined with geotextile



Figure 10D-30

Corrugated drainage pipe with slots, doubled walled pipes may be specified if higher strengths are needed



 $\textbf{Figure 10D-29} \quad \text{Low-strength perforated drainage tubes}$



Soil mechanics testing for documentation

Laboratory soil testing may be required by regulations for design, or a designer may not choose to rely on correlated permeability test values. The NRCS National Soil Mechanics Center Laboratories have the capability to perform the necessary tests. Similar testing is also available at many commercial labs. The

accepted method of permeability testing is by ASTM Standard Test Method D5084, Measurement of Hydraulic Conductivity of Saturated Porous Materials Using a Flexible Wall Permeameter. Figure 10D–31 shows the equipment used for performing the test.

Contact the labs for more detailed information on documentation needed and for procedures for submitting samples.

Figure 10D-31 Equipment used for performing ASTM D5084



Molding a sample for a flexible wall permeability test



Disassembled mold with compacted specimen



Preparing sample in cell for flexible wall permeability test



Molded sample after dissembling mold

If the only tests requested are gradation and Atterberg limit tests, smaller samples are needed. The size of sample that should be submitted depends on the gravel content. The following recommendations should be adhered to:

Estimated gravel content of the sample 1/ (%)	Sample moist weight (lb)
0–10	5
10–50	20
>50	40

The sample includes the gravel plus the soil material that passes the No. 4 sieve (approx. 1/4-inch mesh).

If gradation analysis, Atterberg limits, compaction, and permeability testing are requested, considerably larger samples are required. When all these tests are needed, the sample size should be as follows:

Estimated gravel content of the sample 1/ (%)	Sample moist weight (lb)
0–10	50
10–50	75
>50	100

The sample includes the gravel plus the soil material that passes the No. 4 sieve (approx. 1/4-inch mesh).

Submitting samples at their natural water content is important so designers can compare the natural water content to reference compaction test values. Samples should always be shipped in moisture proof containers for this reason. The best container for this purpose is a 5-gallon plastic pail commonly obtained in hardware stores. These pails have tight fitting lids with a rubber gasket that ensures maintenance of the water content in the samples during shipping. These 5-gallon pail containers are much more robust and less likely to be damaged during shipment than cardboard containers.

If designs rely on a minimum degree of compaction and water content to achieve stated permeability goals in a clay liner, testing of the clay liner during construction may be advisable to verify that design goals have been achieved. Field density and water content measurements are routinely made using procedures shown in NEH, Section 19, Construction Inspection.

Other methods for documenting liner seepage

Performing density/water content tests during construction is a generally accepted method of documenting that a clay liner has been constructed according to specifications. If the liner is found to meet the requirements of the compaction specifications, the assumption is that the permeability values documented from laboratory testing on samples that were compacted at the specified density and water content will be achieved. In some cases, no additional documentation is required. In other cases, regulations require obtaining samples of the completed liner and performing permeability tests on them. Figure 10D–32 shows one way that a Shelby tube type of sample may be obtained without mobilizing a drilling rig. The Shelby tube used is typically a standard tube with a 3-inch outside diameter and 2 7/8-inch inside diameter. This size sample can be placed directly in a flexible wall permeameter for testing, after extrusion in the laboratory.

Another method for obtaining a sample of a compacted clay liner is with a drive sampler like that shown in figure 10D-33.

Figure 10D–32 Shelby tube sample being obtained with backhoe bucket used to force tube into clay liner (Photo credit Jody Kraenzel, NRCS, NE)



Figure 10D-33

Obtaining undisturbed sample of compacted clay liner using thin-walled drive cylinder



In the situation where a storage pond was constructed several years before documentation on quality of construction and permeability was required, studies are sometimes made in an attempt to measure seepage losses directly. One approach that has been used was developed by researchers at Kansas State University. This approach involves installing precise water level monitoring devices and evaporation stations. Seepage losses can be estimated by carefully monitoring the levels in the pond during periods when no waste is introduced into the pond and no rainfall occurs. After estimating the amount of evaporation, and subtracting that from the total decline in the level of the pond during that period, seepage loss can be estimated. Figure 10D–34 shows equipment for measuring evaporation in a pond.



Figure 10D-34

Equipment used to monitor evaporation at an agriculture waste storage lagoon. Measurements are used in total lagoon seepage evaluations.





Summary

- The reduction in the quantity of seepage that occurs as manure solids accumulate in the bottom and on the sides of storage ponds and treatment lagoons is well documented. However, manure sealing is not effective for soils with a low clay content. Its effectiveness is not accepted by all designers and cannot be used in the designs of storage ponds by some State and local regulations.
- Soils can be divided into four permeability groups based on their percent fines (percent finer than the No. 200 sieve) and plasticity index (PI). Soils in groups III and IV may be assumed to have a coefficient of permeability of 1×10⁻⁶ centimeters per second or lower unless they have an unusual clay chemistry (high calcium), or they have a very blocky structure.
- Group I soils will generally require a liner. Soils
 in group II will need permeability tests or other
 documentation to determine whether a desirable permeability rate can be achieved for a
 particular soil.
- If natural clay blankets are present at a site below planned grade of an excavated pond, the seepage rate should be estimated based on measured or estimated permeability values of the low permeability horizons beneath the liner and above an aquifer. If the estimated seepage rate is less than that given in NRCS guidance or State regulations, no special compacted liner may be required. If the soils at grade are not of sufficient thickness and permeability to produce a desirably low seepage rate, a liner should be designed to achieve the seepage rate that is the design goal.
- Guidance is given on factors to consider whether a constructed liner may be required. Four conditions are listed in which a liner should definitely be considered.
- Allowable specific discharge values are discussed and guidance is provided on reasonable values to use for design when other regulatory requirements are not specified.
- Flexibility is built into the design process. The depth of the liquid, the permeability, and thick-

- ness of the soil liner can be varied to provide an acceptable specific discharge.
- The guidelines provided for design of clay liners in this appendix provide designers with the tools to evaluate the probable unit seepage or specific discharge through a clay liner. The methods presented allow a designer to determine what treatment is required to achieve specific discharge or permeability goals.
- Methods provide designers with the ability to evaluate the effect of changes in a proposed design on the estimated unit seepage rate.
- As additional research becomes available, practice standards and guidance in this document may warrant revision.

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